

1 **An improved SWAT vegetation growth module for tropical ecosystem**

2 Alemayehu T.^{1,2*}, van Griensven A.^{1,2} and Bauwens W.¹

3 ¹Vrije Universiteit Brussel (VUB), Department of Hydrology and Hydraulic Engineering, Brussel, Belgium

4 ²IHE Delft Institute for Water Education, Department of Water Science and Engineering, Delft, the Netherlands

5 *Correspondence: t.abitew@un-ihe.org; Tel.: +31-621381512

6 **Abstract.** The Soil and Water Assessment Tool (SWAT) is a globally applied river basin eco-hydrological model in
7 a wide spectrum of studies, ranging from land use change and climate change impacts studies to research for the
8 development of best water management practices. However, SWAT has limitations in simulating the seasonal
9 growth cycles for trees and perennial vegetation in the tropics, where rainfall (via soil moisture) is the major plant
10 growth controlling factor than temperature. Our goal is to improve the vegetation growth module of the SWAT
11 model for simulating the vegetation variables such as the leaf area index (LAI) for tropical ecosystem. Therefore, we
12 present a modified SWAT version for the tropics (SWAT-T) that **uses a straightforward** but robust soil moisture
13 index (SMI) - a quotient of the rainfall (P) and reference evapotranspiration (PET) - to initiate a new growth cycle
14 dynamically within a pre-defined period. Our results for the Mara Basin (Kenya/Tanzania) show that the SWAT-T
15 simulated LAI corresponds well with the Moderate Resolution Imaging Spectroradiometer (MODIS) LAI for ever-
16 green forest, savanna grassland and shrubland, indicating that the SMI is reliable for triggering new growth cycle
17 annually. The water balance components (evapotranspiration and streamflow) simulated by the SWAT-T exhibit a
18 good agreement with a thermal-based evapotranspiration (ET-RS) estimate and observed streamflow. The SWAT-T
19 model with the proposed improved vegetation growth module for tropical ecosystem can be a robust tool **for simu-**
20 **lating the vegetation growth dynamics consistently in hydrologic model applications** including land use and climate
21 change impact studies.

22 **1. Introduction**

23 The Soil and Water Assessment Tool (SWAT; Arnold et al., 1998) is a process-oriented, spatially semi-distributed
24 and time-continuous river basin model. SWAT is one of the most widely applied eco-hydrological models for simu-
25 lating hydrological and biophysical processes under a range of climate and management conditions (Arnold et al.,
26 2012; Bressiani et al., 2015; Gassman et al., 2014; van Griensven et al., 2012; Krysanova and White, 2015). Many
27 studies used SWAT in tropical Africa, to investigate the basin hydrology (e.g. Dessu and Melesse, 2012; Easton et
28 al., 2010; Mwangi et al., 2016; Setegn et al., 2009) as well as to study the hydrological impacts of land use change
29 (e.g. Gebremicael et al., 2013; Githui et al., 2009; Mango et al., 2011) and climate change (Mango et al., 2011;
30 Mengistu and Sorteberg, 2012; Setegn et al., 2011; Teklesadik et al., 2017). Notwithstanding the high number of
31 SWAT model applications in tropical catchments, only a few studies underscored the limitation of its plant growth

32 module for simulating the growth cycles of trees, perennials and annuals in this region of the world (Mwangi et al.,
33 2016; Strauch and Volk, 2013; Wagner et al., 2011).

34 It is worthwhile to note that phenological changes in vegetation affect the biophysical and hydrological processes in
35 the basin hydrology and thus play a key role in integrated hydrologic and ecosystem modeling (Jolly and Running,
36 2004; Shen et al., 2013; Strauch and Volk, 2013; Yang and Zhang, 2016; Yu et al., 2016). The Leaf Area Index
37 (LAI), a vegetation attribute commonly used in eco-hydrological modeling, strongly correlates with a vegetation
38 phenological development. Thus, an enhanced representation of the LAI dynamics can improve the predictive capa-
39 bility of hydrologic models, as noted in several studies (Andersen et al., 2002; Yu et al., 2016; Zhang et al., 2009).
40 Arnold *et al.* (2012) also underscored the need for a realistic representation of the local and regional plant growth
41 processes to simulate reliably the water balance, the erosion, and the nutrient yields using SWAT. For instance, the
42 LAI and canopy height are needed to determine the canopy resistance and the aerodynamic resistance to subsequent-
43 ly compute the potential plant transpiration in SWAT. Therefore, inconsistencies in the vegetation growth could
44 result in uncertain ET estimates as noted in Alemayehu *et al.* (2015).

45 SWAT utilizes a simplified version of the Environmental Policy Impact Climate (EPIC) crop growth module to
46 simulate the phenological development of plants, based on accumulated heat units (Arnold et al., 1998; Neitsch et
47 al., 2011). SWAT uses dormancy, which is a function of daylength and latitude, to repeat the annual growth cycle
48 for trees and perennials. Admittedly, this approach is suitable for temperate region. However, Strauch and Volk
49 (2013) showed that the LAI temporal dynamics are not well represented for perennial vegetation (savanna and
50 shrubs) and evergreen forest in Brazil. Likewise, Wagner et al. (2011) reported a mismatch between the growth
51 cycle of deciduous forest in the Western Ghats (India) and the SWAT dormancy period, and they subsequently
52 shifted the dormancy period to the dry season.

53 Unlike temperate regions where the vegetation growth dynamics are mainly controlled by the temperature, the pri-
54 mary controlling factor in tropical regions is the rainfall (i.e. the water availability) (Jolly and Running, 2004;
55 Lotsch, 2003; Pfeifer et al., 2012, 2014; Zhang, 2005). A study of Zhang et al. (2005) explored the relationship be-
56 tween the rainfall seasonality and the vegetation phenology across Africa. They showed that the onset of the vegeta-
57 tion green-up can be predicted using the cumulative rainfall as a criterion to indicate the season change. Jolly and
58 Running (2004) determined the timing of leaf flush in an ecosystem process simulator (BIOME-BGC) after a de-
59 fined dry season in the Kalahari, using events where the daily rainfall (P) exceeded the reference evapotranspiration
60 (PET). They showed that the modeled leaf flush dates compared well with the leaf flush dates estimated from the
61 Normalized Difference Vegetation Index (NDVI), indicating the reliability of a proxy derived from P and PET to
62 pinpoint a season change of tropical ecosystems. Sacks et al. (2010) studied the relationships between crop planting
63 dates and temperature, P and PET globally, using 30-year average climatological values. They noted that in rainfall
64 limited regions the ratio of P to PET is a better proxy for the soil moisture status than is P alone. Using soil mois-
65 ture index (SMI) derived from the ratio of P to PET to trigger new growth cycle annually in hydrological modeling
66 is appealing because as the SMI can be determined *a priori*. On the other hand, Strauch and Volk (2013) used

67 SWAT model simulated soil moisture in the top soil layers with a certain minimum threshold after a defined dry
68 season to indicate the start of a rainy season (SOS) and thus new vegetation growth cycle. Their results showed
69 improvements in the SWAT simulated LAI seasonal dynamics and reproduced well the Moderate Resolution Imag-
70 ing Spectroradiometer (MODIS) 8-day LAI. However, such approach requires calibrating the SWAT parameters for
71 a realistic representation of the soil water balance dynamics often using observed streamflow. Recently, Yu *et al.*
72 (2016) concluded uncertainty in soil moisture is significantly greater than streamflow simulations of a calibrated
73 hydrologic model.

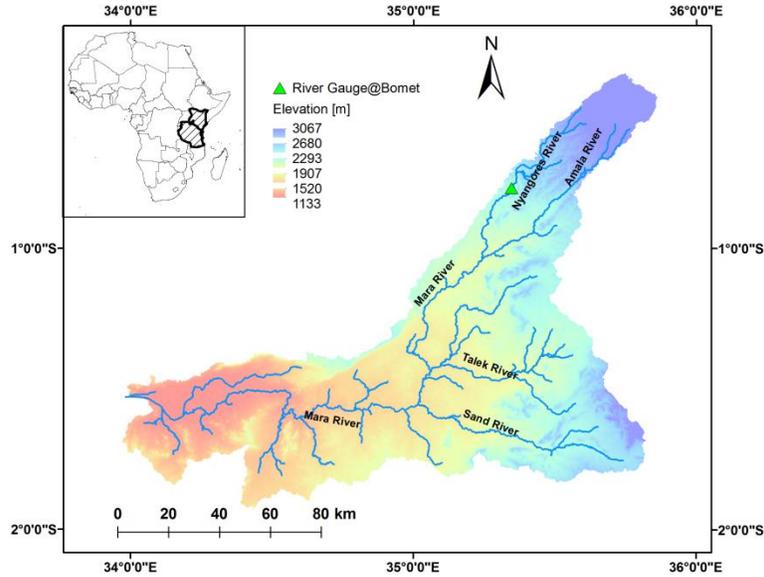
74 The objective this study is to improve the vegetation growth module of SWAT model for trees and perennials in the
75 tropics. Towards this the use of the SMI within a predefined transition months as a dynamic trigger for new vegeta-
76 tion growth cycle will be explored. The modified SWAT (SWAT-T) model will be evaluated using 8-day MODIS
77 LAI and thermal-based ET. Additionally, the model will be evaluated using observed daily streamflow.

78 **2. Materials and methods**

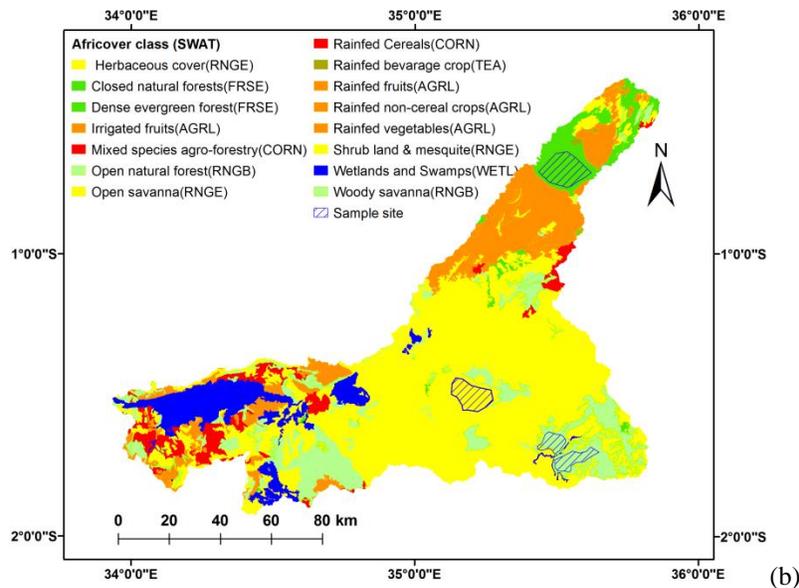
79 **2.1. The study area**

80 The Mara River, a transboundary river shared by Kenya and Tanzania, drains an area of 13,750 km² (Figure 1a).
81 This river originates from the forested Mau Escarpment (about 3000 m.a.s.l.) and meander through diverse agroeco-
82 systems and subsequently crosses the Masai-Mara Game Reserve in Kenya and the Seregenti National Park in Tan-
83 zania and finally feeds the Lake Victoria. The Amala River and the Nyangores River are the only perennial tributar-
84 ies draining the head water region. The Talek River and the Sand River are the two most notable seasonal rivers
85 stemming from Loita Hills.

86 **Rainfall varies spatially mainly due to its equatorial location and its topography.** The rainfall pattern in most part of
87 the basin is bimodal, with a short rainy season (October-December) driven by convergence and southward migration
88 of the Intertropical Convergence Zone (ITCZ) and a long rainy season (March-May) driven by southeasterly trades.
89 In general, rainfall decreases from west to east across the basin while temperature increases southwards. The Mara
90 basin is endowed with significant biodiversity features through a sequence of zones from moist montane forest on
91 the escarpment through dry upland forest to scattered woodland and then the extensive savanna grasslands (Figure
92 1b). Dark volcanic origin soils are common on the escarpment and rangelands while shallow soils that drain freely
93 are found lower down. Poorly drained soils cover the plateau and the plains.



(a)



(b)

94

95

96

97

98 **Figure 1** Location of the Mara Basin (a) and its land cover classes (b). Note the sample sites location for the major natural
 99 **vegetation classes that are used to mask the Moderate Resolution Imaging Spectroradiometer (MODIS) Leaf Area Index**
 100 **(LAI).**

101 2.2. SWAT model description

102 The SWAT (Arnold et al., 1998, 2012; Neitsch et al., 2011) is a comprehensive, process-oriented and physically-
 103 based eco-hydrological model at a river basin scale. SWAT requires specific information about weather, soil proper-
 104 ties, topography, vegetation, and land management practices occurring in the watershed to directly model physical

105 processes associated with water movement, sediment movement, crop growth, nutrient cycling, etc. In SWAT a
 106 basin is partitioned into several sub-basins using topographic information and the sub-basins, in turn, are subdivided
 107 into several Hydrological Response Units (HRUs) with a unique combination of land use, soil and slope class. Each
 108 hydrologic processes are simulated at HRU level on a daily or sub-daily time step and aggregated into sub-basin
 109 level for routing into a river network (Neitsch et al., 2011). SWAT considers five storages: snow, canopy storage,
 110 the soil profile with up to ten layers, a shallow aquifer and a deep aquifer to calculate the water balance (Neitsch et
 111 al., 2011) using the following equation:

$$\Delta S = \sum_{i=1}^t (P - Q_{total} - ET - Losses) \quad (1)$$

112 where ΔS is the change in water storage (mm) and t is time in days. P , Q_{total} , ET and $Losses$ are the daily amounts of
 113 precipitation (mm), the total water yield (mm), the evapotranspiration (mm) and the groundwater losses (mm), re-
 114 spectively. The total water yield represents an aggregated sum of the surface runoff, the lateral flow and the return
 115 flow. In this study, the surface runoff is computed using the Soil Conservation Service (SCS) Curve Number (CN)
 116 method (USDA SCS, 1972).

117 SWAT provides three options for estimating PET: Hargreaves (Hargreaves et al., 1985), Priestley-Taylor (Priestley
 118 and Taylor, 1972), and Penman-Monteith (Monteith, 1965) (Neitsch et al., 2011). The model simulates evaporation
 119 from soil and plants separately as described in Ritchie (1972). The potential soil evaporation is simulated as a func-
 120 tion of PET and leaf area index (LAI) and further reduced with high plant water use while the actual soil water
 121 evaporation is estimated by using exponential functions of soil depth and water content (Neitsch et al., 2011).
 122 SWAT simulated LAI is also required to calculate the potential plant transpiration with formulations that varies
 123 depending on the PET method selection (Alemayehu et al., 2015; Neitsch et al., 2011). The actual plant transpiration
 124 (i.e. the plant water uptake) is reduced exponentially for soil water below field capacity. Therefore, actual evapo-
 125 transpiration in SWAT refers to the sum of evaporation from the canopy, the soil as well as plant transpiration.

126 In this study, we use the Penman-Monteith method (Monteith, 1965) to compute the PET for alfalfa (Neitsch et al.,
 127 2011) as:

$$PET = \frac{\Delta \cdot (H_{net} - G) + \rho_{air} \cdot c_p \cdot [e_z^0 - e_z] / r_a}{\Delta + \gamma \cdot \left(1 + \frac{r_c}{r_a}\right)} \quad (2)$$

128 where PET is the maximum transpiration rate (mm d^{-1}), Δ is the slope of the saturation vapour pressure-temperature
 129 curve ($\text{kPa } ^\circ\text{C}^{-1}$), H_{net} is the net radiation ($\text{MJ m}^{-2} \text{d}^{-1}$), G is the heat flux density to the ground ($\text{MJ m}^{-2} \text{d}^{-1}$), ρ_{air} is
 130 the air density (kg m^{-3}), C_p is the specific heat at constant pressure ($\text{MJ kg}^{-1} \text{ } ^\circ\text{C}^{-1}$), e_z^0 is the saturation vapour pres-
 131 sure of air at height z (kPa), e_z is the water vapor pressure of air at height z (kPa), γ is the psychrometric constant

132 (kPa °C⁻¹), r_c is the plant canopy resistance (s m⁻¹), and r_a is the diffusion resistance of the air layer (aerodynamic
 133 resistance) (s m⁻¹). The plant growth module simulates LAI and canopy height, which are required to determine the
 134 canopy and aerodynamic resistance.

135 2.3. The vegetation growth and Leaf Area Index modeling in SWAT

136 SWAT simulates the annual vegetation growth based on the simplified version of the EPIC plant growth model
 137 (Neitsch et al., 2011). The potential plant phenological development is hereby simulated on the basis of daily accu-
 138 mulated heat units under optimal conditions; however, the actual growth is constrained by temperature, water, nitro-
 139 gen or phosphorous stress (Arnold et al., 2012; Neitsch et al., 2011).

140 Plant growth is primarily based on temperature and hence each plant has its own temperature requirements (i.e.
 141 minimum, maximum and optimum). The fundamental assumption in the heat unit theory is plants have a heat unit
 142 requirements that can be quantified and linked to the time of planting and maturity (Neitsch et al., 2011). The total
 143 number of heat units required for a plant to reach maturity must be provided by the user. The plant growth modeling
 144 includes simulation of the leaf area development, light interception and conversion of intercepted light into biomass
 145 assuming a plant species-specific radiation-use efficiency (Neitsch et al., 2011). The optimal leaf area development
 146 during the initial period of the growth is modeled (Neitsch et al., 2011) as:

$$fr_{LAI_{mx}} = \frac{fr_{PHU}}{fr_{PHU} + \exp(l_1 - l_2 \cdot fr_{PHU})} \quad (3)$$

147 where $fr_{LAI_{mx}}$ is the fraction of the plant's maximum leaf area index corresponding to a given fraction of potential
 148 heat units for the plant, , and l_1 and l_2 are shape coefficients. Once the maximum leaf area index is reached, LAI will
 149 remain constant until the leaf senescence begins to exceed the leaf growth. Afterwards, the leaf senescence becomes
 150 the dominant growth process and hence the LAI follows a linear decline (Neitsch et al., 2011). However, Strauch
 151 and Volk (2013) showed the advantage of using a logistic decline curve, to avoid that the LAI drops to zero before
 152 dormancy occurs. Therefore, we adopted this change to SWAT2012 whereby the LAI during leaf senescence for
 153 trees and perennials is calculated as (Strauch and Volk, 2013):

$$LAI = \frac{LAI_{mx} - LAI_{min}}{1 + \exp(-t)} \quad (4)$$

$$with \ t = 12(r - 0.5) \quad and \quad r = \frac{1 - fr_{PHU}}{1 - fr_{PHU, sen}}, \quad fr_{PHU} \geq fr_{PHU, sen}$$

154 where the term used as exponent is a function of time and t varies from 6 to -6, LAI is the leaf area for a given day
 155 and declines using r as a decline rate, LAI_{mx} and LAI_{min} are the maximum and minimum (i.e. during dormancy) leaf
 156 area index, respectively, $fr_{PHU, sen}$ is the fraction of growing season (PHU) at which senescence becomes the domi-
 157 nant growth process.

158 As detailed in Neitsch *et al.* (2011), the daily LAI calculation for perennials and trees are slightly different, for the
159 latter the years of development is considered.

160 For perennials, the leaf on day i is calculated as:

$$\Delta LAI_i = \left(fr_{LAI_{mx},i} - fr_{LAI_{mx},i-1} \right) LAI_{mx} \cdot \left(1 - \exp(5 \cdot (LAI_{i-1} - LAI_{mx})) \right) \quad (5)$$

161 The total leaf area index, the area of green leaf per unit area of land, is calculated:

$$LAI_i = LAI_{i-1} + \Delta LAI_i \quad (6)$$

162 where ΔLAI_i is the leaf area added on day i , LAI_i and LAI_{i-1} are the leaf area indices for day i and $i-1$ respectively,
163 $fr_{LAI_{mx},i}$ and $fr_{LAI_{mx},i-1}$ are the fraction of the plant's maximum leaf area index for day i and $i-1$, LAI_{mx} is the maximum
164 leaf area index for the plants, yr_{cur} is the age of the tree (years), and $yr_{fulldev}$ is the number of years for tree species to
165 reach full development (years).

166 **2.4. SWAT annual vegetation growth cycle limitation for the tropics**

167 SWAT assumes that trees and perennial vegetation can go dormant as the daylength nears the minimum daylength
168 for the year. Dormancy, which is a function of latitude and daylength during which plants do not grow, is used to
169 repeat the growth cycle each year for trees and perennials. At the beginning of the dormant period, a fraction of the
170 biomass is converted to residue and the leaf area index is set to the minimum value (Neitsch *et al.*, 2011). In the
171 tropics, however, plants growth dormancy is primarily controlled by precipitation (Bobée *et al.*, 2012; Jolly and
172 Running, 2004; Lotsch, 2003; Zhang *et al.*, 2010; Zhang, 2005) and hence the standard SWAT growth module can-
173 not realistically represent the seasonal growth dynamics for trees and perennials. SWAT offers several management
174 settings for the start and the end of growing season based on either heat units (the default) or calendar date schedul-
175 ing. **In fact, the limitation with plant growth dynamics cannot be solved using SWAT management settings as far as**
176 **the latitude and daylength dependent dormancy is activated.**

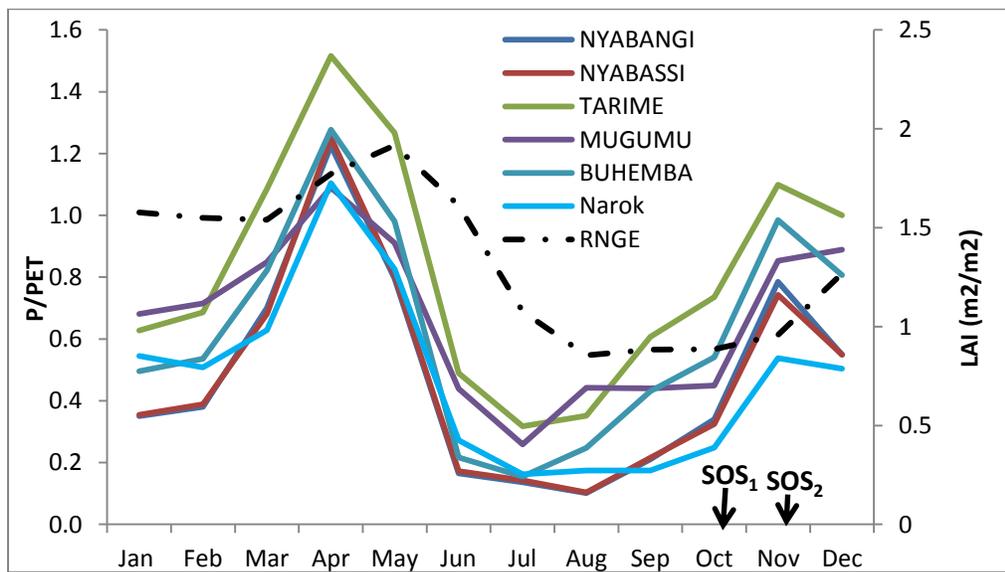
177 **2.5. A soil moisture index-based vegetation growth cycle for the tropics**

178 As several studies demonstrated (Jolly and Running, 2004; Zhang, 2005; Zhang *et al.*, 2006), the water availability
179 in the soil profile is one of the primary governing factors of vegetation growth in tropics. Thus, we propose a soil
180 moisture index (SMI) to trigger new growth cycle for tropical ecosystem in SWAT model within a predefined peri-
181 od. **The SMI is computed as:**

$$SMI = \frac{\sum_{i=1}^N P}{\sum_{i=1}^N PET} \quad (7)$$

182 where P and PET denotes daily rainfall and potential evapotranspiration (mm d⁻¹), N is the number of days of aggre-
 183 gation. In this study we used five days aggregated P and PET (i.e. pentad) to determine the SMI to assure sufficient
 184 soil moisture availability to initiate new growth cycle. The SMI is somewhat similar to the Water Requirement Sat-
 185 isfaction Index (WRSI) (Verdin and Kalver 2002), which is a ratio of ET to PET.

186 Figure 2 presents the SMI seasonal pattern based on long-term climatological P for several gauge stations and PET
 187 from Trabucco and Zomer (2009) across the Mara Basin. It is apparent from Figure 2 that the dry season (mostly
 188 from June - September) shows low SMI values (less than 0.5). Additionally, these patterns resemble well the long-
 189 term monthly average LAI for the savanna ecosystem (the dominant cover in the mid-section of the Mara Basin). In
 190 areas with a humid climate (i.e. the head water regions of the basin), the SMI values are high and the rainfall regime
 191 is different, yet in the relatively drier months (January and February) the SMI is low. As shown in Figure 2, the LAI
 192 and the SMI seasonal dynamics match well with approximately one month lag, indicating the reliability of the SMI
 193 as a proxy for the SOS and hence to trigger the annual vegetation growth cycle. This approach enables SWAT
 194 growth module not only to simulate the vegetation cycle dynamically within a predefined period, also avoids the
 195 need for management setting (“plant” and “kill”).



196

197 **Figure 2** The climatological moisture index (SMI) derived from historical gauge observation across the Mara Basin and
 198 Trabucco and Zomer (2009) global reference evapotranspiration data. Leaf Area Index (LAI) for the savanna ecosystem
 199 (dotted line). SOS₁ and SOS₂ represent the start-of-rainy season (SOS) transition months to trigger growth.

200 To avoid false starts during the dry season, the end of the dry season and the beginning of the rainy season (SOS₁
 201 and SOS₂, respectively) are determined using a long-term monthly climatological P to PET ratio (Figure 2). For
 202 river basins with a single rainfall regime, a single set of SOS months can be used across the basin. However, in ba-
 203 sins with different rainfall regimes, different SOS months need to be set at sub-basin level. In our study area two
 204 distinct rainfall regimes are observed and therefore two different SOS values were needed. For the major part of the
 205 sub-basins October (SOS₁) and November (SOS₂) were used as transitions (Figure 2).

206 **2.6. SWAT-T: the adaptation of the SWAT plant growth module**

207 Based on the rationale elaborated in the preceding sections, we modified the standard SWAT2012 (version 627)
 208 plant growth subroutine for basins located between 20⁰ N and 20⁰ S:

- 209 i) If the simulation day is within SOS₁ and SOS₂ for a given HRU and a new growing cycle is not initiated
 210 yet, the SMI is calculated as the ratio of the pentad P to PET.
- 211 ii) If the SMI exceeds or equals 0.5, a new growing cycle for trees and perennials is initiated. Subsequent-
 212 ly, FR_{PHU} is set to 0 and the LAI is set to the minimum value. Plant residue decomposition and nutrient
 213 release is calculated as if dormancy would occur.
- 214 iii) In case the SMI is still below the threshold (i.e. 0.5) at the end of month SOS₂, a new growing cycle is
 215 initiated immediately after the last date of SOS₂.

216 It is worth noting that SMI threshold could be raised or lowered depending on the climatic condition of the basin.

217 **2.7. Data for model evaluation**

218 *The Leaf Area Index*

219 The remote sensing LAI data used in this study is based on the MODIS TERRA sensor (Table 1). The LAI product
 220 retrieval algorithm is based on the physics of radiative transfer in vegetation canopies (Myneni et al., 2002) and
 221 involves several constants (leaf angle distribution, optical properties of soils and wood, and canopy heterogeneity)
 222 (Bobée et al., 2012). The theoretical basis of the MODIS LAI product algorithm and the validation results are de-
 223 tailed in Myneni et al. (2002). Kraus (2008) validated the MOD15A2 LAI data at Budongo Forest (Uganda) and
 224 Kakamega Forest (Kenya) sites and reported an accuracy level comparable to the accuracy of field measurements,
 225 indicating the reliability of MOD15A2 LAI for evaluating SWAT simulated LAI for the study area.

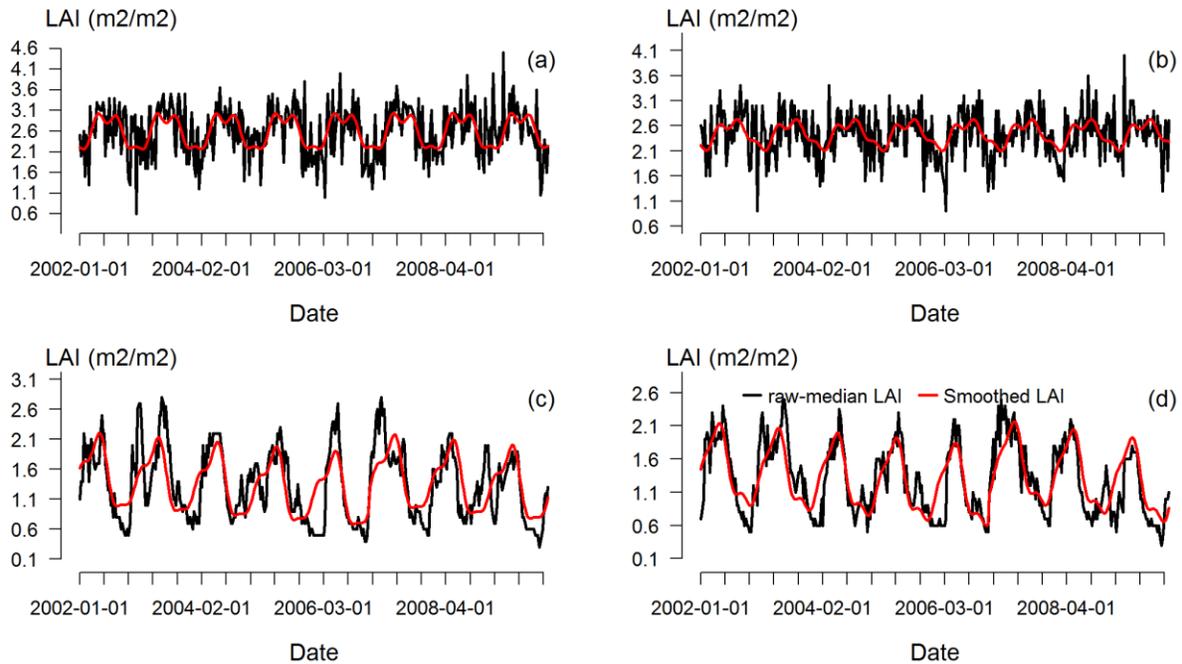
226 Table 1 Summary of the inputs of the SWAT model and the evaluation datasets.

	Spatial/temporal resolution	Source	Description
Rainfall	5 km / 1-day	Roy <i>et al.</i> (2017)	Bias-corrected satellite rainfall for Mara basin
Climate	25 km / 3-hour	Rondell <i>et al.</i> (2004)	Max. and min. temperature, relative
Land cover classes	30 m	FAO (2002)	Land cover classes for East Africa

DEM	30 m	NASA (2014)	Digital elevation model
Soil classes	1 km	FAO (2009)	Global soil classes
Discharge	Daily (2002-2008)	WRMA (Kenya)	River discharge at Bomet
ET	1 km / 8-day	Alemayehu <i>et al.</i> (2017)	ET maps for Mara basin
MOD15A2	1 km / 8-day	LPDAAC(2014)	Global leaf area index

227

228 We selected a representative relatively homogeneous sample sites (i.e. polygons) for evergreen forest (174 km²), tea
229 (123 km²), savanna grassland (136 km²) and shrubland (130 km²) (see Figure 1b) using the Africover classes and
230 Google Earth images. This is useful to reduce the effect of land cover mix while averaging coarse scale (i.e. 1 km)
231 LAI and hence improve the reliability of the LAI timeseries. Subsequently, the MOD15A2 LAI was masked using
232 the polygons of the sample covers and pixels with only quality flag 0, which indicates good quality, were used. Also,
233 pixels with LAI values less than 1.5 during the peak growing months (i.e. period with LAI values mostly above 2.0)
234 were removed. Finally, we extracted the 8-day median LAI time series for each land cover for 2002-2009 and few
235 gaps in the LAI time series were filled using linear interpolation. Notwithstanding with all the quality control efforts,
236 due to the high variability and the inevitable signal noise, we noted breaks and high temporal variation in the LAI
237 timeseries (Figure 3). Verbesselt et al. (2010) developed the Breaks For Additive Seasonal and Trend (BFAST)
238 method that decomposes the Normalized Vegetation Index (NDVI) time series into trend, seasonal, and remainder
239 components. The trend and seasonal components comprise information pertinent to phenological developments as
240 well as gradual and abrupt changes whereas the reminder component comprises noise and error information of the
241 time series. This method has been applied to tropical ecosystems to identify phenological cycles as well as abrupt
242 changes (DeVries et al., 2015; Verbesselt et al., 2010, 2012). In our study, we used the BFAST tool to extract the
243 seasonal development pattern of LAI while excluding the noise and error information from the LAI timeseries. Fig-
244 ure 3 demonstrates the smoothed 8-day LAI time series using BFAST along with the raw-median LAI values. It is
245 apparent from the smoothed LAI time series that the high LAI development occurs during the wet months from
246 March to May, suggesting consistency in the smoothed LAI timeseries. Therefore, the smoothed LAI time series
247 were used to calibrate and evaluate the SWAT-T model vegetation growth module for simulating LAI.



248
 249 **Figure 3** The 8-day raw-median LAI timeseries for evergreen forest (a), tea (b), grassland (c) and shrubland (d) sample
 250 sites. The raw-median LAI is smoothed using the Breaks For Additive Seasonal and Trend (BFAST) method (Verbesselt
 251 et al., 2010).

252 *The evapotranspiration*

253 ET is one of the major components in a basin water balance that is influenced by the seasonal vegetation growth
 254 cycle. Thus, remote sensing-based ET estimates can be used to evaluate (calibrate) the SWAT-T model. Alemayehu
 255 et al. (2017) estimated ET for the Mara River basin using several MODIS thermal imageries and the GLDAS global
 256 weather dataset from 2002 to 2009 at a 8-day temporal resolution based on the Operational Simplified Surface Ener-
 257 gy Balance (SSEBop) algorithm (Senay et al., 2013). The SSEBop mainly depends on the remotely sensed land
 258 surface temperature and the grass reference evapotranspiration (Senay et al., 2013). Alemayehu et al. (2017)
 259 demonstrated that the SSEBop ET explained about 52%, 63% and 81% of the observed variability in the MODIS
 260 NDVI at 16-day, monthly and annual temporal resolution. Also, they suggested that the estimated ET can be used
 261 for hydrological model parameterization. We note the resemblance in the seasonal pattern of the MODIS LAI ana-
 262 lyzed in this study with the SSEBop ET, hereafter referred as remote sensing-based ET (ET-RS). Therefore, we used
 263 this dataset to evaluate the SWAT-T simulated ET at land cover level.

264 *Streamflow*

265 Due to the limited availability of observed streamflow, we used daily observed streamflow series (2002-2008) for
 266 the head water region (700 km²) at Bomet gauging station. The streamflow data is relatively complete with about
 267 11% missing gaps distributed throughout the timeseries.

268 **2.8. Model set up, calibration and evaluation**

269 **2.8.1. The model set up and data used**

270 The Mara River Basin was delineated using a high resolution (30 m) digital elevation model (DEM) (NASA, 2014)
271 in ArcSWAT2012 (revision 627). The basin was subdivided into 89 sub-basins to spatially differentiate areas of the
272 basin dominated by different land use and/or soil with dissimilar impact on hydrology. Each sub-basin was further
273 discretized into several HRUs, which represent unique combinations of soil, land use and slope classes. The model
274 was set up for conditions representing the period 2002-2009. The land cover classes for the basin were obtained
275 from FAO-Africover project (FAO, 2002). Generally speaking, as shown in Figure 1b, the dominant portion of the
276 basin is covered by natural vegetation including savanna grassland, shrubland and evergreen forest. These land cover
277 classes were assigned the characteristics of RNGE, RRGB and FRSE, respectively in SWAT (Neitsch et al.,
278 2011). We extracted the soil classes for the basin from the Harmonized Global Soil Database (FAO, 2008). A soil
279 properties database for the Mara River Basin was established using the soil water characteristics tool (SPAW,
280 <http://hydrolab.arsusda.gov/soilwater>).

281 The list of hydro-climatological and spatial data used to derive the SWAT model are presented in Table 1. In situ
282 measurements of rainfall and other climate variables are sparse and thus bias-corrected multi-satellite rainfall analy-
283 sis data from Roy et al. (2017) were used. The bias-correction involves using historical gauge measurements and a
284 downscaling to a 5 km resolution. Detailed information on the bias-correction and downscaling procedures can be
285 found in Roy *et al.* (2017). Weather data needed to compute the PET using the Penman-Monteith (Monteith, 1965)
286 method was obtained from the Global Land Data Assimilation System (GLDAS) (Rodell et al., 2004). **To remove**
287 **the biases in the PET estimates compared to observation based long-term (1950-2000) seasonal PET (Trabucco and**
288 **Zomer, 2009) estimates, we slightly adjusted the solar radiation for each month at sub-basin level.**

289 **2.8.2. Model calibration and evaluation approach**

290 The main purpose of this study is to explore the potential of the SMI to trigger new vegetation growth cycle for the
291 tropical ecosystem within a predefined period annually. We initially evaluated the effects of the vegetation growth
292 module modification by comparing against the standard SWAT model growth module with varying management
293 settings. This analysis involved uncalibrated simulations of the SWAT models with the default SWAT model pa-
294 rameters, meaning the models differs only with how vegetation growth is simulated. It is worth noting that the aim
295 of these simulations is mainly to expose the inconsistencies in the vegetation growth module structure. Afterwards,
296 we calibrated the parameters related to the simulation of the LAI, the evapotranspiration and the streamflow manual-
297 ly by trial-and-error and expert knowledge for the SWAT-T model. Firstly, SWAT parameters that control the
298 shape, the magnitude and the temporal dynamics of LAI were adjusted to reproduce the MODIS LAI at 8-day for
299 each land cover classes. Then, we adjusted parameters that mainly control streamflow and evapotranspiration (ET)
300 simulation simultaneously using the daily observed streamflow and 8-day ET-RS. Perhaps, the manual adjustment
301 may not be as robust as an automatic calibration as the latter explores a larger parameters space. However, the man-

302 ual calibration is sufficient to illustrate the impact of the modification on the vegetation growth cycle and its effect
303 on the water balance components. The SWAT-T model calibration and validation was done for 2002-2005 and
304 2006-2009, respectively.

305 **2.8.3. The model performance metrics**

306 The Pearson correlation coefficient (r) and the Percent of PBIAS (%bias) were used to evaluate the agreement be-
307 tween the simulated and the remote sensing-based estimates of LAI and evapotranspiration for each land cover class
308 and the streamflow. Additionally, the model performance was evaluated using the Kling-Gupta Efficiency (KGE)
309 (Gupta et al., 2009), which provides a comprehensive assessment by taking into account of the variability, the bias and
310 the correlation in a multi-objective sense.

311 **3. Results and discussion**

312 **3.1. Consistency assessment of the vegetation growth module without calibration**

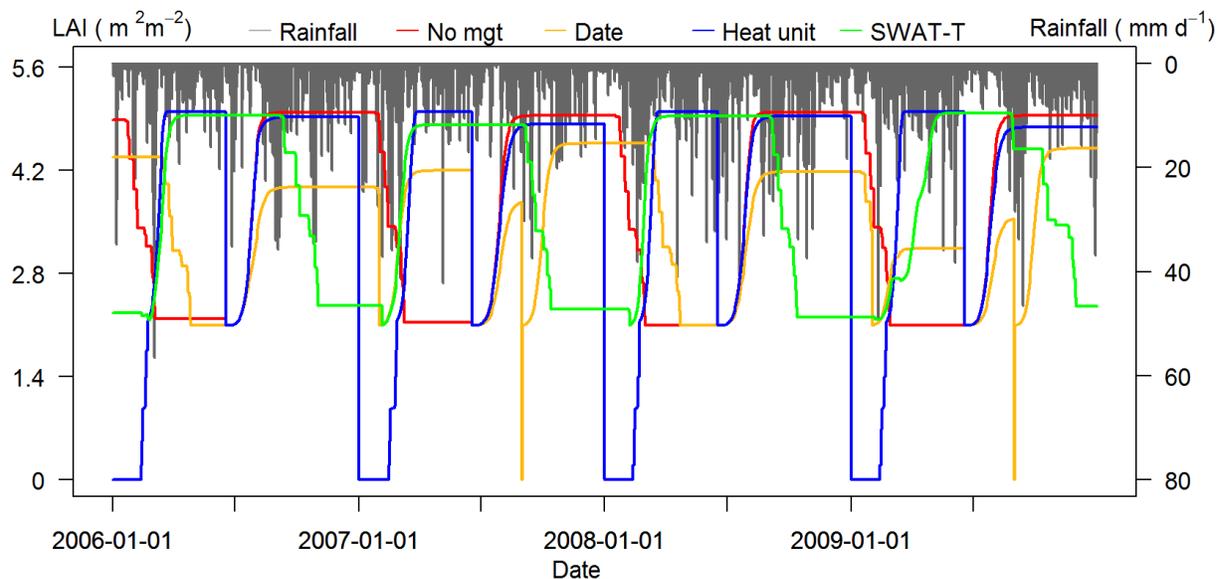
313 **3.1.1. LAI simulation**

314 To highlight the added value of the modified vegetation growth module in SWAT-T for simulating the seasonal
315 growth pattern of trees and perennials, we compared the daily simulated LAI of the standard SWAT2012 (revision
316 627) model and SWAT-T model. At this stage, the models were uncalibrated (i.e. based on default SWAT param-
317 eters). This is useful to explore the effect of the vegetation growth module structural modification on the consistency
318 of simulated LAI annual cycle. We note from the simulation results considerable inconsistencies in the growth cycle
319 of the simulated daily LAI mainly due to the vegetation growth model structure and management settings. For in-
320 stance, Figure 4 and Figure 5 present the simulated daily LAI for FRSE and RNGE based on the standard SWAT
321 model under different management settings and the SWAT-T model. Strauch and Volk (2013), Kilonzo (2014) and
322 Mwang et al. (2016) reported similar observations. The default management setting in the standard SWAT model
323 for starting the new growth cycle (i.e. planting) and ending the growth cycle is scheduled using the FR_{PHU} (Heat
324 unit). Thus, the start and the end of the vegetation growth cycle management settings occurs at FR_{PHU} 0.15 and 1.2,
325 respectively. With this management setting, the simulated LAI is zero at the beginning of each simulation year for
326 all types of vegetation cover. Mwang et al. (2016) improved the SWAT LAI simulation with this management set-
327 ting using FR_{PHU} of 0.001 to start the growing season and minimum LAI of 3.0 for evergreen forest. Yet, this
328 change is region specific and cannot be transferred. As shown in Figure 4 and Figure 5, this can also be partly im-
329 proved using a date scheduling (Date) for the start and the end of the vegetation growth cycle (i.e. instead of heat
330 unit). Additionally, all the management setting can be removed (no mgt) and vegetation is growing since the start of
331 the simulation (i.e. IGRO=1).

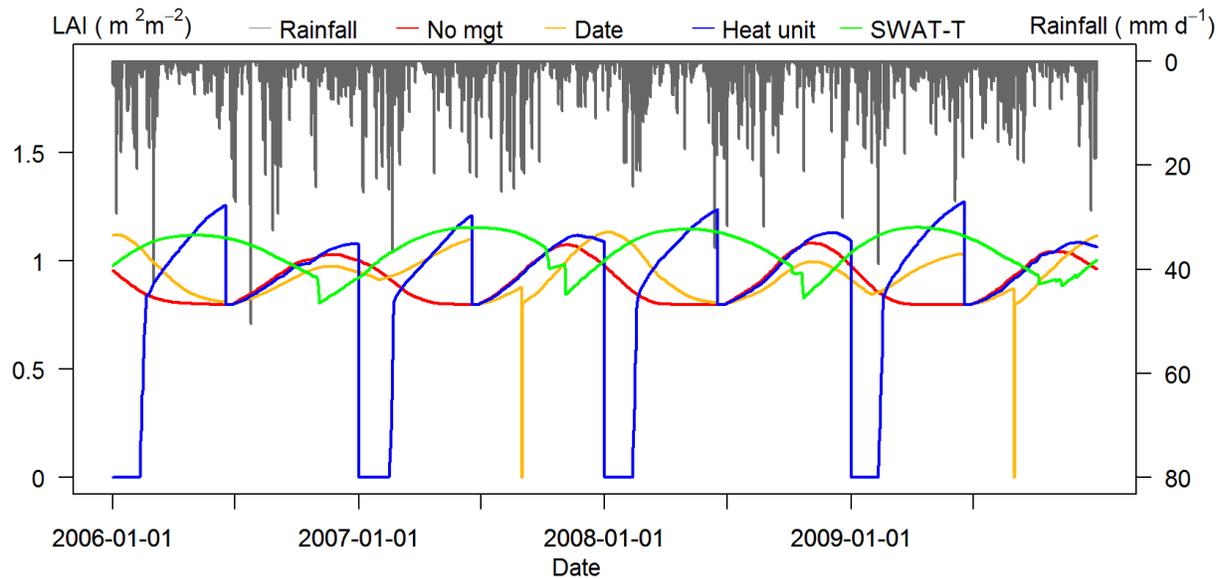
332 The forested head-water region experiences a unimodal rainfall regime, with March-August being the rainy season.
333 In contrast, a bimodal rainfall regime prevails (March-May and October-December) on the remaining part of the
334 basin. Despite the changes in the management settings, it is apparent that the standard SWAT model has inherent

335 limitation to simulate vegetation growth cycle for tropics that are consistent with seasonal rainfall distribution
336 (Figure 4 and Figure 5). Also, the vegetation growth cycle resets annually on 28th June due to dormancy.

337 In contrast, the simulated LAI cycles for FRSE, tea, RNGE and RRGB cover types using the SWAT-T model (i.e.
338 the modified vegetation growth module) reveal a consistent annual cycle and are associated with the seasonal rain-
339 fall pattern (see Figure 4 and Figure 5).



340
341 **Figure 4** The LAI as simulated by the SWAT-T and the standard SWAT models for different management settings for
342 evergreen forest (FRSE) using default SWAT parameter. See management settings explanations in the texts.

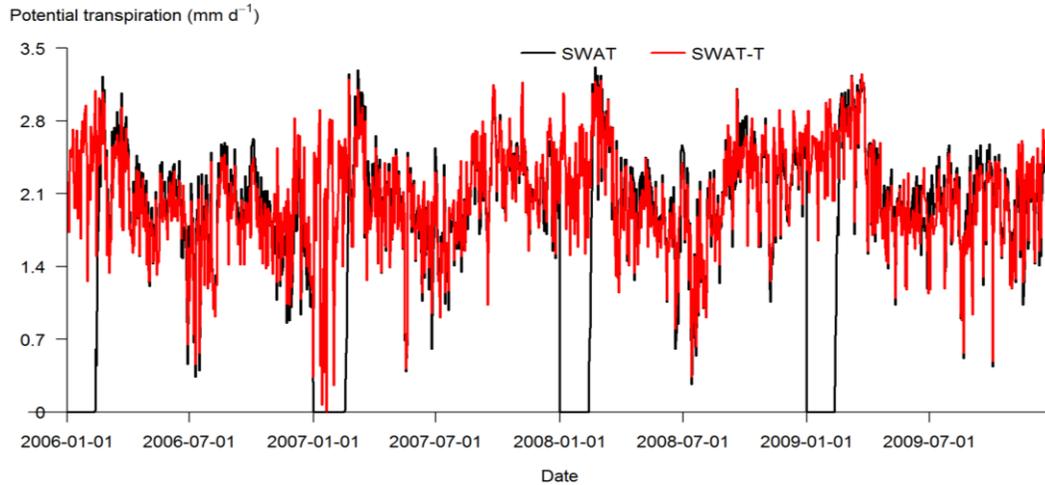


343

344 **Figure 5** The LAI as simulated by the SWAT-T and the standard SWAT models for different management settings for
 345 grassland (RNGE) using default SWAT parameter. See management settings explanations in the texts.

346 **3.1.2. Implication of inconsistent LAI simulation**

347 The LAI is required to compute potential transpiration, potential soil evaporation and plant biomass, among others
 348 in SWAT (Neitsch et al., 2011). For instance, to compute the daily potential plant transpiration in SWAT, the can-
 349 opy resistance and the aerodynamic resistance are determined using the simulated actual daily LAI and canopy
 350 height, respectively (Neitsch et al., 2011). Therefore, the aforementioned limitations of the annual vegetation growth
 351 cycle in the standard SWAT model growth module also influence directly the accuracy of transpiration. For in-
 352 stance, Figure 6 depicts the comparison of the standard SWAT and the SWAT-T simulated daily potential transpira-
 353 tion timeseries for grassland based on the Penman-Monteith approach. We observe 14% (12%) of the standard
 354 SWAT simulated daily potential transpiration timeseries (2002-2009) for FRSE (RNGE) being zero, suggesting a
 355 considerable inconsistency. However, the SWAT-T reduced considerably (i.e. less than 2% for FRSE and RNGE)
 356 the inconsistent zero daily potential transpiration, indicating the improvements in the vegetation growth module.
 357 Several studies have shown the effect of PET method selection in SWAT on simulated ET and other water balance
 358 components (Alemayehu et al., 2015; Maranda and Anctil, 2015; Wang et al., 2006). Alemayehu et al. (2015) re-
 359 ported significant differences in both potential and actual transpiration with the choice of PET method using cali-
 360 brated SWAT model, which partly ascribed to the unrealistic LAI growth cycle. We notice the SWAT-T simulated
 361 potential transpiration is consistent regardless of the PET method selection in SWAT (results not shown here) and
 362 therefore, the improved vegetation growth module in the SWAT-T could reduce the uncertainty arising from the
 363 module structure and thus minimize the uncertainties in model simulation outputs.



364

365 **Figure 6 Inter-comparison of Penman-Monteith-based daily potential transpiration simulated by the SWAT-T and the**
 366 **standard SWAT models for grassland. Note that the heat unit scheduling is used in the standard SWAT model.**

367 **3.2. Evaluation of the calibrated SWAT-T model**

368 **3.2.1. Performance of the LAI simulation**

369 Table 2 presents the list of SWAT model parameters that are adjusted during the calibration process. Initially, the
 370 minimum LAI (ALAI_MIN) for each land cover classes were set based on the long-term MODIS LAI. Also, the
 371 PHU was computed using the long-term climatology, as suggested in Strauch and Volk (2013). The shape coeffi-
 372 cients for the LAI curve ($FRGW_1$, $FRGW_2$, $LAIMX_1$, $LAIMX_2$ and $DLAI$) and the remaining parameters were adjust-
 373 ed during the calibration period by a trail-and-error process such that the SWAT-T simulated 8-day LAI mimics the
 374 MODIS 8-day LAI.

375 Figure 7 presents the comparison of 8-day MODIS LAI with the calibrated SWAT-T simulated LAI aggregated over
 376 several land cover classes for the calibration and validation period. We evaluated the degree of agreement qualita-
 377 tively -by visual comparison- and quantitatively -by statistical measures. From the visual inspection it is apparent
 378 that the intra-annual LAI dynamics (and hence the annual growth cycle of each land cover class) from the SWAT-T
 379 model correspond well with the MODIS LAI data. This observation is supported by correlation as high as 0.94
 380 (FRSE) and 0.92 (RNGB) during the calibration period (Table 3). As shown in Table 3, the model also shows simi-
 381 lar performance during the validation period with low average biases and correlation as high as 0.93 (FRSE). Over-
 382 all, the results indicate that the SMI can indeed be used to dynamically trigger a new growing season within a pre-
 383 defined period.

384 Despite the overall good performance of SWAT-T in simulating LAI, we observed biases for FRSE and Tea mainly
 385 during the rainy season over the calibration and validation period (see Figure 7 top row). This is partly attributed to
 386 the cloud contamination of the MODIS LAI, as shown in Figure 3a and Figure 3b, in the mountainous humid part of
 387 the basin , as note in Krause (2008). Also, the senescence seems to occur slightly early for Tea, as shown in Figure

388 3b, thereby we note a mismatch between SWAT simulated LAI and MODIS LAI. This indicate the need to further
 389 adjust the Fraction of total PHU when leaf area begins to decline (DLAI).

390 Table 2 List of SWAT parameters used to calibrate LAI, ET and streamflow with their default and calibrated values.

Parameter	Parameter definition (unit)	Variable	Default (calibrated)		
			FRSE	RNGE	RNGB
<i>BIO_E</i>	Radiation-use efficiency((kg/ha)/(MJ/m ²))	LAI	15 (17)	34 (10)	34 (10)
<i>BLAI</i>	Maximum potential leaf area index (m ² /m ²)	LAI	5 (4.0)	2.5 (3.5)	2 (3.5)
<i>FRGW₁</i>	Fraction of PHU corresponding to the 1 st point on the optimal leaf area development curve	LAI	0.15 (0.06)	0.05 (0.2)	0.05 (0.2)
<i>LAIMX₁</i>	Fraction of BLAI corresponding to the 1 st point on the optimal leaf area development curve	LAI	0.7 (0.15)	0.1 (0.1)	0.1 (0.1)
<i>FRGW₂</i>	Fraction of PHU corresponding to the 2 nd point on the optimal leaf area development curve	LAI	0.25 (0.15)	0.25 (0.5)	0.25 (0.5)
<i>LAIMX₂</i>	Fraction of BLAI corresponding to the 2 nd point on the optimal leaf area development curve	LAI	0.99 (0.30)	0.7 (0.99)	0.7 (0.99)
<i>DLAI</i>	Fraction of total PHU when leaf area begins to decline	LAI	0.99 (0.30)	0.35 (0.99)	0.35 (0.99)
<i>T_OPT</i>	Optimal temperature for plant growth (°C)	LAI	30 (25)	25 (30)	25 (30)
<i>T_BASE</i>	Minimum temperature for plant growth (°C)	LAI	0 (5)	12 (5)	12 (5)
<i>ALAI_MIN</i>	Minimum leaf area index for plant during dormant period (m ² .m ²)	LAI	0.75 (2.0)	0 (0.75)	0 (0.75)
<i>PHU</i>	Total number of heat units needed to bring plant to maturity	LAI	1800 (3570)	1800 (4100)	1800 (4100)
<i>SOL_Z¹</i>	Soil layer depths (mm)	ET	300 [1000] (480 [1600])	300[1000] (480 [1600])	300[1000] (480 [1600])
<i>SOL_AWC²</i>	Soil available water (mm)	ET/flow	0.26-0.31 [0.27-0.29] (0.18-0.21)	0.26-0.31 [0.27-0.29] (0.18-0.21)	0.26-0.31 [0.27-0.29] (0.18-0.21)

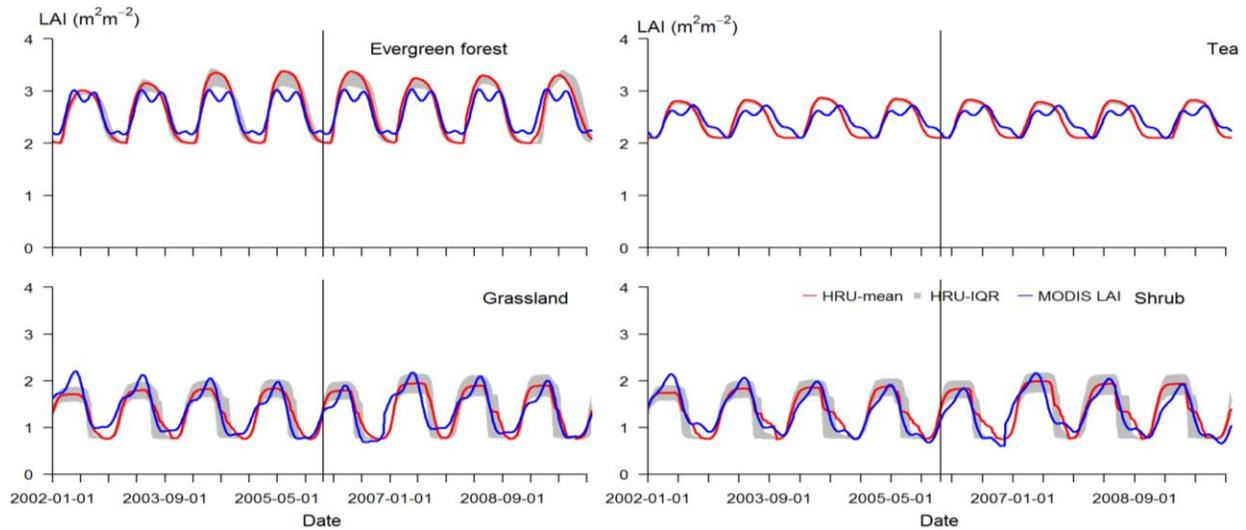
			[0.18-0.20])	[0.18-0.20])	[0.18-0.20])
<i>ESCO</i>	Soil evaporation compensation factor (-)	ET	0.95 (0.88)	0.95 (1)	0.95 (1)
<i>EPCO</i>	Plant uptake compensation factor (-)	ET	1 (1)	1 (1)	1 (1)
<i>GSI</i>	Maximum stomatal conductance at high solar radiation and low vapor pressure deficit (m.s ⁻¹)	ET	0.002 (0.006)	0.005 (0.0035)	0.005 (0.004)
<i>REVAPMN</i>	Depth of water in the aquifer for revap (mm)	ET	750 (100)	750 (100)	750 (100)
<i>CN2³</i>	Initial SCS curve number II value (-)	flow	55 [70] (38 [48])	69 [79] (81 [92])	61 [74] (71 [87])
<i>SURLAG</i>	Surface runoff lag time (day)	flow	4(0.01)	4(0.01)	4(0.01)
<i>ALPHA_BF</i>	Baseflow recession constant (day)	flow	0.048 (0.2)	0.048 (0.2)	0.048 (0.2)
<i>GWQMN</i>	Shallow aquifer minimum level for base flow	flow	1000 (50)	1000 (50)	1000 (50)
<i>GW_REVAP</i>	Groundwater 'revap' coefficient (-)	ET	0.02 (0.1)	0.02 (0.02)	0.02 (0.02)
<i>RCHRG_DP</i>	Deep aquifer percolation fraction (-)	flow	0.05 (0.3)	0.05 (0.1)	0.05 (0.1)

391 ¹SOL_Z values for the top [and lower] soil layers depth

392 ²SOL_AWC values range for the top [and lower] soil layers depending on soil texture and bulk density

393 ³CN2 values for soil hydrologic group B[C]

394



395

396 **Figure 7** The MODIS LAI and the SWAT-T model simulated HRU weighted aggregated 8-day LAI time series (2002-
 397 2009). The gray sheds indicate the boundaries of the 25th and 75th percentiles. The vertical line marks the end of the cali-
 398 bration period and the beginning of the validation period.

399 **Table 3** Summary of the performance metrics for the SWAT-T for simulating LAI, ET and flow. Note that the for LAI
 400 and ET the performance is at 8-day whilst daily for flow.

	LAI calibration (validation)				ET calibration (validation)				Streamflow calibration (vali- dation)
	FRSE	Tea	RNGE	RNGB	FRSE	Tea	RNGE	RNGB	Flow
r	0.94 (0.93)	0.83 (0.83)	0.89 (0.86)	0.92 (0.88)	0.71 (0.68)	0.67 (0.64)	0.72 (0.77)	0.66 (0.72)	0.72 (0.76)
%bias	1.5 (0)	0.1 (0.2)	-3.7 (-0.4)	-1.3 (4.6)	3.7 (6.6)	-1.7 (0.5)	7.8 (11)	1.2 (2.9)	3.5 (15.5)
KGE	0.50 (0.62)	0.42 (0.44)	0.86 (0.85)	0.88 (0.86)	0.71 (0.67)	0.62 (0.62)	0.69 (0.74)	0.66 (0.72)	0.71 (0.71)

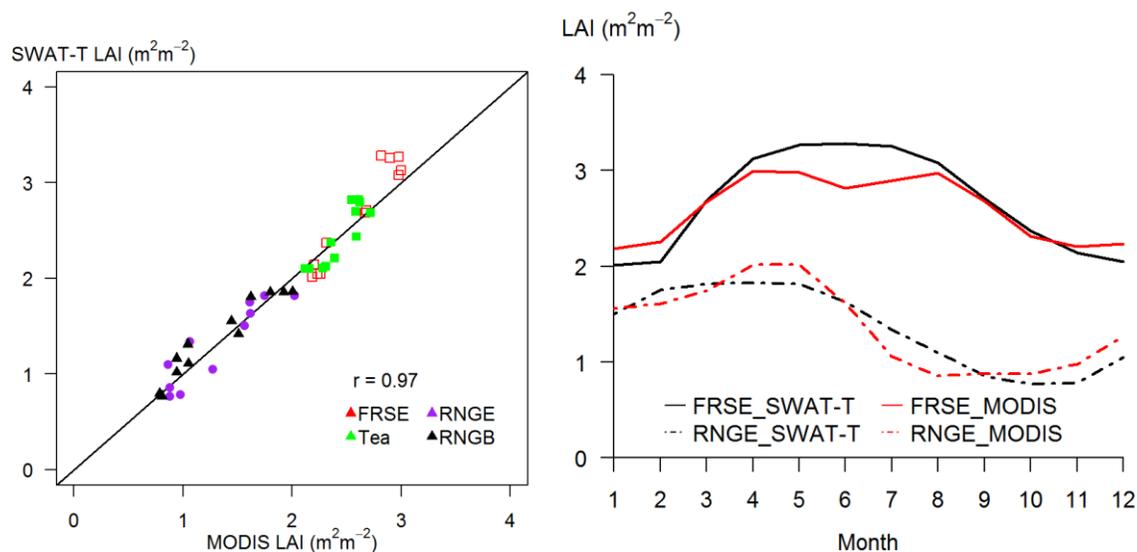
401

402 3.2.2. The seasonal vegetation growth pattern

403 The seasonal pattern of LAI for FRSE, Tea, RNGE and RNGB are analysed using 8-day LAI timeseries (2002-
 404 2009) from calibrated SWAT-T model and MODIS LAI. Generally, not surprisingly, the seasonal dynamics of
 405 SWAT-T simulated LAI and MODIS LAI agrees well (Figure 8 left) with pooled correlation of 0.97.

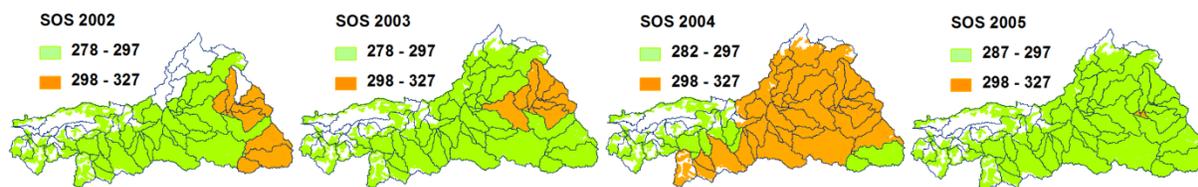
406 As shown in Figure 8 (right), the SWAT-T simulated monthly average LAI shows a higher seasonal variation as
 407 compared to the variation observed from MODIS LAI for FRSE with amplitude (i.e. peak-to-trough difference) is
 408 about 47.7% and 31%, respectively of the average annual MODIS LAI. The seasonal variation from MODIS LAI is
 409 comparable with the results of Myneni *et al.* (2007) who noted 25% seasonal variation in the Amazon forest. We
 410 notice a correlation up to 0.66 between the seasonal LAI and rainfall in the humid part of the basin. Our observa-
 411 tions are in agreement with Kraus (2008), that reported the association of LAI dynamics for forest sites located in
 412 Kenya and Uganda with inter-annual climate variability.

413 In part of the basin where there is a marked dry season, the seasonal LAI dynamics exhibit a notable seasonal varia-
 414 tion, with amplitude (i.e. peak-to-trough difference) that is up to 79% of the mean annual LAI ($1.4 \text{ m}^2/\text{m}^2$) for
 415 RNGE. Unlike the LAI of FRSE and Tea in the humid part, the seasonal rainfall pattern is strongly correlated (up to
 416 0.81) with lagged LAI for RNGE and RNGB. This results is in agreement with several studies that noted that the
 417 LAI dynamics for natural ecosystem in the Sub-Saharan Africa are associated with the rainfall distribution pattern
 418 (Bobée et al., 2012; Kraus et al., 2009; Pfeifer et al., 2014).



419
 420 **Figure 8** The long-term (2002-2009) average seasonal LAI pooled scatter plot (left) and temporal dynamics (right).
 421 **FRSE: evergreen forest; RNGE: grassland; RNGB: shrubland.**

422 In addition to improving the seasonal dynamics of LAI in SWAT without the need of management settings, the SMI
 423 accounts for the year-to-year shifts in the SOS due to climatic variations. This is particularly important for long-term
 424 land use change and climate change impact studies. Figure 9 demonstrates the year-to-year shifts as well as the spa-
 425 tial variation in the SOS dates for part of the Mara River Basin dominated by savanna grassland. Generally, the
 426 season change tends to occur in the month of October (i.e. Julian date 278-304). *Yet, we acknowledge the need of*
 427 *further verification studies in basins with sufficient forcing data and field measurements.*



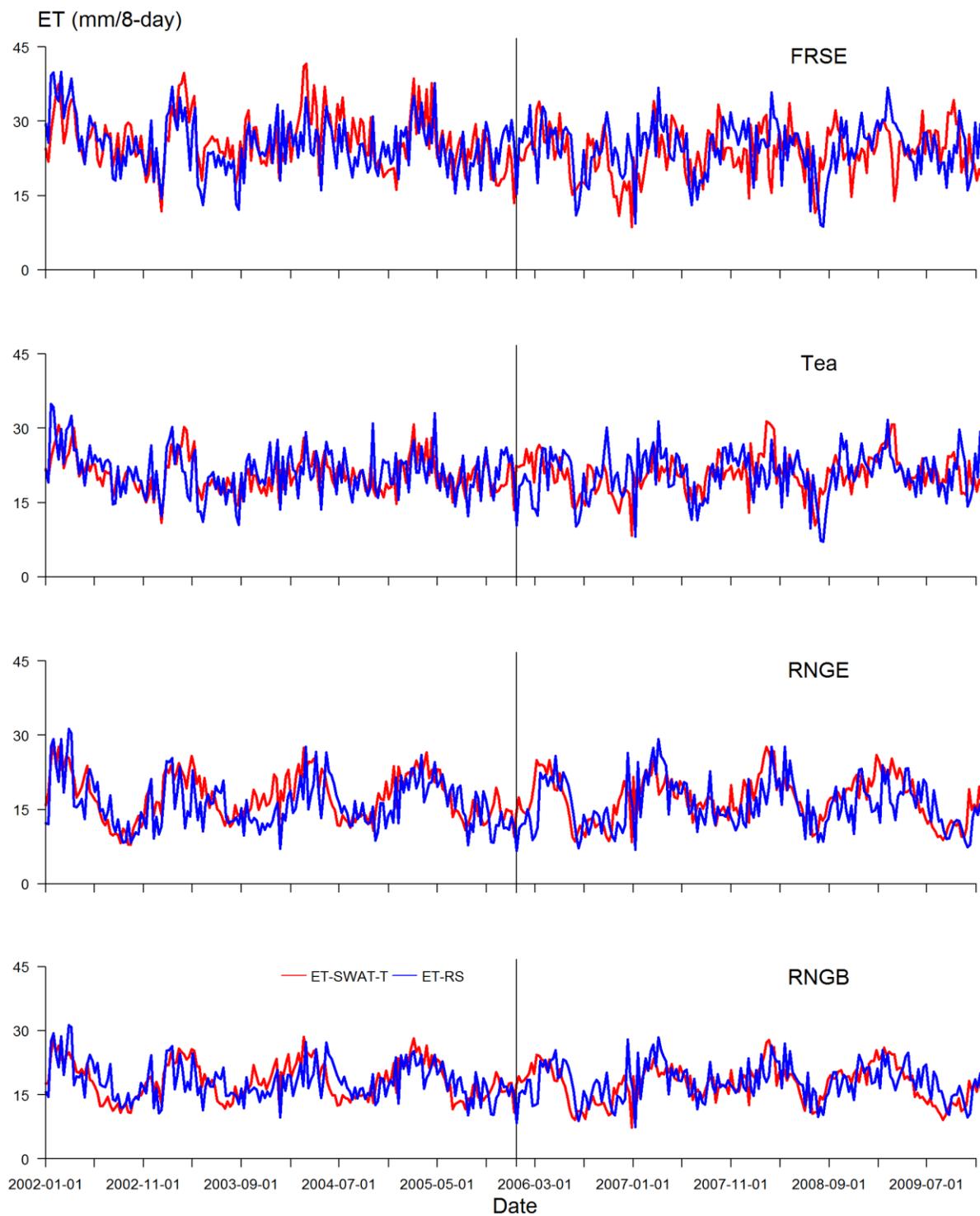
428
 429 **Figure 9** The inter-annual and spatial variation of the start of the rainy season for the savanna vegetation in the Mara
 430 River basin for 2002-2005. Note that Julian dates are used and the mapping is done at HRU scale.

431 3.2.3. The spatial simulation of the evapotranspiration

432 As presented in Table 2, several SWAT parameters were calibrated by comparing SWAT-T model simulated ET
433 with ET-RS. The higher water use by FRSE as compared to other land cover classes is reflected by a lower ESCO,
434 and a higher GW_REVAP and GSI (Table 2). The lower ESCO indicates an increased possibility of extracting soil
435 water to satisfy the atmospheric demand at a relatively lower soil depth. Also, the higher GW_REVAP points to an
436 increased extraction of water by deep-rooted plants from the shallow aquifer or pumping. Similar findings were
437 reported by Strauch and Volk (2013).

438 Figure 10 presents the comparison of 8-day ET-RS and SWAT-T simulated ET for the calibration (2002 - 2005)
439 and validation (2006 - 2009) periods for FRSE, Tea, RNGE and RRGB. Visually, the ET simulated by the SWAT-T
440 fairly agrees with the ET-RS for all the covers. As shown in Table 3, the statistical performance indices show a
441 modest performance in simulating ET for the dominant cover types in the basin. The average model biases for the
442 simulated ET ranges from 7.8% (RNGE) to 1.2% (RRGB) during the calibration period. Additionally, the correla-
443 tion between 8-day ET from the SWAT-T and the ET-RS varies from 0.67 (Tea) to 0.72 (grassland). Overall, we
444 mark similar performance measures during the calibration and validation period, suggesting a fair representation of
445 the processes pertinent to ET.

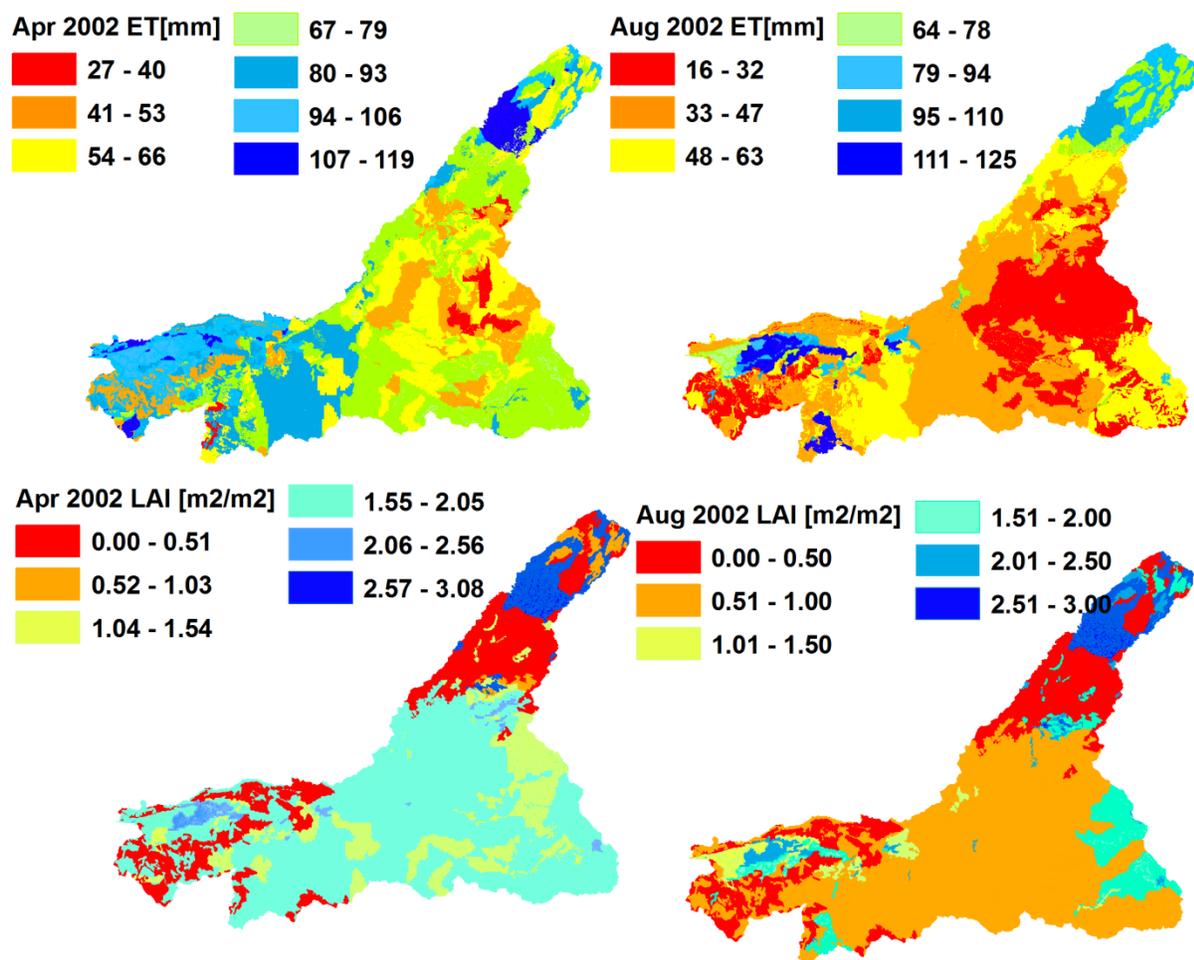
446 The variability of the ET is controlled by several -biotic and abiotic- factors. The 8-day ET time series as simulated
447 by the SWAT-T model illustrates the variation in the temporal dynamics of ET in the study area. For land cover
448 types located in the humid part of the basin (FRSE and tea) there is no clear temporal pattern (Figure 10). In con-
449 trast, the areas covered by RNGE and RRGB show a clear seasonality in the simulated ET. These observations are
450 consistent with the seasonality of the simulated LAI, as discussed section 3.2.2.



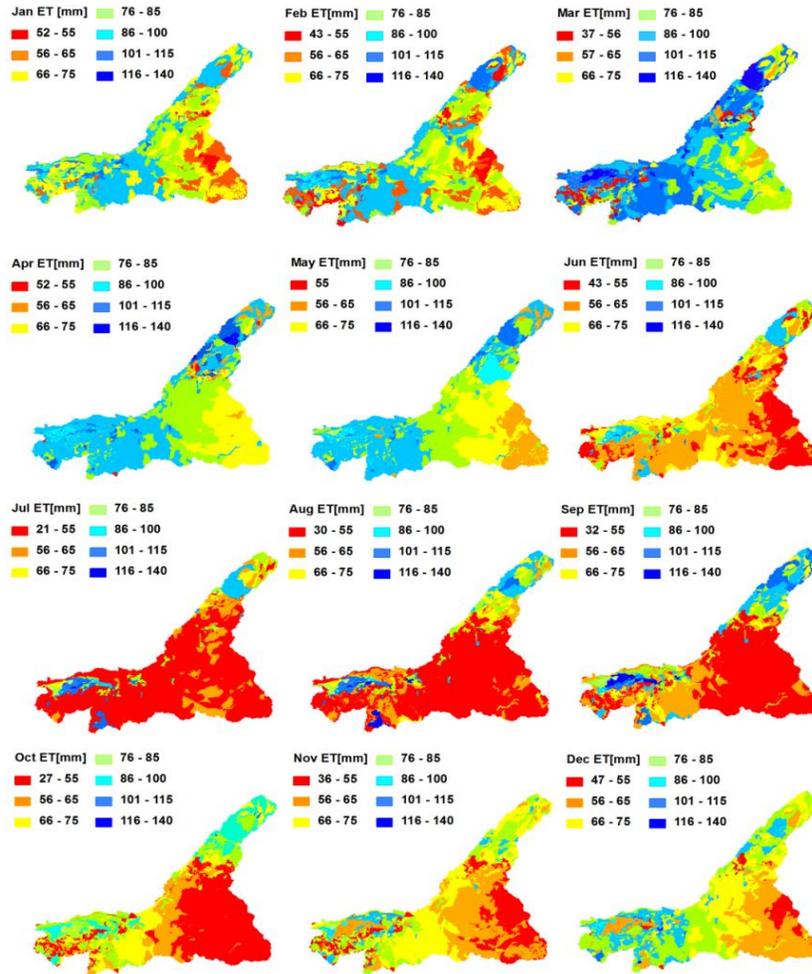
451

452 **Figure 10** The comparison of remote sensing-based evapotranspiration (ET-RS) and SWAT-T simulated ET (ET-SWAT-T) aggregated per land cover classes. Note that for SWAT-T HRU level ET is aggregated per landcover. The vertical
 453 black line marks the end of the calibration period and the beginning of the validation period.
 454

455 To shed light on the consistency of SWAT-T model simulated LAI and ET, we selected simulation outputs for April
 456 and August at HRU level (Figure 11 and Figure 12). Figure 11 (upper row) exhibits the monthly ET at HRU level
 457 for the wet month (April) and dry month (August) in 2002. The lower portion of the basin, with dominant savanna
 458 cover, experiences a monthly ET between 16 and 63 mm/month in August and between 41 and 93 mm/month in
 459 April. These estimates are also well reflected in the spatial distribution of the average monthly simulated LAI
 460 (Figure 11 lower row). We notice that the linear relationship between ET and LAI is stronger, in general, for grass-
 461 land and shrubs than for evergreen forest and tea. The lower correlation for tea and evergreen forest could be partly
 462 attributed to the high evaporation contribution of the wet soil, as the upper portion of the basin receives ample rain-
 463 fall year round. Also, the tea harvesting activities in the upper part of the basin is not taken into account in the mod-
 464 el. We also note visually that during the wet month the spatial variability of ET is higher than that of the LAI (Figure
 465 11). Further comparison research is needed to evaluate the added value of the improved vegetation growth module
 466 on spatial ET simulations. This will be addressed in our ongoing research on ET evaluation.



467
 468 **Figure 11** SWAT-T simulated monthly ET (upper row) and LAI (lower row) for April (wet) and August (dry) 2002 at
 469 HRU level.

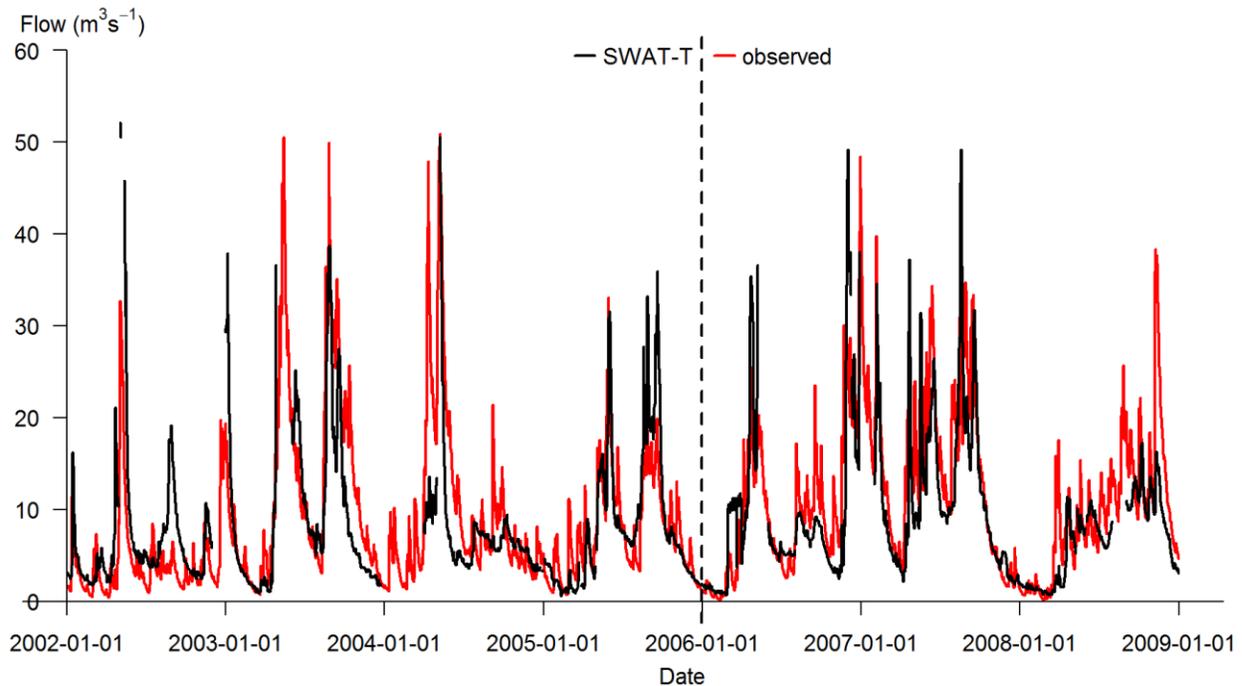


470

471 **Figure 12** The average seasonal and spatial distribution of ET (2002-2009) in the Mara Basin, as simulated by the SWAT-
 472 **T** model at HRU level.

473 **3.2.4. The performance of the streamflow simulations**

474 Figure 13 presents the comparison of daily SWAT-T simulated streamflow with observation for the calibration and
 475 validation periods. Visually, the simulated hydrograph fairly reproduced the observations. The average biases of the
 476 SWAT-T simulated daily streamflow compared to observations are 3.5 and 15.5% during the calibration and valida-
 477 tion periods, respectively (Table 3). **The correlation between the daily observed and simulated streamflows is about**
 478 **0.72 (0.76) during calibration (validation) period.** Additionally, the overall comprehensive assessment using KGE is
 479 about 0.71, suggesting the ability of the calibrated SWAT-T model in reproducing observed streamflow responses.
 480 However, the model tends to underestimate the baseflow and this is more pronounced during the validation period.
 481 This is probably associated with the overestimation of the ET for evergreen forest (6.6%) during the validation,
 482 since ET has a known effect on the groundwater flow.



483

484 **Figure 13 Observed and simulated flows for the Nyangores River at Bomet.**

485 **4. Summary and conclusions**

486 We presented an innovative approach to improve the simulation of the annual growth cycle for trees and perennials -
 487 and hence improve the simulation of evapotranspiration and streamflow- for tropical conditions in SWAT. The ro-
 488 bustness of the changes made to the standard SWAT2012 version 627 have been assessed by comparing the model
 489 outputs with remotely sensed 8-day composite Moderate Resolution Imaging Spectroscopy (MODIS) LAI data, as
 490 well as with thermal-based evapotranspiration (ET-RS) and observed streamflow data. Towards this, we presented a
 491 straightforward but robust soil moisture index (SMI), a quotient of rainfall (P) and reference evapotranspiration
 492 (PET), to trigger a new growing season within a defined period. The new growing season starts when the SMI index
 493 exceeds or equals a certain user defined threshold.

494 The structural improvements in the LAI simulation have been demonstrated by comparing uncalibrated simulation
 495 of LAI using standard SWAT model and SWAT-T model. The results indicated that the modified module structure
 496 for the vegetation growth exhibits temporal progression patterns that are consistent with the seasonal rainfall pattern
 497 in the Mara Basin. Further, we noted better consistency in the SWAT-T model simulated potential transpiration for
 498 perennial and trees, suggesting the usefulness of the vegetation growth module modification in reducing the model
 499 structural uncertainty. Our calibrated SWAT-T model results also show that the calibrated SWAT-T simulated LAI
 500 corresponds well with the MODIS LAI for various land cover classes with correlation of up to 0.94, indicating the
 501 realistic representation of the start of the new growing season using the SMI after a pre-defined period. The im-
 502 provement in the vegetation growth cycle in SWAT is also supported with a good agreement of simulated ET with

503 ET-RS, particularly for the grassland. Additionally, the daily flow simulated with the SWAT-T mimics well the
504 observed flows for the Nyangores River. Therefore, the SWAT-T developed in this study can a robust tool for simu-
505 lating the vegetation growth dynamics consistently in hydrologic model applications including land use and climate
506 change impact studies.

507 **5. Acknowledgments**

508 We would like to thank Tirthankar Roy, the University of Arizona, for providing bias-corrected satellite rainfall
509 products. We also would like to thank the Water Resource Management Authority (WRMA) of Kenya for provision
510 of streamflow data. The technical help on FORTRAN coding from Befekadu Woldegeorgis, Vrije Universiteit Brus-
511 sel, is very much appreciated.

512 **6. Data Availability**

513 The modified SWAT model for Tropics is available upon request from the first author.

514 **7. References**

515 Alemayehu, T., van Griensven, A. and Bauwens, W.: Evaluating CFSR and WATCH Data as Input to SWAT for the
516 Estimation of the Potential Evapotranspiration in a Data-Scarce Eastern-African Catchment, *J. Hydrol. Eng.*, 21(3),
517 5015028, doi:10.1061/(ASCE)HE.1943-5584.0001305, 2015.

518 Alemayehu, T., Griensven, A. van, Senay, G. B. and Bauwens, W.: Evapotranspiration Mapping in a Heterogeneous
519 Landscape Using Remote Sensing and Global Weather Datasets: Application to the Mara Basin, East Africa,
520 *Remote Sens.*, 9(4), 390, doi:10.3390/rs9040390, 2017.

521 Andersen, J., Dybkjaer, G., Jensen, K. H., Refsgaard, J. C. and Rasmussen, K.: Use of remotely sensed precipitation
522 and leaf area index in a distributed hydrological model, *J. Hydrol.*, 264(1–4), 34–50, doi:10.1016/S0022-
523 1694(02)00046-X, 2002.

524 Arnold, J. G., Srinivasan, R., Muttiah, R. S. and Williams, J. R.: Large area hydrologic modeling and assessment
525 part I: model development, *J. Am. Water Resour. Assoc.*, 34(1), 73–89, doi:10.1111/j.1752-1688.1998.tb05961.x,
526 1998.

527 Arnold, J. G., D. N. Moriasi, P. W. Gassman, K. C. Abbaspour, M. J. White, R. Srinivasan, C. Santhi, R. D. Harmel,
528 A. van Griensven, M. W. Van Liew, N. Kannan and M. K. Jha: SWAT: Model Use, Calibration, and Validation,
529 *Trans. ASABE*, 55(4), 1491–1508, doi:10.13031/2013.42256, 2012.

530 Bobée, C., Otlé, C., Maignan, F., De Noblet-Ducoudré, N., Maugis, P., Lézine, A. M. and Ndiaye, M.: Analysis of

531 vegetation seasonality in Sahelian environments using MODIS LAI, in association with land cover and rainfall, J.
532 Arid Environ., 84, 38–50, doi:10.1016/j.jaridenv.2012.03.005, 2012.

533 Bressiani, D. de A., Gassman, P. W., Fernandes, J. G., Garbosa, L. H. P., Srinivasan, R., Bonumá, N. B. and
534 Mendiondo, E. M.: A review of soil and water assessment tool (SWAT) applications in Brazil: Challenges and
535 prospects, Int. J. Agric. Biol. Eng., 8(3), 1–27, doi:10.3965/j.ijabe.20150803.1765, 2015.

536 Dessu, S. B. and Melesse, A. M.: Modelling the rainfall-runoff process of the Mara River basin using the Soil and
537 Water Assessment Tool, Hydrol. Process., 26(26), 4038–4049, doi:10.1002/hyp.9205, 2012.

538 DeVries, B., Verbesselt, J., Kooistra, L. and Herold, M.: Robust monitoring of small-scale forest disturbances in a
539 tropical montane forest using Landsat time series, Remote Sens. Environ., 161, 107–121,
540 doi:10.1016/j.rse.2015.02.012, 2015.

541 Easton, Z. M., Fuka, D. R., White, E. D., Collick, a. S., Biruk Ashagre, B., McCartney, M., Awulachew, S. B.,
542 Ahmed, a. a. and Steenhuis, T. S.: A multi basin SWAT model analysis of runoff and sedimentation in the Blue
543 Nile, Ethiopia, Hydrol. Earth Syst. Sci., 14(10), 1827–1841, doi:10.5194/hess-14-1827-2010, 2010.

544 FAO: Africover Regional Land Cover Database, <http://www.africover.org>, 2002.

545 FAO/IIASA/ISRIC/ISSCAS/JRC: Harmonized World Soil Database (version 1.1). FAO, Rome, Italy and IIASA,
546 Laxenburg, Austria., 2009.

547 FAO, I.-C.: Harmonized World Soil Database (version 1.0), FAO, Rome, Italy and IIASA, Laxenburg, Austr.,
548 2008.

549 Gassman, P. W., Sadeghi, A. M. and Srinivasan, R.: Applications of the SWAT Model Special Section: Overview
550 and Insights, J. Environ. Qual., 43(1), 1, doi:10.2134/jeq2013.11.0466, 2014.

551 Gebremicael, T. G., Mohamed, Y. A., Betrie, G. D., van der Zaag, P. and Teferi, E.: Trend analysis of runoff and
552 sediment fluxes in the Upper Blue Nile basin: A combined analysis of statistical tests, physically-based models and
553 landuse maps, J. Hydrol., 482, 57–68, doi:10.1016/j.jhydrol.2012.12.023, 2013.

554 Githui, F., Mutua, F. and Bauwens, W.: Estimating the impacts of land-cover change on runoff using the soil and
555 water assessment tool (SWAT): case study of Nzoia catchment, Kenya / Estimation des impacts du changement
556 d'occupation du sol sur l'écoulement à l'aide de SWAT: étude du cas du bassi, Hydrol. Sci. J., 54(5), 899–908,
557 doi:10.1623/hysj.54.5.899, 2009.

558 van Griensven, a., Ndomba, P., Yalaw, S. and Kilonzo, F.: Critical review of SWAT applications in the upper Nile
559 basin countries, Hydrol. Earth Syst. Sci., 16(9), 3371–3381, doi:10.5194/hess-16-3371-2012, 2012.

560 Gupta, H. V, Kling, H., Yilmaz, K. K. and Martinez, G. F.: Decomposition of the mean squared error and NSE
561 performance criteria: Implications for improving hydrological modelling, *J. Hydrol.*, 377(1–2), 80–91,
562 doi:10.1016/j.jhydrol.2009.08.003, 2009.

563 Jolly, W. M. . and Running, S. W.: Effects of precipitation and soil water potential on drought deciduous phenology
564 in the Kalahari, *Glob. Chang. Biol.*, 10(3), 303–308, doi:10.1046/j.1529-8817.2003.00701.x, 2004.

565 Kilonzo, F.: Assessing the Impacts of Environmental Changes on the Water Resources of the Upper Mara, Lake
566 Victoria Basin. PhD Thesis, Vrije Universiteit Brussel (VUB)., 2014.

567 Kraus, T.: Ground-based Validation of the MODIS Leaf Area Index Product for East African Rain Forest
568 Ecosystems., 2008.

569 Kraus, T., Schmidt, M., Dech, S. W. and Samimi, C.: The potential of optical high resolution data for the assessment
570 of leaf area index in East African rainforest ecosystems, *Int. J. Remote Sens.*, 30(19), 5039–5059, doi:Doi
571 10.1080/01431160903022878, 2009.

572 Krysanova, V. and White, M.: Advances in water resources assessment with SWAT—an overview, *Hydrol. Sci. J.*,
573 (August), 1–13, doi:10.1080/02626667.2015.1029482, 2015.

574 Lotsch, A.: Coupled vegetation-precipitation variability observed from satellite and climate records, *Geophys. Res.*
575 *Lett.*, 30(14), 1774, doi:10.1029/2003GL017506, 2003.

576 LPDAAC: Land Processes Distributed Active Archive Center (LPDAAC) of NASA, [online] Available from: url:
577 https://lpdaac.usgs.gov/data_access/data_pool (Accessed 5 December 2014), 2014.

578 Mango, L. M., Melesse, a. M., McClain, M. E., Gann, D. and Setegn, S. G.: Land use and climate change impacts
579 on the hydrology of the upper Mara River Basin, Kenya: results of a modeling study to support better resource
580 management, *Hydrol. Earth Syst. Sci.*, 15(7), 2245–2258, doi:10.5194/hess-15-2245-2011, 2011.

581 Maranda, B. and Ancil, F.: SWAT Performance as Influenced by Potential Evapotranspiration Formulations in a
582 Canadian Watershed, *Trans. ASABE*, 58(6), 1585–1600, doi:10.13031/trans.58.11290, 2015.

583 Mengistu, D. T. and Sorteberg, a.: Sensitivity of SWAT simulated streamflow to climatic changes within the
584 Eastern Nile River basin, *Hydrol. Earth Syst. Sci.*, 16(2), 391–407, doi:10.5194/hess-16-391-2012, 2012.

585 Monteith, J. L.: Evaporation and the environment, The state and movement of water in living organisms, in XIXth
586 symposium, Cambridge University Press, Swansea., 1965.

587 Mwangi, H. M., Julich, S., Patil, S. D., McDonald, M. a. and Feger, K.-H.: Modelling the impact of agroforestry on

588 hydrology of Mara River Basin in East Africa, *Hydrol. Process.*, n/a-n/a, doi:10.1002/hyp.10852, 2016.

589 Myneni, R. ., Hoffman, S., Knyazikhin, Y., Privette, J. ., Glassy, J., Tian, Y., Wang, Y., Song, X., Zhang, Y., Smith,
590 G. ., Lotsch, A., Friedl, M., Morisette, J. ., Votava, P., Nemani, R. . and Running, S. .: Global products of vegetation
591 leaf area and fraction absorbed PAR from year one of MODIS data, *Remote Sens. Environ.*, 83(1–2), 214–231,
592 doi:10.1016/S0034-4257(02)00074-3, 2002.

593 Myneni, R. B., Yang, W., Nemani, R. R., Huete, A. R., Dickinson, R. E., Knyazikhin, Y., Didan, K., Fu, R., Negron
594 Juarez, R. I., Saatchi, S. S., Hashimoto, H., Ichii, K., Shabanov, N. V, Tan, B., Ratana, P., Privette, J. L., Morisette,
595 J. T., Vermote, E. F., Roy, D. P., Wolfe, R. E., Friedl, M. a, Running, S. W., Votava, P., El-Saleous, N., Devadiga,
596 S., Su, Y. and Salomonson, V. V: Large seasonal swings in leaf area of Amazon rainforests, *Proc. Natl. Acad. Sci.*,
597 104(12), 4820–4823, doi:10.1073/pnas.0611338104, 2007.

598 NASA: United States Geological Survey Earth Explorer. Available online: <http://earthexplorer.usgs.gov/> (accessed
599 on 9 Sept 2015)., [online] Available from: <http://earthexplorer.usgs.gov/>, 2014.

600 Neitsch, S. L., Arnold, J. G., Kiniry, J. R. and Williams, J. R.: Soil & Water Assessment Tool Theoretical
601 Documentation Version 2009. Texas Water Resources Institute Technical Report No. 406 Texas A&M University
602 System College Station, TX, pp. 647., 2011.

603 Pfeifer, M., Gonsamo, A., Disney, M., Pellikka, P. and Marchant, R.: Leaf area index for biomes of the Eastern Arc
604 Mountains: Landsat and SPOT observations along precipitation and altitude gradients, *Remote Sens. Environ.*,
605 118(2012), 103–115, doi:10.1016/j.rse.2011.11.009, 2012.

606 Pfeifer, M., Lefebvre, V., Gonsamo, A., Pellikka, P. K. E., Marchant, R., Denu, D. and Platts, P. J.: Validating and
607 linking the GIMMS leaf area index (LAI3g) with environmental controls in tropical Africa, *Remote Sens.*, 6(3),
608 1973–1990, doi:10.3390/rs6031973, 2014.

609 Ritchie, J. T.: Model for predicting evaporation from a row crop with incomplete cover, *Water Resour. Res.*, 8(5),
610 1204–1213, doi:10.1029/WR008i005p01204, 1972.

611 Rodell, M., Houser, P. R., Jambor, U., Gottschalck, J., Mitchell, K., Meng, C.-J., Arsenault, K., Cosgrove, B.,
612 Radakovich, J., Bosilovich, M., Entin, J. K., Walker, J. P., Lohmann, D. and Toll, D.: The Global Land Data
613 Assimilation System, *Bull. Am. Meteorol. Soc.*, 85(March), 381–394, doi:10.1175/BAMS-85-3-381, 2004.

614 Roy, T., Serrat-Capdevila, A., Gupta, H. and Valdes, J.: A platform for probabilistic Multimodel and Multiproduct
615 Streamflow Forecasting, *Water Resour. Res.*, (3), 1–24, doi:10.1002/2016WR019752, 2017.

616 Sacks, W. J., Deryng, D., Foley, J. A. and Ramankutty, N.: Crop planting dates: an analysis of global patterns, *Glob.*
617 *Ecol. Biogeogr.*, 19, no-no, doi:10.1111/j.1466-8238.2010.00551.x, 2010.

618 Senay, G. B., Bohms, S., Singh, R. K., Gowda, P. H., Velpuri, N. M., Alemu, H. and Verdin, J. P.: Operational
619 Evapotranspiration Mapping Using Remote Sensing and Weather Datasets: A New Parameterization for the SSEB
620 Approach, *JAWRA J. Am. Water Resour. Assoc.*, 49(3), 577–591, doi:10.1111/jawr.12057, 2013.

621 Setegn, S. G., Srinivasan, R., Melesse, A. M. and Dargahi, B.: SWAT model application and prediction uncertainty
622 analysis in the Lake Tana Basin, Ethiopia, *Hydrol. Process.*, 24(3), 357–367, doi:10.1002/hyp.7457, 2009.

623 Setegn, S. G., Rayner, D., Melesse, A. M., Dargahi, B. and Srinivasan, R.: Impact of climate change on the
624 hydroclimatology of Lake Tana Basin, Ethiopia, *Water Resour. Res.*, 47(4), n/a-n/a, doi:10.1029/2010WR009248,
625 2011.

626 Shen, C., Niu, J. and Phanikumar, M. S.: Evaluating controls on coupled hydrologic and vegetation dynamics in a
627 humid continental climate watershed using a subsurface-land surface processes model, *Water Resour. Res.*, 49(5),
628 2552–2572, doi:10.1002/wrcr.20189, 2013.

629 Strauch, M.: SWAT plant growth modification for improved modeling of tropical vegetation SWAT is increasingly
630 used in the tropics ..., 2013.

631 Strauch, M. and Volk, M.: SWAT plant growth modification for improved modeling of perennial vegetation in the
632 tropics, *Ecol. Modell.*, 269, 98–112, doi:10.1016/j.ecolmodel.2013.08.013, 2013.

633 Teklesadik, A. D., Alemayehu, T., van Griensven, A., Kumar, R., Liersch, S., Eisner, S., Tecklenburg, J., Ewunte, S.
634 and Wang, X.: Inter-model comparison of hydrological impacts of climate change on the Upper Blue Nile basin
635 using ensemble of hydrological models and global climate models, *Clim. Change*, doi:10.1007/s10584-017-1913-4,
636 2017.

637 Trabucco, A. and Zomer, R. J.: Global Aridity Index (Global-Aridity) and Global Potential Evapo-Transpiration
638 (Global-PET) Geospatial Database. CGIAR Consortium for Spatial Information. Published online, available from
639 the CGIAR-CSI GeoPortal, 2009.

640 USDA SCS: Section 4 Hydrology, *National Engineering Handbook*. Washington., 1972.

641 Verbesselt, J., Hyndman, R., Newnham, G. and Culvenor, D.: Detecting trend and seasonal changes in satellite
642 image time series, *Remote Sens. Environ.*, 114(1), 106–115, doi:10.1016/j.rse.2009.08.014, 2010.

643 Verbesselt, J., Zeileis, A. and Herold, M.: Near real-time disturbance detection using satellite image time series,
644 *Remote Sens. Environ.*, 123(2012), 98–108, doi:10.1016/j.rse.2012.02.022, 2012.

645 Wagner, P. D., Kumar, S., Fiener, P. and Schneider, K.: Hydrological Modeling with SWAT in a Monsoon -Driven
646 environment: Experience from the Western Ghats, India, *Trans. ASABE*, 54(5), 1783–1790, 2011.

647 Wang, X., Melesse, A. M. and Yang, W.: Influences of Potential Evapotranspiration Estimation Methods on
648 SWAT's Hydrologic Simulation in a Northwestern Minnesota Watershed, *Trans. ASABE*, 49(6), 1755–1771,
649 doi:10.13031/2013.22297, 2006.

650 Yang, Q. and Zhang, X.: Improving SWAT for simulating water and carbon fluxes of forest ecosystems, *Sci. Total*
651 *Environ.*, 569–570, 1478–1488, doi:10.1016/j.scitotenv.2016.06.238, 2016.

652 Yu, X., Lamačová, A., Duffy, C., Krám, P. and Hruška, J.: Hydrological model uncertainty due to spatial
653 evapotranspiration estimation methods, *Comput. Geosci.*, 90(2016), 90–101, doi:10.1016/j.cageo.2015.05.006,
654 2016.

655 Zhang, K., Kimball, J. S., Nemani, R. R. and Running, S. W.: A continuous satellite-derived global record of land
656 surface evapotranspiration from 1983 to 2006, *Water Resour. Res.*, 46(9), 1–21, doi:10.1029/2009WR008800, 2010.

657 Zhang, X.: Monitoring the response of vegetation phenology to precipitation in Africa by coupling MODIS and
658 TRMM instruments, *J. Geophys. Res.*, 110(D12), D12103, doi:10.1029/2004JD005263, 2005.

659 Zhang, X., Friedl, M. A. and Schaaf, C. B.: Global vegetation phenology from Moderate Resolution Imaging
660 Spectroradiometer (MODIS): Evaluation of global patterns and comparison with in situ measurements, *J. Geophys.*
661 *Res. Biogeosciences*, 111(4), 1–14, doi:10.1029/2006JG000217, 2006.

662 Zhang, Y., Chiew, F. H. S., Zhang, L. and Li, H.: Use of Remotely Sensed Actual Evapotranspiration to Improve
663 Rainfall–Runoff Modeling in Southeast Australia, *J. Hydrometeorol.*, 10(4), 969–980, doi:10.1175/2009JHM1061.1,
664 2009.

665