

1 Field comparison of dry deposition samplers for collection of atmospheric 2 mineral dust: results from single-particle characterization

3 Andebo Waza¹, Kilian Schneiders¹, Jan May², Sergio Rodríguez^{3,4}, Bernd Epple², Konrad Kandler¹

4 ¹Atmospheric Aerosol, Institute for Applied Geosciences, Technische Universität Darmstadt, D-64287
5 Darmstadt, Germany

6 ²Institute for Energy Systems & Technology, Technische Universität Darmstadt, D-64287 Darmstadt, Germany

7 ³Izaña Atmospheric Research Centre, AEMET, Tenerife, Spain.

8 ⁴Estación Experimental de Zonas Áridas, EEZA CSIC, Almería, Spain.

9 *correspondence to andebo.waza@geo.tu-darmstadt.de

10 Abstract

11 Frequently, passive dry deposition collectors are used to sample atmospheric dust deposition. However,
12 there exists a multitude of different instruments with different, usually not well-characterized sampling
13 efficiencies. **As a result**, the acquired data might be considerably biased with respect to their size
14 representativity, and as consequence, also composition. In this study, individual particle analysis by
15 automated scanning electron microscopy coupled with energy-dispersive X-ray was used to characterize
16 different, commonly used passive samplers with respect to their size-resolved deposition rate and
17 concentration. This study focuses on the microphysical properties, i.e. **the aerosol concentration and**
18 **deposition rates as well as the particle size distributions**. In addition, computational fluid dynamics
19 modeling was used in parallel to achieve deposition velocities from a theoretical point of view.

20 **SEM calculated deposition rate measurements made using different passive samplers show a**
21 **disagreement among the samplers. Modified Wilson and Cooke (MWAC) and Big Spring Number Eight**
22 **(BSNE) - both horizontal flux samplers - collect considerably more material than Flat plate and the**
23 **Sigma-2, which are vertical flux samplers. The collection efficiency of MWAC increases for large**
24 **particles in comparison to Sigma-2 with increasing wind speed, while such an increase is less observed**
25 **in the case of BSNE. A positive correlation is found between deposition rate and PM₁₀ concentration**
26 **measurements by an optical particle spectrometer. The results indicate that a BSNE and Sigma-2 can be**
27 **good options for PM₁₀ measurement, whereas MWAC and Flat plate samplers are not a suitable choice.**
28 **A negative correlation was observed in between dust deposition rate and wind speed. Deposition**
29 **velocities calculated from different classical deposition models do not agree with deposition velocities**
30 **estimated using computational fluid dynamics simulations (CFD). The deposition velocity estimated**

31 from CFD was often higher than the values derived from classical deposition velocity models. Moreover,
32 the modeled deposition velocity ratios between different samplers do not agree with the observations.

33 **Key words:** Mineral dust particles, passive samplers, SEM-EDX, single particle analysis, computational
34 fluid dynamics

35 1 Introduction

36 Mineral dust aerosol in the climate system has received considerable scientific attention mainly due to
37 its direct effect on the radiative budget and indirect one on cloud microphysical properties (Arimoto,
38 2001; Huang et al., 2010). Mineral dust particles also play a key part with respect to gas phase chemistry
39 by providing a reaction surface e.g. ozone depletion (Nicolas et al., 2009; Prospero et al., 1995).
40 Moreover, dust aerosol also plays an important role in biogeochemical cycles by supplying important
41 and limiting nutrients to Ocean surfaces (Jickells et al., 2005). Mineral dust is emitted mainly from the
42 arid and semi-arid regions of the world and believed to have a global source strength ranging from 1000-
43 3000 Tgyr⁻¹ (Andreae, 1995). They form the single largest component of global atmospheric aerosol mass
44 budget, contributing about one third of the total natural aerosol mass annually (Penner et al., 2001).

45 Deposition measurement data of mineral dust are useful to validate numerical simulation models and to
46 improve our understanding of deposition processes. However, the scarcity and the limited
47 representativity of the deposition measurement data for validation pose a major challenge to assess dust
48 deposition at regional and global scales (Schulz et al., 2012; WMO, 2011). This is in part linked to the
49 uncertainties evolving from the use of different and non-standardized measurement techniques.

50 Commonly, deposition is measured by passive techniques, which provide an acceptor area for the
51 depositing atmospheric particles. The advantage of these passive samplers is that they operate passively,
52 resulting in simple and thus cheaper instruments, so that many locations can be sampled at a reasonable
53 cost (Goossens and Buck, 2012). The usual lack of a power supply allows also for unattended remote
54 setups. However, the most important disadvantage is that collection efficiency and deposition velocity is
55 determined by the environmental conditions not under operator control, and in remote setups also
56 frequently also unknown. That implies, in addition, that the sampler shape can have a strong and variable
57 impact of the collection properties.

58 While there is previous work describing and modeling single samplers (Einstein et al., 2012; Wagner and
59 Leith, 2001a, b; Yamamoto et al., 2006) and a few comparison studies (Goossens and Buck, 2012;
60 Mendez et al., 2016), most previous studies (Goossens and Buck, 2012; López-García et al., 2013) only

61 compare total mass, thereby neglecting size dependence and potential comparison biases. Also, a
62 systematic assessment of the impact of wind conditions is not commonly carried out, but for example
63 Mendez et al. (2016) showed that the efficiency of the BSNE and MWAC samplers for collecting PM₁₀
64 varies with wind speed, and Goossens and Buck (2012) found that PM₁₀ concentrations from BSNE and
65 DustTrak samplers have comparable values for wind speeds from 2–7 m/s.

66 The purpose of this study is to assess the particle collection properties of different deposition and other
67 passive samplers based on single particle measurements, and to assess their agreement with theory. From
68 the available data, also relations of the collected particle microphysics and composition homogeneity
69 between the samplers will be presented, which can be used as estimators for the comparability of previous
70 literature data based on the different techniques. To the best of our knowledge, this is the first study to
71 analyze dry deposition measurements collected using passive samplers by means of a single-particle
72 SEM-EDX Analysis approach (particularly in the size fraction larger than 10 μm).

73 2 Material and methods

74 2.1 Sampling location and time

75 Sahara and Sahel provide large quantities of soil dust, resulting in a westward flow of mineral dust
76 particles over the North Atlantic Ocean accounting for up to 50% of global dust budget (Goudie and
77 Middleton, 2001). Owing to proximity to the African continent, the Canary Islands are influenced by dust
78 particles transported from Sahara and Sahel regions. Therefore, Tenerife is one of the best locations to
79 study relevant dust aerosol in a natural environment.

80 For this study, we conducted a two month (July to August 2017) aerosol collection and dry deposition
81 sampling campaign at Izaña Global Atmospheric Watch observatory (Bergamaschi et al., 2000;
82 Rodríguez et al., 2015) (28.3085°N, 16.4995°W). Sampling inlet were placed at a height of 2 m above
83 ground, on top of a measurement installation. The installation was made on a 160 m² flat concrete
84 platform. The trade wind inversion, which is a typical meteorological feature of the **station, shields** most
85 of the time the observatory from local island emissions (García et al., 2016). Therefore, the Izaña Global
86 Atmospheric Watch observatory is an ideal choice for in-situ measurements under “free troposphere”
87 conditions (Bergamaschi et al., 2000; García et al., 2016).

88 2.2 Wind measurements

89 An ultra-sonic anemometer (Young model 81000, R. M. Young Company, Traverse City, MI, USA) was
90 installed at approximately 2 m height above the ground to obtain the 3-D wind velocity and direction. It
91 was operated with a time resolution of 10 Hz to get basic information on turbulence structure.

92 2.3 Particle sampling

93 Samples were collected from different, commonly used samplers, namely Big Spring Number Eight
94 (BSNE) (Fryrear, 1986), Modified Wilson and Cooke (MWAC) (Wilson and Cook, 1980), Sigma-2
95 (VDI2119, 2013) and Flat plate (UNC-derived) (Ott and Peters, 2008). In addition, the free-wing
96 impactor (FWI) (Kandler et al., 2018) was used to collect coarser particles. The BSNE, MWAC, FWI
97 and Filter Sampler were mounted on wind vanes to align them to the ambient wind direction. [Samples](#)
98 [were collected continuously, and substrates were exchanged at intervals of 24 hours. The sampling](#)
99 [duration for FWI \(12 mm Al-stub\) was 30 min only to avoid overloading. The sampling duration for filter](#)
100 [sampler was set to be one hour. It has to be noted that the PM₁₀ from optical measurements for this](#)
101 [particular 0.5 or 1 hour only deviates by 2 % and 0.2 % respectively from the 24-h-average.](#)

102 2.3.1 Flat plate sampler

103 The Flat plate sampler used in this work was taken from the original Flat plate geometry used in Ott and
104 Peters (2008). Briefly, the geometry contains two round brass plates (top plate diameter 203 mm, bottom
105 plate 127 mm, thickness 1 mm each) mounted in a distance of 16 mm. Unlike the original design, the
106 geometry of the current work has a cylindrical dip in the lower plate, which recedes the sampling substrate
107 – a SEM stub with a thickness of 3.2 mm – from the airflow, thereby reducing the flow disturbance. A
108 preliminary study with the modified and original setup side-by-side in a rural environment had shown
109 that this recession approximately doubles the collection efficiency for large particles. In this design, larger
110 droplets (> 1 mm) are prevented by this setup from reaching the SEM stub surface at the local wind
111 speeds Ott and Peters (2008). As described in Wagner and Leith (2001a); (Wagner and Leith, 2001b),
112 the main triggers for particle deposition on the substrates for this sampler are diffusion, gravity settling,
113 and turbulent inertial forces, of which only the latter two are relevant in our study.

114 2.3.2 Sigma-2 sampler

115 The Sigma-2 sampling device is described in Dietze et al. (2006); (Schultz, 1989; VDI2119, 2013).
116 Briefly, the geometry consists of a cylindrical sedimentation tube with a height of about 27 cm made of
117 antistatic plastic, which is topped by a protective cap with diameter of 158 mm. At its top, the cap has

118 four rectangular inlet windows (measuring 40 mm x 77 mm, all at the same height) at its side providing
119 away for passive entrance of particles to the collection surface. Once entered the tube, particles are
120 assumed settle down to the collection surface due to gravitation (Stokes' law) (VDI2119, 2013). The
121 samplers designed in a way that it protects the sample from direct radiation, wind and precipitation.

122 2.3.3 The Modified Wilson and Cooke (MWAC) sampler

123 The MWAC sampler is based on an original design developed by Wilson and Cook (1980). The sampler
124 consists of a closed polyethylene bottle, serving as settling chamber, to which an inlet tube and an outlet
125 tube have been added. The MWAC sampling bottles are 95 mm long with a diameter of 48 mm. The two
126 inlet and outlet plastic tubes with inner and outer diameter 8 and 10 mm respectively, pass air through
127 the cap into the bottle and then out again. The large volume of the bottle relative to the inlet diameter
128 makes the dust particles entering the bottle to be deposited in the bottle due to the flow deceleration the
129 total bottle area, and due to impaction below the exit of the inlet tube. The air then discharges from the
130 bottle via the outlet tube. MWAC is one of the most commonly used samplers (Goossens and Offer,
131 2000) and has a high sampling efficiency for large particles (Mendez et al., 2016).

132 2.3.4 The Big Spring Number Eight (BSNE) sampler

133 The BSNE sampler, originally designed by Fryrear (1986), is intended to collect airborne dust particles
134 from the horizontal flux (Goossens and Offer, 2000). Briefly, the particle laden air passes through a
135 rectangular inlet (21 mm wide and 11 mm high, with total area of 231 mm²). Once inside the sampler,
136 air speed is reduced by continuous cross section increase (angular walls) and the particles settle out on a
137 collection surface. Air discharges through a mesh screen.

138 2.3.5 Free-wing impactor (FWI)

139 A free rotating wing impactor (Jaenicke and Junge, 1967; Kandler et al., 2018; Kandler et al., 2009) was
140 used to collect particles larger than approximately 5 µm. A FWI has a sticky impaction surface attached
141 to a rotating arm that moves through air; particles deposit on the moving plate due to their inertia. The
142 rotating arm is moved at constant speed by a stepper motor, which is fixed on a wind vane, aligning the
143 FWI to wind direction. The particle size cut-off is defined by the impaction parameter, i.e. by rotation
144 speed, wind speed and sample substrate geometry. Details of working principle of FWI can be obtained
145 from Kandler et al. (2018)

146 2.3.6 Filter sampler

147 A filter sampler with Nucleopore filters (Whatman® Nucleopore™ Track-Etched Membranes diam. 25
148 mm, pore size 0.4 µm, polycarbonate) mounted on a wind vane was used for iso-axial particle collection.
149 An inlet nozzle of 6 mm was used to achieve pseudo-isokinetic conditions. Sample flow (0.75 m³/h
150 volumetric at ambient conditions) was measured by a mass flow meter (MASS-STREAM, M+W
151 instruments, Leonhardsbuch, Germany). The filter sampler was operated at least two times a day.

152 2.4 Upward-downward deposition rate sampler

153 It is important to compare the upward and downward rates to understand the turbulent and the
154 gravitational share in aerosol deposition rate measurement. Following an approach by Noll and Fang
155 (1989), it was assumed that turbulent transport is the main mechanism for upward-directed deposition
156 rate while turbulent transport and sedimentation are the mechanism of for the downward one. Therefore,
157 a sampler with an upward- and a downward-facing substrate in analogy to the Flat plate sampler was
158 designed. Air is flowing between two circular steel plates thick 1 mm with a diameter of 127 mm. In the
159 centers of the plates, two substrates are mounted opposite to each other. The substrate holders are
160 recessed, so that their adhesive collection surface is in plane with the steel surface. The construction is
161 mounted into a frame with a distance of 16 mm between the plates / substrates.

162 2.5 Ancillary Aerosol Data

163 Additional information regarding the aerosol particle size distributions has been obtained by using an
164 optical particle counter (OPC, GRIMM, Ainring, Germany), which is operationally available at the Izaña
165 Global Atmospheric Watch observatory (Bergamaschi et al., 2000; Rodríguez et al., 2015).

166 2.6 SEM-Analysis

167 All aerosol samples (except the filter sampler) were collected on pure carbon adhesive substrates (Spectro
168 Tabs, Plano GmbH, Wetzlar, Germany) mounted to standard SEM aluminum stubs. The filter samples
169 were stored in standard 'Petrislides' (Merck KGaA, Darmstadt, Germany). All adhesive samples were
170 stored in standard SEM storage boxes (Ted Pella Inc, Redding, CA, USA) in dry conditions at room
171 temperature. Individual particle analysis by automated scanning electron microscopy (SEM; FEI ESEM
172 Quanta 400 FEG, FEI, Eindhoven, The Netherlands; operated at 12.5 kV, lateral beam extension 3 nm
173 approx., spatial resolution 160 nm) was used to characterize particles for size and composition. A total
174 of 316,000 particles from six samplers was analyzed. 26 samples from BSNE (53,000 particles), 23
175 samples from MWAC (49,000), 23 samples from SIGMA-2 (39,000), 18 samples from Flat plate (12
176 mm) (24,000), 22 samples from Flat plate (25 mm) (21,000), 13 samples from Filter (80,000) and 12

177 samples from FWI-12 mm (50,000) were analyzed. Each sample was characterized at areas selected by
178 a random generator, until a total of 3,000 particles with projected area diameters greater than 1 μm was
179 reached. For particle identification, the backscattered electron image (BSE-image) has been used, as dust
180 particles contain elements with higher atomic number than carbon and therefore appear as detectable
181 bright spots in the BSE-image.

182 Chemistry information was derived by energy-dispersive X-ray analysis (EDX; Oxford X-Max 120,
183 Oxford Instruments, Abingdon, United Kingdom). The internal ZAF-correction of the detector / software
184 system – based on inter-peak background radiation absorption measurements for correction – was used
185 for obtaining quantitative results.

186 2.7 Particle size determination

187 The image analysis integrated into the SEM-EDX software determines the size of particles as a projected
188 area diameter.

$$189 \quad d_g = \sqrt{\frac{4B}{\pi}} \quad (1)$$

190 Where B and d_g are the area covered by the particle on the sample substrate and the projected area
191 diameter respectively.

192 Following Ott et al. (2008), the volumetric shape factor, S_v is determined from the count data as:

$$193 \quad S_v = \frac{P^2}{4\pi A} \quad (2)$$

194 Where P and A are the perimeter and the projected area of the particle respectively.

195 The volume-equivalent diameter (sphere with the same volume as the irregular shaped particle) is then,
196 calculated from the projected area diameter via the volumetric shape factor (Ott et al., 2008) and is
197 expressed by particle projected area and perimeter as

$$198 \quad d_v = \frac{4\pi B}{P^2} d_g = \frac{1}{P^2} \sqrt{64\pi B^3} \quad (3)$$

199 The aerodynamic diameter (d_a) is calculated from projected area diameter through the use of a volumetric
200 shape factor and aerodynamic shape factor (Wagner and Leith, 2001b)

$$201 \quad d_a = \sqrt{[d_v (\rho_p / \rho_0) 1 / S_d]} \quad (4)$$

202 With S_d the aerodynamic shape factor; ρ_p and ρ_0 are particle density and unit density respectively. For
203 this work, a value of $S_d = 1.41$ was used (Davies, 1979). Cunningham's slip correction was neglected in
204 this study, as all particles considered were super-micron size.

205 **2.8 Mass and number deposition rate calculation**

206 The mass deposition rate (MDR) and number deposition rate (NDR) are calculated from deposited
207 particle numbers per area, individual particle size and, in case of MDR, density. The particle density was
208 assumed to be equal the bulk material density of the dominating identified compound for each particle
209 (Kandler et al., 2007). A window correction (Kandler et al., 2009) was applied to the particle deposition
210 rate as:

$$211 \quad C_w = \frac{w_x w_y}{(w_x - d_p)(w_y - d_p)} \quad (5)$$

212 Where w_x and w_y are the dimensions of the analysis rectangle.

213 The MDR of the samples is then determined as

$$214 \quad MDR = \frac{1}{Ati} \sum_k \rho d_p^3 C_w(d_p, k) \quad (6)$$

215 Similarly, The NDR of the samples is determined as

$$216 \quad NDR = \frac{1}{Ati} \sum_k C_w(d_p, k) \quad (7)$$

217 With A is the total analyzed area, t is the sample collection time, ρ particle density and k is index of the
218 particle.

219 Size distributions for all properties were calculated for the logarithmic-equidistant intervals of 1-2 μm ,
220 2-4 μm , 4-8 μm , 8-16 μm , 16-32 μm , and 32-64 μm .

221 **2.9 Modeling atmospheric concentrations and size distributions from flux measurements**

222 Concentrations are calculated from the deposition rate using different deposition velocity models for
223 different samples, namely the models of Stokes and Piskunov (Piskunov, 2009). The basic relationship
224 between concentration and deposition rate was already given by Junge (1963), as the ratio of deposition
225 rate to concentration:

$$226 \quad V_d = F/C \quad (8)$$

227 With F is deposition rate and C is concentration. Note that the formulation is independent of the type of
228 concentration, i.e. it can be equally applied to number or mass concentrations.

229 All different approaches now give different formulations for the deposition velocity, based on a set of
230 assumptions and neglections.

231 2.9.1 Stokes settling

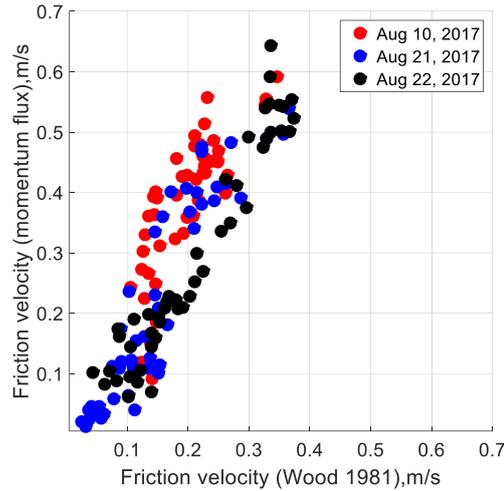
232 Terminal settling velocity (V_{ts}) is calculated according to Stokes' law.

$$233 \quad V_{ts} = \frac{d_p^2 g (\rho_p - \rho_a)}{18\mu} \quad (9)$$

234 Where d_p is the particle size, g is the gravitational acceleration (9.81 m/s^2); ρ_p the density of particle; ρ_a
235 the air density; μ is the dynamic viscosity of air ($1.8 \cdot 10^{-5} \text{ kg/(ms)}$).

236 2.9.2 Turbulent deposition and more complex deposition models

237 To calculate the turbulent impaction velocity, which depends of the wind speed, the friction velocity is
238 needed. Friction velocity (u_*), which is a measure of wind generated turbulence is one most important
239 variables affecting deposition velocity (Arya, 1977). Mainly two different approaches have been used to
240 estimate u_* . On one hand the momentum flux or the eddy covariance (EC) approach (Ettling, 1996),
241 which directly estimates u_* from the correlations between the measured horizontal and vertical wind
242 velocity fluctuation, and on the other the law of the wall (LoW) approach (Shao et al., 2011), which
243 estimates u_* from the wind profile. The latter can be approximated from free-stream velocity and
244 roughness assumptions (Wood, 1981), where the flow inside the sampler is assumed to be in the
245 hydraulically smooth regime (Schlichting, 1968). **Figure 1** shows correlations between u_* estimated
246 using Wood (1981) and Ettling (1996) approaches. Obviously, the approaches lead to different results,
247 for which no clear explanation is available (Dupont et al., 2018).



248
 249 **Figure 1:** Comparison of the friction velocities obtained from the momentum flux and the [Wood 1981](#)
 250 approach for different days with different wind speeds (average wind speed =2.9 m/s, 2.1 m/s, 3.1
 251 m/s for Aug 10, Aug 21, and Aug 22, 2017, respectively).

252 For the current work, the friction velocity is calculation is based on Wood (1981) approach:

$$253 \quad u_* = (u/\sqrt{2}) [(2\log_{10}(Re) - 0.65)^{-1.15}] \quad (10)$$

254 Where Re is the flow Reynolds number at the sampling stub location and is given as

$$255 \quad Re = uX/V \quad (11)$$

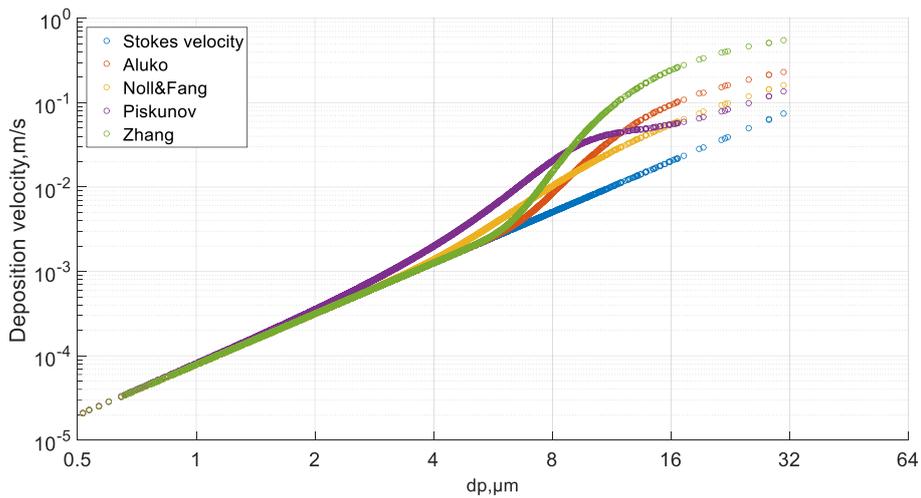
256 X is the distance from the lower plate edge to the center of the sampling stub (6.3 cm) and V is kinematic
 257 viscosity.

258 The reason why we opted to use the Wood (1981) over the Ettlign (1996) approach is a) its simplicity, as
 259 it requires only average wind speeds instead of 3D high resolution ones, and therefore will be more
 260 commonly applicable; and b) the fact that the momentum approach yields sometimes uninterpretable
 261 data, in particular in case of buoyancy-driven flow. For some case studies, both approaches are compared
 262 below.

263 There are a variety of models estimating the particles deposition speed (Aluko and Noll, 2006; Noll and
 264 Fang, 1989; Noll et al., 2001; Piskunov, 2009; Slinn and Slinn, 1980; Wagner and Leith, 2001a) (see
 265 **Figure 2**). And these different deposition velocity models yield different results, which could be due to
 266 negligence of unaccounted forces ([Lai and Nazaroff, 2005](#)) or due to the way how friction velocity is

267 determined or can be related to suppositions by different models (Kandler et al., 2018). Unless otherwise
268 stated, the particle density used in deposition velocity calculation is 2600 kg/m^3 .

269 It can be noted that a particular deposition model therefore may not be suitable in different cases for
270 describing the deposition velocity precisely, so as a result concentrations derived from deposition rate
271 measurements are likely to be biased (Giardina and Buffa, 2018; Kandler et al., 2018).



272
273 **Figure 2:** Deposition velocities for single particles to a smooth surface (Flat plate sampler) calculated
274 by using set of different classical deposition models for Tenerife samples (Aug 9, 2017; average wind
275 speed =3.0 m/s).

276 2.9.3 Deposition models applied to the samplers

277 **Table 1** shows the different deposition velocity models applied to the various samplers. The Piskunov
278 deposition velocity model is made for flat surfaces, and therefore it is applied to BSNE and Flat plate,
279 where deposition occurs to such surfaces. For the Sigma-2 sampler, it is assumed that each particle settles
280 with the terminal settling velocity (Tian et al., 2017), and therefore Stokes' velocity was used for
281 calculation of concentrations. In the case of MWAC, a different approach was required due to its
282 geometry. It is internally in principle an impactor design with the incoming tube pointing at the substrate,
283 but is operating at very low flow speed and therefore Reynolds numbers. As a result, it cannot be
284 described by the impactor theory only. Therefore, we assumed that the deposition velocity can't become
285 smaller than the one prescribed by the Piskunov model. As a result, we derived a velocity model based
286 on wind speed (or a reduced wind speed) and calculated the collection efficiency assuming the MWAC

287 to act as impactor for particles in the range of the cut-off diameter and larger. For smaller particles, we
 288 assumed that flow is like a flow over a smooth surface, so the Piskunov deposition velocity model was
 289 applied (e.g., as soon as the deposition velocity from impactor considerations becomes smaller than the
 290 Piskunov one, the latter was used).

291 **Table 1:** A summary of different deposition velocity models applied to the samplers

Sampler	Deposition velocity model
Sigma-2	Stokes' velocity
Flat plate	Piskunov
BSNE	Piskunov
MWAC	Combination of Piskunov and Impaction curve

292

293 **2.10 Determining the size distributions for mass concentration from the free-wing impactor**
 294 **measurements**

295 Considering the windows correction and the collection efficiency dependence on the impaction speed
 296 and geometry, the overall collection efficiency E is calculated according to Kandler et al. (2018). After
 297 calculating the collection efficiency, the atmospheric concentration is calculated from deposition rate and
 298 deposition velocity as

299
$$C = \frac{M}{V_d} = \frac{M}{E V_{imp}} \quad (12)$$

300 With E being the collection efficiency and v_imp the impaction velocity, calculated from ambient wind
 301 speed and rotation speed.

302 **2.11 Determining the size distributions for mass concentration from the filter sampler**
 303 **measurements**

304 Apparent number concentrations are determined from the particle deposition rate and the volumetric flow
 305 rate calculated from the mass flow for ambient conditions. The inlet efficiency (Eff_{in}) – accounting for
 306 the difference in wind speed and inlet velocity - is calculated as a function of Stokes number (Stk)
 307 (Belyaev and Levin, 1974). The ambient concentration N_{out} is calculated by dividing the measured
 308 number concentration by the inlet efficiency.

309 **2.12 Statistical uncertainty**

310 Owing to the discrete nature of the particle size measurement, the uncertainty coming from counting can
311 pose a significant contribution to the uncertainty of mass deposition rate measurement (Kandler et al.,
312 2018). It is, therefore, important to assess the uncertainties in our mass deposition rate measurements,
313 which is done in accordance to the previous work (Kandler et al., 2018). For the mass deposition rate,
314 the statistical uncertainty is assessed by a bootstrap simulation approach using Monte Carlo
315 approximation (Efron, 1979). In this work, the bootstrap simulations and the two-sided 95 % confidence
316 intervals calculation were performed by using Matlab's bootstrap function (MATLAB R2016a
317 (MathWorks, Inc). Here, MATLAB function uses a non-parametric bootstrap algorithm (Neto, 2015) to
318 compute the 95% bootstrap confidence interval.

319 **3 Computational fluid dynamics (CFD) simulation**

320 Computational fluid dynamics (CFD) simulations were conducted to predict the deposition of particles
321 on to different passive samplers (MWAC, Sigma-2 and Flat plate). A discrete phase model without
322 interaction with continuous phase was used to calculate the trajectories of the particles. The CFD software
323 ANSYS-FLUENT 18.2 was used for performing the numerical simulations.

324 **3.1 Evaluating the mean flow field**

325 In a first step the geometry of samplers was created using ANSYS DesignModeler. In a second step, an
326 enclosure around the geometry was generated. To ensure that there are no large gradients normal to the
327 boundaries at the domain boundary, the domain was created depending on the width, the height and the
328 length of the geometries. The space in front of the geometry is two times the height of the sampler, the
329 space behind the sampler is ten times the height, the space left and right of the geometry is five times the
330 width of the geometry and the space below and above the sampler is five times the height.

331 Afterwards a mesh was created using the ANSYS Meshing program. For the enhanced wall treatment
332 the first near-wall node should be placed at the dimensionless wall distance of $y^+ \approx 1$. The dimensionless
333 wall distance is given by

$$334 \quad y^+ = \frac{u_* y}{\nu} \quad (13)$$

335 With y the distance to the wall, ν the kinematic viscosity of the fluid and u_* the friction velocity which
336 is defined for this purpose by

$$337 \quad u_* = \sqrt{\tau_w / \rho} \quad (14)$$

338 With τ_w the wall, shear stress and ρ the fluid density at the wall. The wall is then subdivided into a
339 viscosity-affected region and a fully turbulent region depending on the turbulent Reynolds number Re_y

$$340 \quad Re_y = \frac{\rho y \sqrt{k}}{\mu} \quad (15)$$

341 With y the wall-normal distance from the wall to the cell centers, k the turbulence kinetic energy and μ
342 the dynamic viscosity of the fluid. If $Re_y > 200$ the k-epsilon model is used. $Re_y < 200$ the one-equation
343 of Wolfstein is employed (Chmielewski and Gieras, 2013; Fluent, 2015). The flow field was calculated
344 by solving the Reynolds Averaged Navier Stokes's equations with the software ANSYS Fluent. Standard
345 k-epsilon model was used to calculate the Reynolds-stresses. The boundary conditions at the sides of the
346 domain were set to symmetric. The inlet boundary condition was set to 2, 4 or 8 m/s with air as fluid
347 (Density: 1.225 kg/m³, viscosity: 1.7849*10⁻⁵ kg/(ms)). The outlet boundary condition was set to pressure
348 outlet.

349 The turbulence intensity T_i was calculated as

$$350 \quad T_i = \frac{\left(\frac{2}{3}k\right)^{1/2}}{v} \quad (16)$$

351 With k the turbulence intensity and v the velocity at the inlet of the domain.

352 Detail of the sampler construction and geometry are found in the electronic supplement (see **Figure S**
353 **18, S 19 and S 20**). Different cases were calculated for the Flat plate sampler (deposition area of 12 and
354 25 mm), for the Sigma-2 and for the MWAC (**Figure 3**). For the Flat plate, a mesh with 3,920,000 cells
355 was generated, for the Sigma-2 one with 7,600,000 cells and for the MWAC one with 4,620,000. After
356 the meshing, the flow fields were calculated. **Figure 3** shows as example the velocity magnitude in the
357 middle of the domain for a velocity of 4 m/s at the inlet.

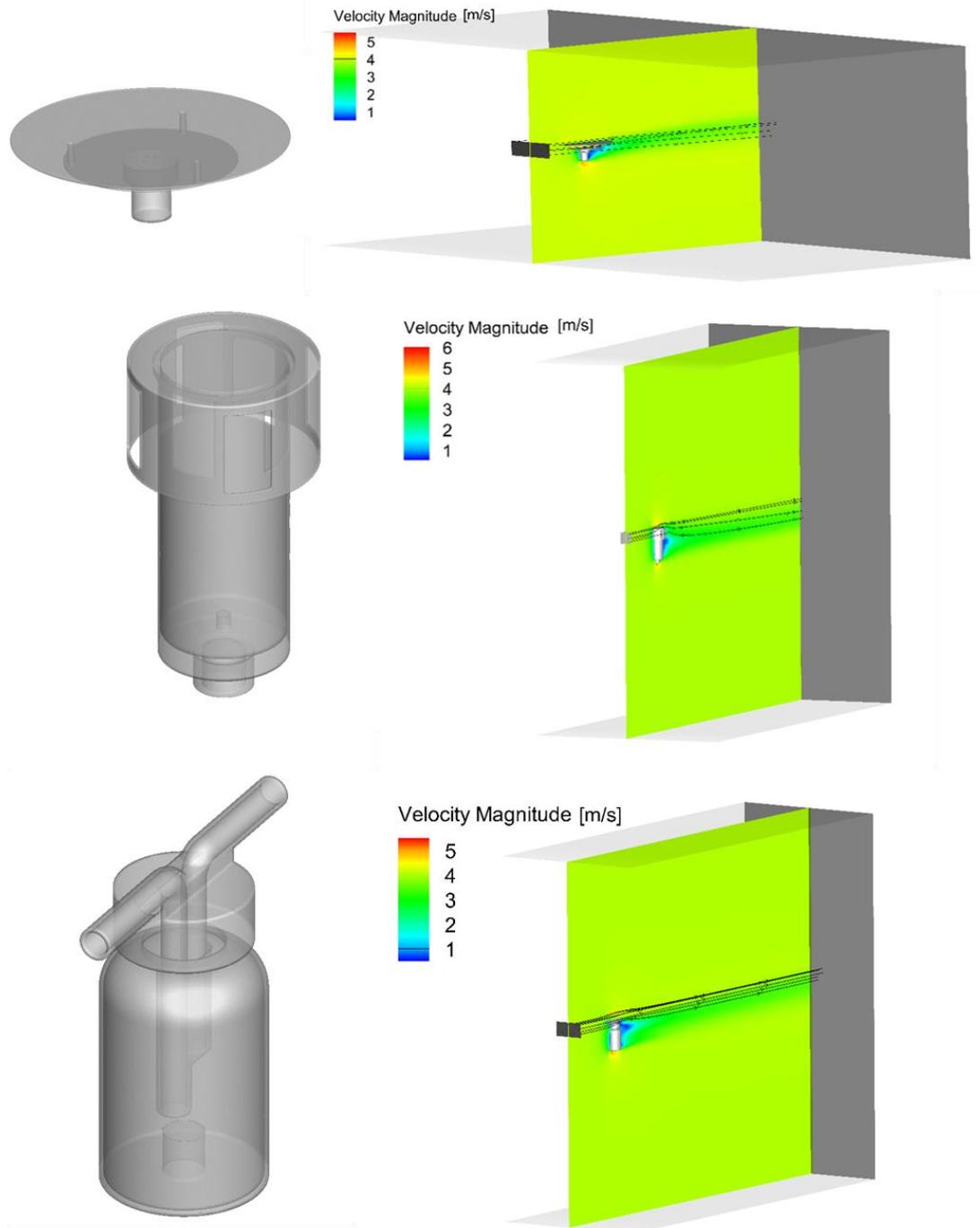
358 In the last step, particles were injected into the velocity field and their trajectories computed. For all
359 samplers, the deposition area boundary condition was set to "trap" and the walls were defined as
360 reflecting boundaries. Different particle sizes (1, 2.5, 5, 10, 20 and 50 μ m, Stokes' diameter) for three
361 different wind speeds (2, 4, 8 m/s) were investigated. The particles density was set to a value of 2600
362 kg/m³ to match an approximate dust bulk density. The particle concentration was 4*10⁸ /m² in all cases,
363 while the injection area was adjusted to the geometries (**Figure 3**).

364 The number of particles trapped in the deposition area was determined. The deposition velocity V_d was
365 calculated by

366
$$V_d = \frac{N_{pt}v}{A_d C_p} \quad (17)$$

367 with N_{pt} the number of trapped particle at the deposition area, v the velocity of the air at the inlet
 368 boundary of the domain, A_d the deposition area and C_p the particle concentration at the particle injection
 369 area (Sajjadi et al., 2016).

370

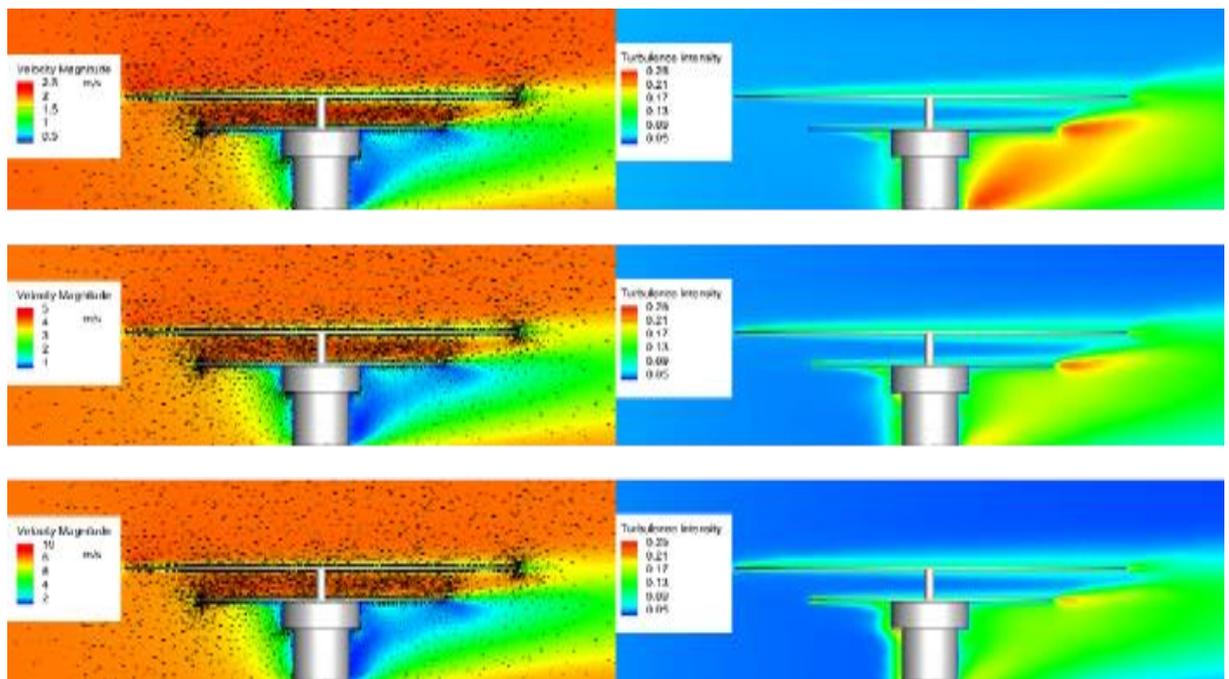


371

372 **Figure 3:** Geometries of Flat plate sampler (top), Sigma-2 sampler (middle), MWAC sampler (bottom).
 373 CFD modeling domain and velocity magnitude, inlet velocity: 4 m/s (right); in addition, the injection
 374 area is shown in black (Flat plate sampler: width 0.2 m, height 0.05 m; Sigma-2-sampler: width 0.2 m,
 375 height 0.1 m; Bottle sampler: width 0.1 m, height 0.05 m) along with exemplary streamtraces.

376 3.2 Velocity contours and vectors for the samplers

377 For the Flat plate sampler, stream velocities and turbulence intensities are shown in **Figure 4**. The
 378 formation of the boundary layer at the wall of the sampler is clearly visible at all velocities. At the central
 379 sampling location, the flow between the plates has the same velocity as the free stream, so for the
 380 analytical deposition models, the lower plate can be treated as single surface. The highest velocity is
 381 found at the sharp edge at the bottom of the sampler. Due to the high velocity gradients in this part there
 382 is also the highest turbulence intensity in the domain. As expected, the turbulent wake becomes smaller
 383 with increasing wind speed.



384
 385 **Figure 4:** Flat plate Sampler: Velocity magnitude and turbulence intensity at wind speed 2 m/s (top),
 386 Flat plate Sampler: Velocity magnitude and turbulence intensity at wind speed 4 m/s (middle), Flat
 387 plate Sampler: Velocity magnitude and turbulence intensity at wind speed 8 m/s (bottom).

388 3.2.1 Sigma 2 Sampler

389 The cross section of the velocities for the Sigma-2 are shown for the 4 m/s case in **Figure S 21** in the
 390 electronic supplement. Apparently the velocity magnitude inside the sampler is much smaller than

391 outside. In the vertical settling tube, the turbulence intensity is low, justifying the idea of Stokes settling
392 inside. Owing to the open, but bulky geometry, there is a flow into the interior at the back. The highest
393 velocities and turbulence intensities are found at the sharp edges at the top and bottom of the sampler.

394 **Figure S 22** (in the electronic supplement) shows the cross section of the velocities for the MWAC in
395 the 4 m/s case. Furthermore, the velocity field and the velocity vectors in the cross sections across and
396 along the inlet tube are shown in **Figure S 23** in the electronic supplement. In the tubes the typical pipe
397 flow is formed. In the figures showing the cross sections along the inlet tube a symmetrical flow over the
398 pipe cross section is visible. Finally, **Figure S 24** (in the electronic supplement) shows the mean flow
399 velocity in the MWAC tube is shown as a function of the outside velocity for the three cases. The fitting
400 curve shows that the mean velocity in the pipe increases linearly with the external velocity.

401 4 Results and Discussion

402 4.1 Methodical aspects (Field Measurements)

403 4.1.1 Mass deposition rate comparison

404 Mineral dust was the dominating particle type during this campaign, consisting of different silicates,
405 quartz, calcite, dolomite, gypsum, similar to previous findings for this location (Kandler et al., 2007).
406 Therefore, hygroscopicity was not taken into account, as due to the mostly non-hygroscopic compounds
407 and the moderate humidities their impact was rated low. Details on the composition will be reported in a
408 companion paper.

409 The mass and number deposition rates (given per unit time and sample surface area) along with daily
410 average temperature and wind speed are presented as daily values. Details for all days and all samplers
411 can be found in the electronic supplement (see **Table S 1, S 2, S 3 and S 4**). All data shown in this section
412 are calculated from SEM measurements. Particle sizes are reported as aerodynamic diameter, if not
413 otherwise stated.

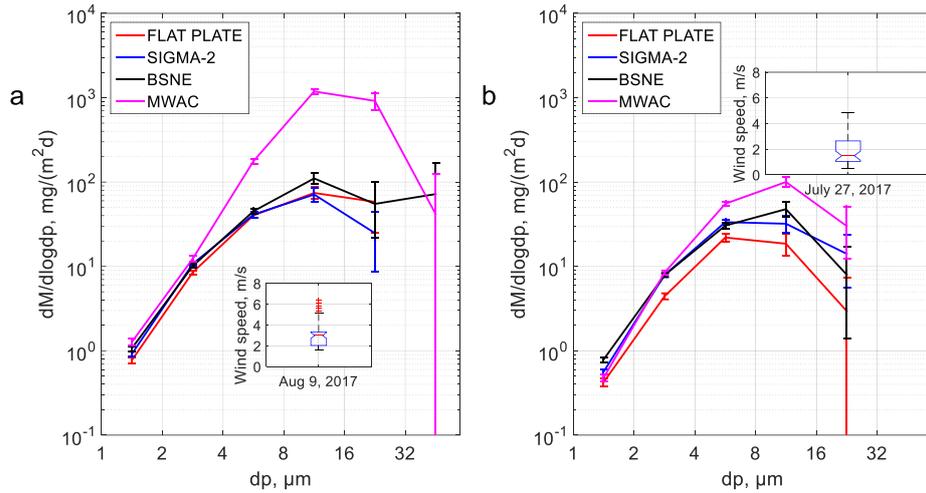
414 **Figure 5** shows as example mass deposition rates for different samplers during a dust event and a non-
415 dust event day. For all samplers, the mass deposition rate size distributions peaked in the 16-32 μm
416 diameter interval. This result is in support of the conclusion that atmospheric dry deposition is dominated
417 by coarse particles owing to their high deposition velocities (Davidson et al., 1985; Holsen et al., 1991).
418 There is a considerable difference among different samplers affecting mainly the size range with the
419 highest [mass deposition rates](#), whereas the difference is small for smaller particles. MWAC and BSNE –
420 both horizontal flux samplers – collect coarser material than the Flat plate and Sigma-2 samplers, which

421 in contrary measure the vertical flux. In particular, the MWAC sampler exhibits considerably higher
422 coarse particle mass deposition rates, probably owing to its impactor-like design.

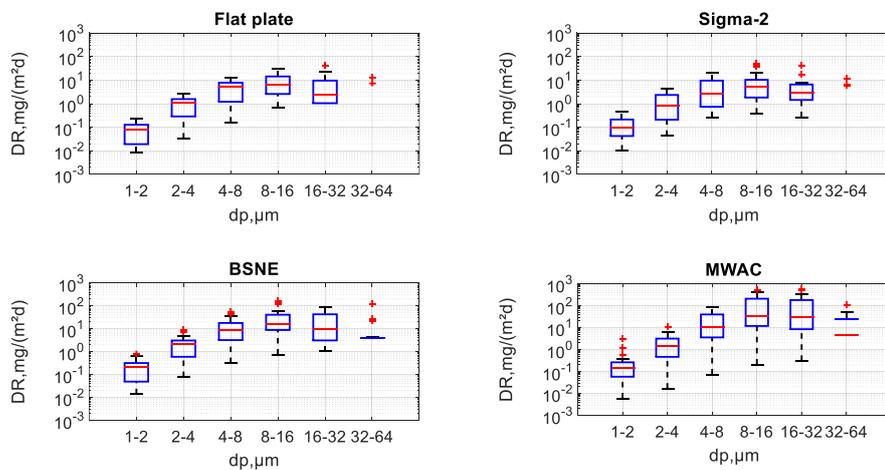
423 **Table 2:** The campaign maximum and minimum and median mass deposition rates measured by the
424 samplers

Samplers	Maximum deposition rate (mg/ (m ² d))	Minimum deposition rate (mg/ (m ² d))	Median deposition rate (mg/ (m ² d))
MWAC	1240	0.6	4.8
BSNE	310	0.2	3.1
Flat plate	80	2.0	1.1
Sigma-2	117	1.9	1.1

425
426 As consequence the vertical flux instruments collect much less material than the horizontal flux ones
427 (**Table 2**), which is in accordance with previous findings (Goossens, 2008). In the present study,
428 horizontal to vertical flux mass ratio is approximately between 2.8 and 4.4 (with single size intervals
429 ranging between 2 and 50), while Goossens (2008) reported it to be in between 50 and 160. This
430 difference in the ratio might come from the different approaches. Goossens (2008) used water as a
431 deposition surface while in our study we used a SEM sampling substrate. Furthermore, from **Figure 5**,
432 we can clearly see that that there is a strong variation in mass deposition rates between dust event days
433 and non-dust event days (full dataset is shown in **Figure 6**). Generally, the temporal variation is higher
434 than the difference between the samplers so a strict comparison between this and the previous study can't
435 be done.



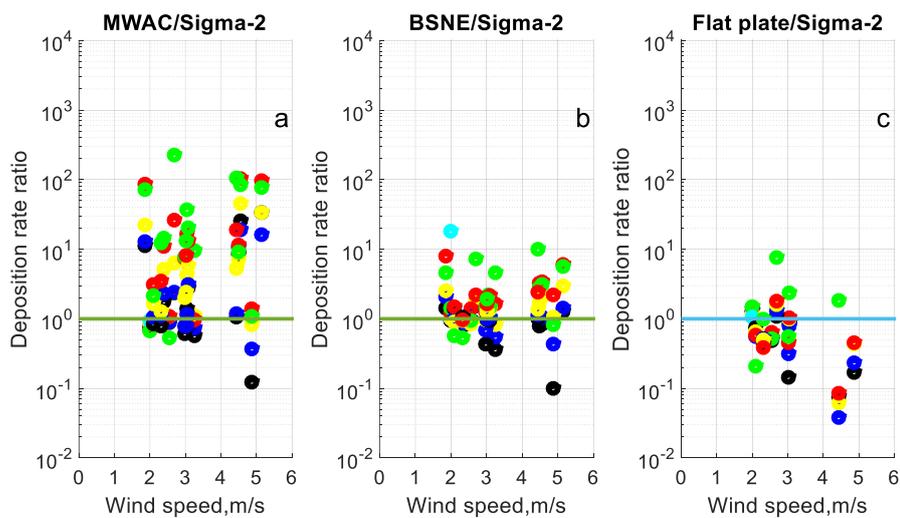
436
 437 **Figure 5:** Size resolved mass deposition rate measured by different passive samplers: a) dust event day;
 438 b) non-dust event day. Data are derived from SEM measurements. Error bars show bootstrapped 95%
 439 confidence interval. The inserts show box plots for the wind speed distribution based on 30-min
 440 intervals.



441
 442 **Figure 6:** box-plots of size resolved deposition rate (campaign data; Flat plate, Sigma-2, MWAC and
 443 BSNE samplers). On each blue box, the central mark is the median, the edges of the box are the 25th
 444 and 75th percentiles. The red vertical lines show the standard deviation.

445 From the structure of the deposition models, a wind speed dependency for the deposition velocity should
 446 be expected. The average wind speed during the campaign was about 3.5 m/s with the lowest daily median
 447 around 1.5 m/s and the highest 7 m/s. A daily box-plot of 30-min averaged wind speed at Izaña is shown
 448 in **Figure S 1** in the electronic supplement.

449 **Figure 7** shows the mass deposition rate ratio of MWAC, BSNE and Flat plate to Sigma-2 as function
 450 of wind speed. The Sigma-2 sampler was chosen for comparison, as due to its settling tube design, it is
 451 expected to have the least wind sensitivity. The results show highly scattered values. The collection
 452 efficiency of MWAC for large particles has an increasing tendency in comparison to Sigma-2 slightly
 453 with increasing wind speed, while there is barely a trend visible for the BSNE. Both – being horizontal
 454 flux samplers – collect considerably more material than the Sigma-2. For the Flat plate, the deposition
 455 velocity in relation to the Sigma-2 has a weak decreasing trend for higher wind speeds, but generally, the
 456 deposition speed is similar. Overall, the relation of Sigma-2 to BSNE shows the closest agreement, while
 457 the scatter is higher for the other combinations. [More information on the relation between the other
 458 instruments is shown in Figure S 2, S 3, S 4, S 5, S 6, S 7 and S 8 in the electronic supplement.](#)

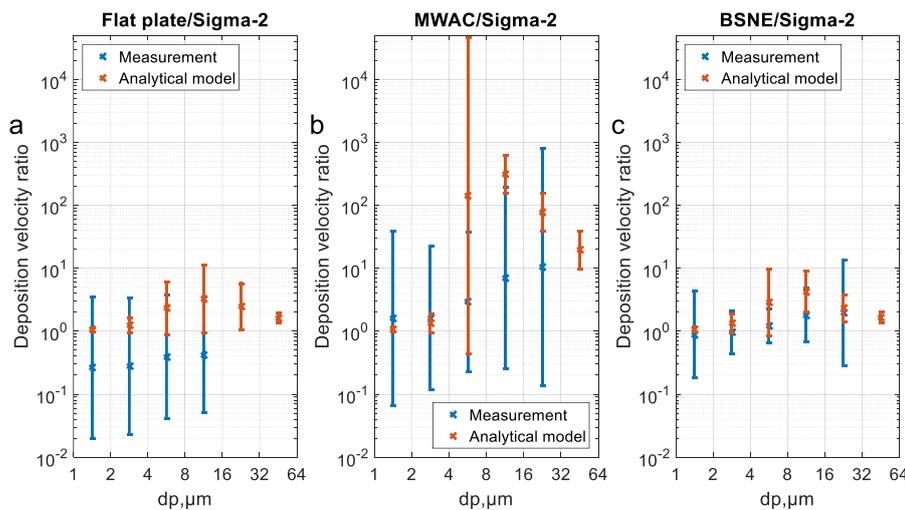


459
 460 **Figure 7:** Deposition rate ratio as function of wind speed for different days (MWAC/ Sigma-2 (a),
 461 BSNE/Sigma-2 (b) and BSNE/ Sigma-2 (c)). Different colors represent deposition rate measured in
 462 different size intervals (black: 1-2 μm ; blue: 2-4 μm ; yellow: 4-8 μm ; red: 8-16 μm ; green: 16-32 μm ;
 463 cyan: 32-64 μm).

464 The deposition rate ratios obtained from the measurements are identical to the deposition velocity ratios,
 465 when the sampling time and concentration are the same (eq. (7)); the latter is achieved by the close and

466 parallel sampling. Therefore, the experimentally determined ratios can now be compared to the
 467 deposition velocity ratios derived from the theoretical considerations. **Figure 8** shows the according
 468 comparison. Note that this consideration allows for the assessment of relative model performance and
 469 sampler efficiency, but lacking a ‘true’ reference, it does not allow for determining the most accurate
 470 sampler.

471 While for BSNE and Sigma-2 observation and model fit comparatively well, the deposition velocity is
 472 misestimated for the Flat plate/Sigma-2 pairing for all particle sizes (overestimate for Flat plate
 473 deposition velocity or/and underestimate for Sigma-2). For MWAC/Sigma-2, there is a clear size
 474 dependency, indicating that probably the impactor model overestimates the deposition velocity; the latter
 475 might be due to unaccounted particle losses (e.g., inlet efficiency). MWAC, BSNE and Sigma-2 agree
 476 with respect to deposition velocity better based on the measurement data than predicted by the theory. It
 477 may be connected to the non-stationarity of the atmosphere, which is not accounted for by the models,
 478 i.e. the permanent wind speed fluctuations smoothing out detail differences of a stationary flow. The Flat
 479 plate sampler, however, has a lower-than-predicted deposition velocity.



480

481 **Figure 8:** Comparison of geometric mean ratio of deposition velocities for different sampler pairs
 482 derived from measured deposition rates (blue) and from corresponding deposition models (orange).

483 (a) Flat plate/Sigma-2; (b) MWAC/Sigma-2; (c) BSNE/Sigma-2. Error bars show geometric standard
 484 deviations. According deposition models are listed in Table 1.

485 4.1.2 Dependence of PM₁₀ dust deposition on atmospheric concentration and wind speed

486 **Figure S 16** (in the electronic supplement) and **Table 3** display for the approximate PM₁₀ size range the
 487 correlation between **number deposition rates**, atmospheric particle number concentrations measured by
 488 OPC and the wind speed for different samples. For this consideration, only the overlapping size range –
 489 1-10 µm aerodynamic diameter – was used. As expected, there is in all cases a positive correlation
 490 between concentrations and number deposition rates (see **Figure S 16** in the electronic supplement). In
 491 particular, for the BSNE and the Simga-2, robust correlations with a trend to underestimation at higher
 492 concentrations exist. While the models predict a positive correlation of wind speed and deposition rate,
 493 this is not observed in the measurements. Instead, a non-significant anti-correlation can be observed, if
 494 at all (e.g., for Flat plate; r^2 : 0.319, p-value = 0.070, slope=-0.261), indicating a cross-influence of wind
 495 speed and concentration. E.g. **higher concentrations of dust aerosol particles might be meteorologically**
 496 **linked to lower wind speeds due to a different transport situation**. Such a general behavior was observed
 497 previously for example by different techniques for a dust transport region (Kandler et al., 2011). An
 498 ambiguous wind-dependency has been reported for other places (Xu et al., 2016). In this study, the main
 499 driver of the deposition rate during is obviously the dust concentration.

500 **Table 3:** Summary of the regression analysis for the correlations between the dust deposition rate and
 501 the atmospheric concentrations (**PM₁₀ size range; measured by the OPC**), and for the correlations
 502 between the **dust** deposition rates and the wind speeds. Significant relationships are shown in bold.

	Deposition rate vs concentration			Deposition rate vs wind speed		
	r^2	p-value	slope (m/d)	r^2	p-value	Slope ($1.16 \cdot 10^5$ /(m ³))
Flat plate	0.600	0.0052	0.492	0.319	0.070	-0.261
MWAC	0.155	0.335	0.146	0.308	0.153	-0.157
BSNE	0.937	1.00*10⁻⁶	0.832	0.017	0.706	-0.052
Sigma-2	0.925	3.39*10⁻⁵	0.725	0.0125	0.775	-0.069

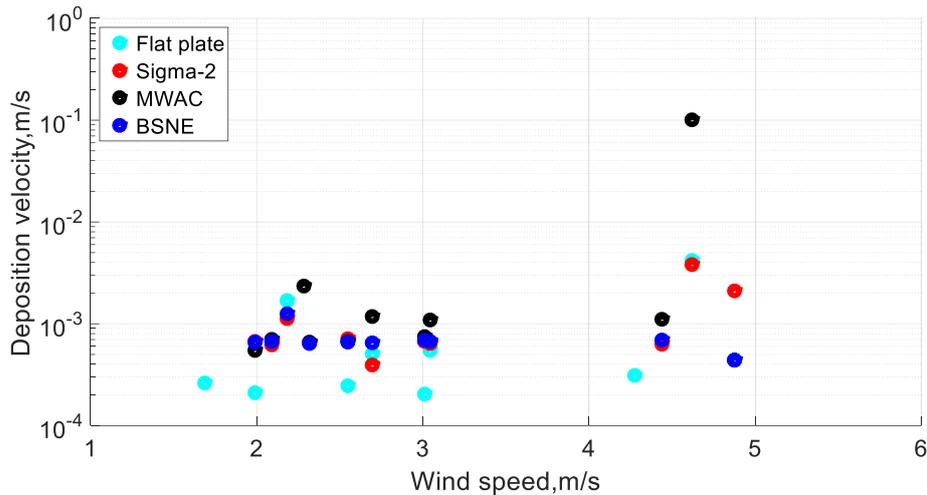
503
 504 In a second step it was tested, whether the application of each sampler's assigned deposition model can
 505 increase the correlation between the measurements by the deposition samplers and the OPC observations,
 506 i.e. whether the meteorological parameters accounted for in the models can decrease the deviation.
 507 Therefore, in analogy to the previous correlation, the concentrations modeled from each sampler's SEM

508 data were correlated with the OPC data for the size range between 1 and 10 μm in particle diameter (see
509 **Table S7** in the electronic supplement). However, no increase in correlation quality is observed,
510 indicating that – like already observed from the varying ratio calculations above – the deposition models
511 fail to describe the deposition behavior in detail.

512 From the correlation relations in **Table 3**, it can be learned that MWAC is least suitable for estimating
513 PM_{10} , which fully agrees well with previous studies (Mendez et al., 2016). However, the correlation
514 analysis here shows that BSNE is actually a suitable instrument for a PM_{10} estimation, which is in contrast
515 to the wind-tunnel observation of Mendez et al. (2016). This discrepancy might be owed to the different
516 approaches. While in the previous work the loss of concentration from the passing aerosol was measured,
517 in this study a gain of deposition was investigated. As result, for lower deposition velocities (discussed
518 below), the former approach will yield high uncertainties. Similar to BSNE, Flat plate and Sigma-2
519 appear good estimators for PM_{10} , which is in accordance with previous studies (Dietze et al., 2006).

520 **4.1.2.1 Size-resolved apparent deposition velocity in the PM_{10} size range**

521 **Figure 9** displays the apparent deposition velocity (calculated as the ratio of the number deposition rate
522 to the concentration of the OPC) as function of the wind speed. Obviously, also here there is not clear
523 trend. The apparent deposition velocities range between $2 \cdot 10^{-4}$ - 10^{-1} m/s. *As can be clearly seen from the*
524 *plot, the effect of wind speed on deposition velocity is negligible, as indicated already in Table 3. While*
525 *this is in contradiction to the models, one has to keep in mind that the (a) the observed wind speeds are*
526 *comparatively low here, and (b) the considered size range is not the one most affected by the wind speed.*
527 *An effect of the wind speed might therefore be much stronger at higher wind speeds and for larger*
528 *particles.*



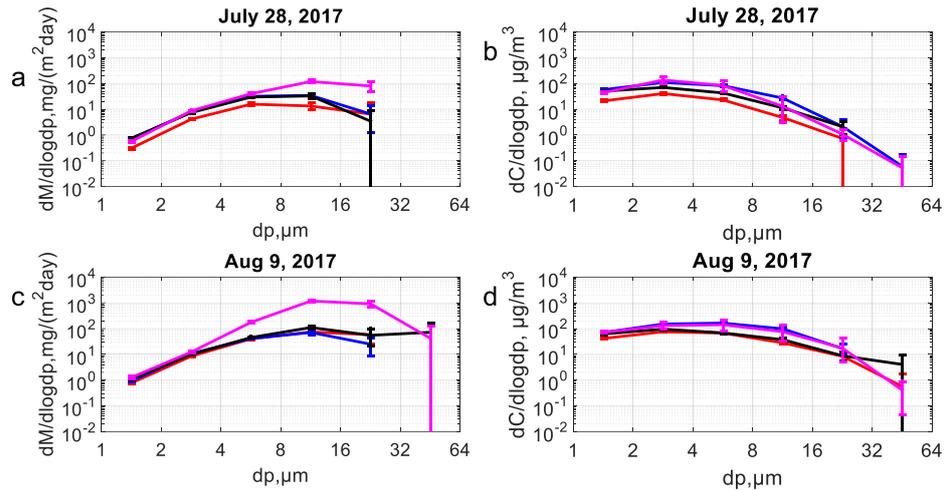
529
 530 **Figure 9:** Apparent deposition velocity: ratio of number deposition rate determined from SEM
 531 measurements to the number concentration observed by the OPC as function of wind speed. For the
 532 consideration, only the overlapping size range (approximately 1-10 μm) was used.

533 4.1.3 Atmospheric mass concentrations derived from deposition rates

534 **4.1.3.1 Consistency between samplers and corresponding models**

535 **Figure 10** compares a mass deposition rate size distribution with the according concentrations derived
 536 by the modeled deposition velocities. Calculating the mass concentrations from different passive
 537 samplers with different models leads in most cases to a better agreement between the measurements,
 538 taking into account the statistical uncertainties (see **Figure S 9** in the electronic supplement). This
 539 indicates that the deposition velocity models selected for the samplers are generally suitable, despite the
 540 deviations in single cases.

541



542

543 **Figure 10:** Comparing different samplers with respect to (a, c) dust mass deposition rates and (b, d)
 544 dust mass concentrations calculated from application of the corresponding model as function of particle
 545 size. Different colors indicate different samplers (red: Flat plate; blue: Sigma-2; black: BSNE and
 546 magenta: MWAC). Bars show the 95% confidence intervals. [According deposition models for](#)
 547 [concentration calculation are given in the Table 1.](#)

548 4.1.3.2 Size-resolved comparison with active samplers

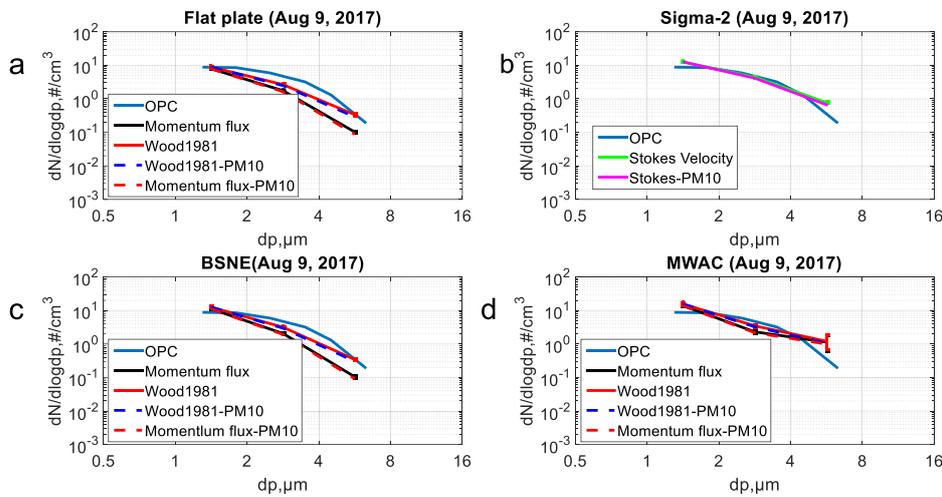
549 The calculated number concentrations in the size interval between 1–10 μm are compared with the
 550 concentrations measured using the OPC. Similarly, the mass concentration size distributions above the
 551 PM_{10} size range are validated using the FWI measurements.

552 **Figure 11** (see also **Figure S 11 and S 12** in the electronic supplement) shows a comparison of number
 553 concentration size distributions calculated from deposition rate measurements of [the Flat plate, Sigma-2,](#)
 554 [BSNE and MWAC](#) samplers with the number size distributions measured by the OPC for different days.
 555 Overall, the number concentrations obtained from OPC measurements are slightly higher than the ones
 556 from the deposition rates, mainly for the size range 2-5 μm and for dust days. This reflects the tendency
 557 of a relative underestimation of the concentrations by the passive techniques for high concentrations,
 558 which was already visible in the correlation analysis above. In general, **Figure 11** shows that the deviation
 559 of calculated values from OPC measured values is significant.

560 In this context, **Figure 11** shows also the low influence of the two techniques used for u_* estimation.

561 While [the number concentrations](#) derived using the friction velocity estimated from Wood (1981)

562 formulation are slightly higher and therefore in better agreement with the OPC data, the difference
 563 appears to be negligible in general, probably owing to the relatively low wind speeds in this study.



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Figure 11: Comparison of the number concentrations calculated from the deposition measurements with the number concentrations measured by the OPC. Number size distributions are obtained by converting the SEM number deposition rates to number concentrations using the different deposition velocity models (Table 1), in analogy to the mass size distributions. The blue curve shows the concentration measurements by the OPC. For the concentrations obtained from the number deposition rates, two different approaches for the friction velocity are shown. The black curve shows the concentration curve calculated using the momentum flux approach without PM₁₀ inlet correction (i.e. the atmospheric concentration). The red curve shows the concentration curve calculated using the Wood approach without PM₁₀ correction. The dotted blue curve shows the concentration curve calculated using the Wood approach with PM₁₀ inlet correction (simulating the concentration the OPC would supposedly measure). The dotted red curve shows the concentration curve calculated using the Momentum flux approach with PM₁₀ inlet correction. In case of the Sigma-2, the green curve shows the concentrations calculated using the Stokes velocity and the magenta curve the ones using Stokes velocity with the PM₁₀ inlet correction. Bars show the central 95 % confidence interval. (a): Flat plate; (b) Sigma-2; (c): BSNE; (d) MWAC. The average wind speed on Aug 9, 2017 was 3.0 m/s.

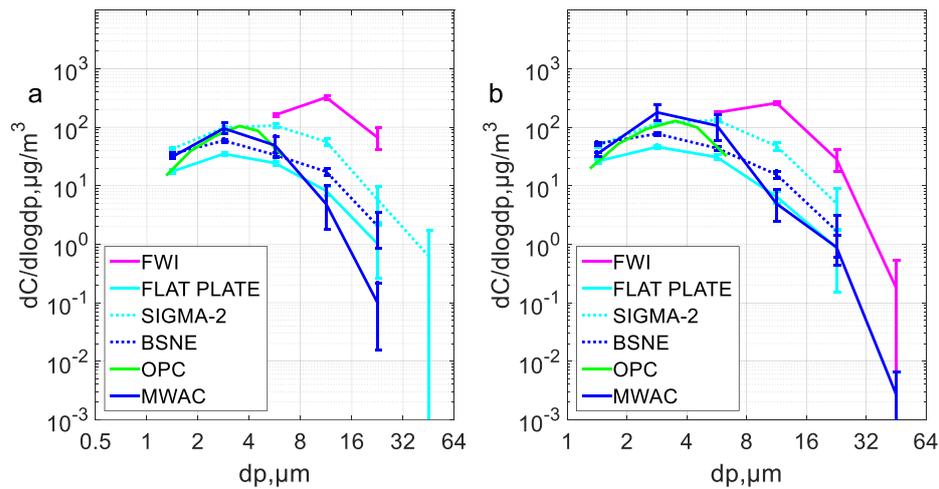
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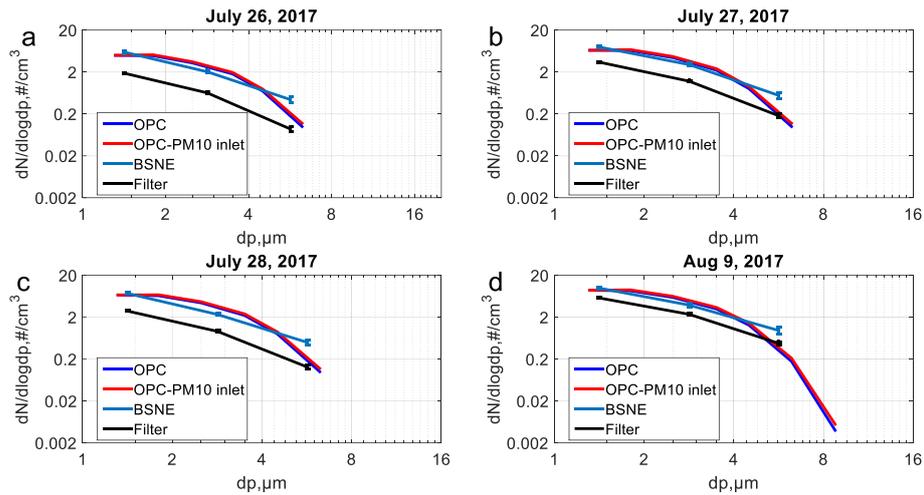
Figure 12 (see also **Figure S 10** in the electronic supplement) shows the comparisons for the larger particles between the deposition-derived number concentrations and the ones from the FWI. Here, a significant inconsistency occurs between the mass size distributions from passive samplers and the ones

583 from FWI. In particular, the size range larger than 10 μm seems to be generally underestimated by the
 584 passive samplers. While for particles around 10 μm , this could be partly to a badly-defined collection
 585 efficiency curve of the FWI ((Kandler et al., 2018); 50 % cut-off at 11 μm) and the according correction,
 586 this can't be the reason for the particles larger than 16 μm , where this efficiency approaches unity. Here,
 587 the deposition velocity for the samplers is apparently overestimated. [A possible explanation would be](#)
 588 [inlet losses of the passive sampler, but this need to be subject of further research.](#)



589
 590 **Figure 12:** Daily average mass size distributions obtained from the passive sampler techniques in
 591 comparison to an active sampler (FWI). [Mass concentration size distributions were calculated from the](#)
 592 [SEM mass flux measurements using the corresponding deposition velocity models.](#) Samples were
 593 collected on 26th of July (a) and 27th of July (b). The mass concentration measurements shown by the
 594 FWI are daily averages (3 samples per day). Error bars show the central 95% confidence interval.

595 In a last step, the deposition-derived concentrations are compared to these determined from the iso-axial
 596 filter sampler. **Figure 13** shows that, while the calculated size distributions are in good agreement with
 597 the OPC ones, the filter-derived seem to relatively underestimate the concentrations. A correlation
 598 analysis (R-squared: 0.681, p-value = 0.0854 and slope = 2.0394) suggests, that there is a weak positive
 599 correlation between calculated number concentration from filter samples and the OPC measured
 600 concentration. [The same comparison to other samplers is shown in Figure S 13, S 14 and S 15 in the](#)
 601 [electronic supplement. The reasons for this weak correlation – in particular in comparison to the ones](#)
 602 [from Sigma-2 and BSNE – remain unexplained by now.](#)



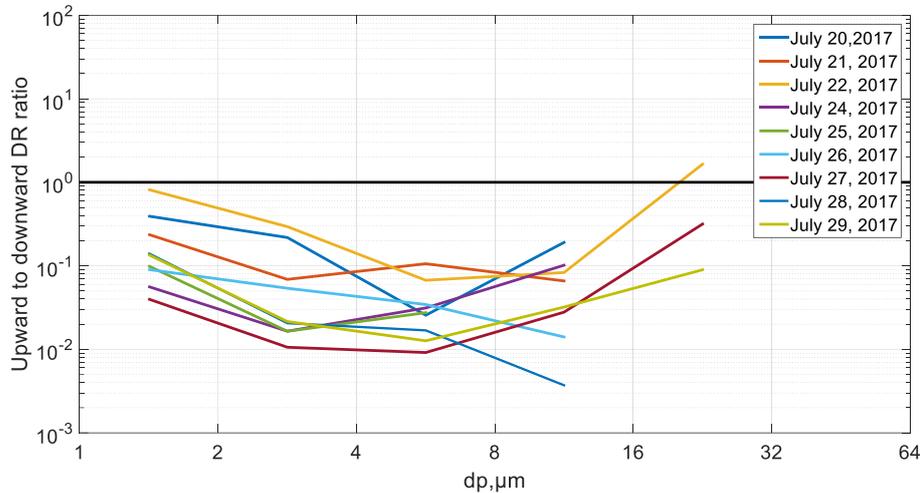
603

604 **Figure 13:** Number concentration size distributions obtained from the SEM analysis of the filter
 605 sampler, in comparison to BSNE and OPC for different measurement days (a: July 26, 2017; b: July 27,
 606 2017; c: July 28, 2017; d: July 29, 2017). [Number concentration size distributions from deposition are](#)
 607 [obtained by converting the SEM obtained number deposition rates to number concentration using](#)
 608 [different deposition velocity models.](#) The red curve shows the OPC with PM₁₀ inlet efficiency
 609 correction (representing the atmospheric concentration). The number concentration measurements
 610 shown for the filter sampler are daily averages. Error bars show the central 95% confidence interval.

611 4.1.4 Estimating the turbulent versus gravitational transport fraction

612 The size-resolved upward and downward deposition rates were derived from the upward-/downward
 613 facing deposition sampler [by the same type of SEM analyses.](#) Results of the size resolved-mass and
 614 number deposition rate measurements along with daily average temperatures and wind speeds are given
 615 in the electronic supplement (see **table S 5 and S 6**). The upward deposition rate is always less than the
 616 downward deposition rate. This is expected because the upward facing substrate (i.e. measuring the
 617 downward-directed deposition rate) collects particles deposited by gravitational settling and turbulent
 618 inertial impaction, while the downward facing substrate (for the upward-directed deposition rate) collects
 619 particles only by means of turbulent impaction. **Figure 14** shows the ratio of upward to downward mass
 620 deposition rate as function of particle size. The deviation is greatest for the particle size range around 8
 621 μm, which is strongly affected by turbulence (Noll and Fang, 1989). However, nearly no trend of
 622 increasing ratio with increasing wind speed can be found here (see **Figure S 17** in the electronic
 623 supplement). Besides the wind speed magnitude, different properties were calculated from the sonic wind

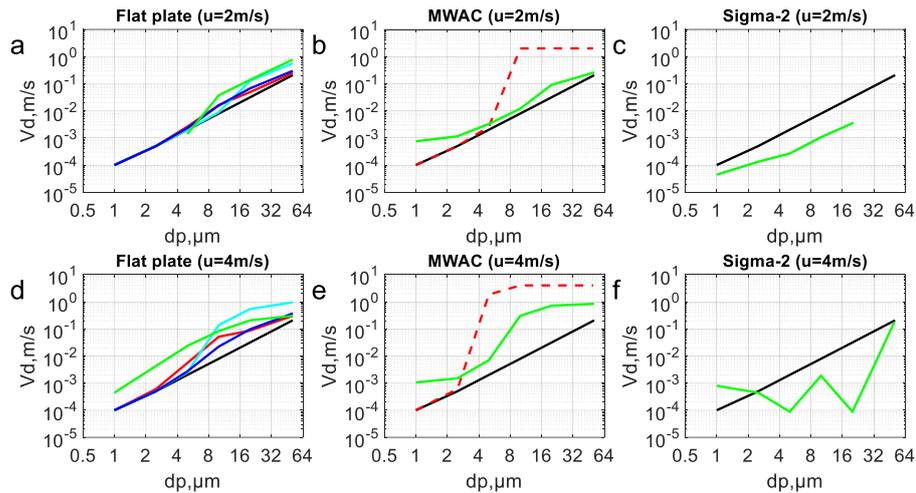
624 data (e.g., turbulent intensity, Monin-Obukhov length, relative standard deviation of wind speed, average
625 vertical component), but none of them was able to explain the observed variations in the deposition rate
626 ratio.



627
628 **Figure 14:** Ratio of upward- to downward-directed mass deposition rate as function of particle size.
629 The deposition rate is measured using the upward-downward-Flat plate sampler (with 25 mm stub).

630 4.2 Computational fluid dynamics (CFD) simulation

631 Using computational fluid dynamics (CFD), deposition velocities of particles for different passive
632 samplers were predicted and compared to the analytical deposition velocity models used for the different
633 samplers (see **Figure 15** and **Figure S 25** in the electronic supplement). While for the Flat plate and
634 MWAC sampler the curves agree qualitatively (i.e. showing deposition speeds higher than Stokes
635 velocity at particles sizes 4-16 μm , which are supposedly strongest affected by turbulence), for the Sigma-
636 2, they are largely contrary except for the lowest wind velocity. The latter might be owed to the fact that
637 in a flow model, the non-omnidirectional construction of the Sigma-2 might lead to preferred airflows,
638 which are not relevant in a more variable and turbulent atmosphere. However, also for the former ones,
639 the deposition velocity curves are quantitatively largely different. **Figure S 26** (in the electronic
640 supplement) shows a comparison of the CFD-derived particle deposition velocities at different wind
641 speed values for different samplers.

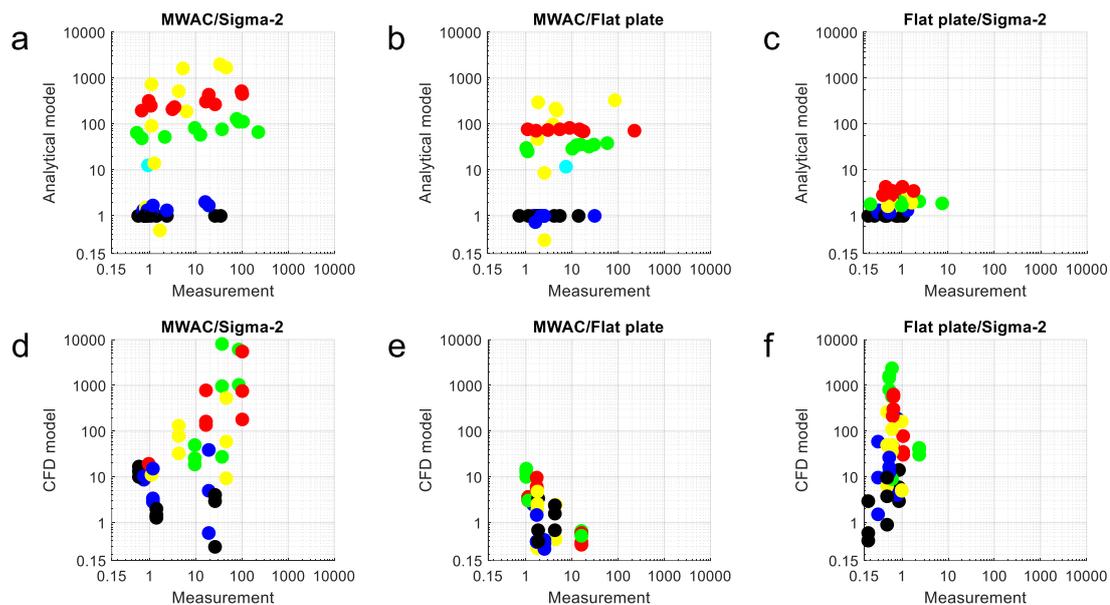


642

643 **Figure 15:** Deposition velocities calculated for different samplers by analytical and CFD approaches.
 644 The red curve shows the deposition velocity calculated using the Piskunov model, the dotted red curve
 645 shows the combination of the Piskunov and the impaction curve model, the black curve shows the
 646 Stokes deposition velocity, the blue curve the Noll and Fang model, the cyan the Zhang model, and the
 647 green curve finally the deposition velocity from CFD. Panels **a-c** are calculated for 2 m/s wind speed,
 648 **d-f** for 4 m/s.

649 4.3 Comparison of measured deposition rate ratios to analytically and CFD modeled ones

650 As there is no reference instrument for dry deposition sampling, the separate approaches are compared
 651 in a relative way. **Figure 16 a-c** show comparisons of the deposition velocity ratios derived from the
 652 analytical models with the according measured deposition velocity ratios (equalling the according
 653 deposition rate ratios), **d-f** the respective correlation of the ratios derived from CFD modeling with the
 654 measurement. As the CFD models could only be calculated for a limited number of flow velocities,
 655 deposition velocity values were interpolated between the calculated cases. Generally, the agreement is
 656 very poor. Practically, no variation observed in the measurement data can be explained by model
 657 variation, independently of the type of model. While this might be explained to a smaller extent by the
 658 propagating measurement uncertainties for the largest particles with low counting statistics, for the
 659 smaller ones this systematic deviation must have other reasons.



660
 661 **Figure 16:** Comparison of the observed deposition velocity ratios with modeled ones by the analytical
 662 deposition models (upper row, a-c) and by the CFD models (lower row, d-f). (a, d) MWAC/Sigma-2;
 663 (b, e) MWAC/Flat plate; (c, f) Flat plate/Sigma-2. Multiple daily measurements are shown in each plot.
 664 Different colors represent different size intervals. 1-2 μm : Black, 2-4 μm : Blue, 4-8 μm : Yellow, 8-16
 665 μm : Red, 16-32 μm : Green, 32-64 μm : Cyan.

666 5 Summary and Conclusions

667 Parallel dust aerosol deposition measurements by means of deposition and other passive samplers were
 668 conducted at Izaña Global Atmospheric Watch observatory continuously from 14th of July to 24th of
 669 August 2017. In addition, active aerosol collection was done with a free-wing impactor and an iso-axial
 670 filter sampler. Additional information regarding the aerosol particle size distributions has been obtained
 671 by an OPC. 316,000 single particles from 6 different samplers were analyzed by SEM-EDX, yielding
 672 size resolved deposition rates.

673 As known from previous studies, the total deposition rate was dominated by coarse particles (16-32 μm).
 674 A high temporal daily variability in total dust deposition rate was observed. The size resolved deposition
 675 rate measurements of different passive samplers varied significantly between the samplers under the same
 676 conditions. This was in principle expected from the different sampler construction. Applying suitable
 677 deposition velocity models, atmospheric concentrations were calculated from different sampler
 678 deposition rates. The resulting concentrations on average are in better agreement between the samplers
 679 than the deposition rates. However, discrepancies beyond the measurement uncertainty remain
 680 unexplained by the deposition models, in particular with respect to dependency on wind speed, which is

681 predicted by the models, but not observed. The estimation of an appropriate deposition velocity from
682 different models for calculating atmospheric concentrations remains obviously a challenge. In particular,
683 when considering the size-resolved deposition velocities and deposition rate ratios, great discrepancies
684 show up. While for an integrated bulk measurement or the PM₁₀ size range at least a qualitative agreement
685 between the samplers can be reached, no model – analytical nor CFD – is able to explain the observed
686 sampler-specific variations in deposition rate. Clearly, a better physical understanding is needed here.

687 In the PM₁₀ size range, a good agreement is found between the calculated concentrations for samples
688 from different passive and the concentrations measured using an OPC. For particle sizes above PM₁₀, the
689 comparison of the deposition-derived size distributions with impactor and filter measurements shows
690 considerable underestimates of the deposition samplers, which must be interpreted as overestimation of
691 the large-particle deposition velocities.

692 Deposition velocities from different analytical deposition models are compared to ones calculated using
693 computational fluid dynamics simulations for different samplers. The comparison shows that two
694 methods largely disagree. Moreover, all theory-based deposition velocities (analytical as well as CFD
695 approaches) fail to represent the observed measurement differences between the samplers. This obviously
696 points to the need of better understanding the physics of dry deposition in general.

697 The correlation analysis between dust deposition rate, dust concentrations and wind speed reveals that
698 the variation in deposition rate is mainly controlled by changes in concentration; variations in wind speed
699 play a minor role for wind speeds lower than 6 m/s. However, the situation might be different for higher
700 wind speeds (Kandler et al., 2018).

701 The correlation analysis between deposition rates and OPC measurements demonstrated that BSNE and
702 Sigma-2 can be a good option for PM₁₀ measurement, while the MWAC is not a suitable option. Apart
703 from that result with respect to the PM₁₀ size range, no recommendation for a least biased general purpose
704 deposition sampler can be derived from our study.

705 Moreover, as the results show that the different samplers can't deliver consistent results between the
706 sampler types, a recommendation must be that if a certain sampler type is chosen for a study, it should
707 not be modified or replaced by another one for sake of consistency of results, even if it was shown that
708 the results do not agree well for example with active sampling. The results show, nevertheless, that
709 passive sampling techniques coupled with an automated single particle analysis provides insights into the
710 variation of size distribution, deposition rate and concentration of atmospheric particles.

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720 7 Author contribution

721 A. W. conducted the field measurements and did data evaluation and interpretation. K. S. helped with the
722 field measurements, carried out the SEM analyses and did data processing. J. M. and B. E. executed the
723 CFD model setup and calculations. S. R. operated the OPC including the data processing and the
724 meteorological base measurements. K. K. designed the experiment, designed and prepared the sampling
725 equipment and did data processing and interpretation. All authors contributed to the data discussion and
726 manuscript preparation.

727 8 Data availability

728 The data sets used for this publication are available from the Pangaea repository free of charge
729 (<https://doi.pangaea.de/10.1594/PANGAEA.901413>)

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