

Study the impact of three Asian industrial regions on PM_{2.5} in Taiwan and the process analysis during transport

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10 **Abstract.** The outflow of East Asian haze (EAH) has gathered much attention in recent years. For downstream areas, it is meaningful to understand the impact of crucial upstream sources and the process analysis during transport. This study evaluated the impact of PM_{2.5} from the three biggest industrial regions in Asian continent: Bohai Rim industrial region (BRIR), Yangtze River Delta industrial region (YRDIR), and Pearl River Delta industrial region (PRDIR) on Taiwan and discussed the processes during transport with the help of air quality modeling. The simulation results revealed the contributions of monthly average
15 PM_{2.5} from BRIR and YRDIR were 0.7~1.1 μg m⁻³ and 1.2~1.9 μg m⁻³ (~5 % and 7.5% of total concentration) on Taiwan, respectively in January 2017. When the Asian anticyclone moved from Asian continent to the West Pacific, e.g. on Jan 9th 2017, the contributions from BRIR and YRDIR to northern Taiwan could reach 6~8 and 9~12 μg m⁻³. The transport of EAH from BRIR and YRDIR to low latitude regions was horizontal advection (HADV), vertical advection (ZADV), and vertical diffusion (VDIF) over Bohai Sea and East China Sea. Over Taiwan Strait and northern South China Sea, cloud processes
20 (CLDS) was the major contribution to PM_{2.5} due to high relative humidity environment. Along the transport from high latitude regions to low latitude regions, Aerosol chemistry (AERO) and Dry deposition (DDEP) were the major removal processes. When the EAH intruded northern Taiwan, the major processes to the gains of PM_{2.5} at northern Taiwan were HADV and AERO. The stronger the EAH was the easier the EAH could influence central and southern Taiwan. Although PRDIR was located at the downstream of Taiwan under northeast wind, the PM_{2.5} from PRDIR could transport upward above boundary layer and
25 moved eastwards. When the PM_{2.5} plume moved overhead Taiwan blocked by mountains, PM_{2.5} could transport downward via boundary layer mixing (VDIF) and further enhanced by the passing cold surge. In contrast, for the simulation of July 2017, the influence from three industrial regions was almost negligible unless there was special weather system like thermal lows, which may carried pollutants from PRDIR to Taiwan, but the occurrence was rare.

1. Introduction

30 The damage of PM_{2.5} (aerodynamic diameter is equal or less than 2.5 μm) on respiratory system has been proved (Kagawa, 1985; Schwartz et al., 1996 ; Zhu et al., 2011). The short-term human exposure to PM_{2.5} could inflict cardiovascular and respiratory diseases, reducing lung functions, and increasing respiratory symptoms such as rapid breath, cough, and asthma. While the long-term influences include the mortality from heart or lung disease, cardiovascular illness (Pope et al., 2004 ; Brook et al., 2004 ; Ohura et al., 2005), and overuse of medical resources (Atkinson et al., 2001). Environmentally, the PM_{2.5}
35 not only absorbs and scatters solar radiation but also impairs visibility (Na et al., 2004), influences the balance of radiation and global climate (Hu et al., 2017), and the heterogeneous reactions of oxidants in the troposphere (Tie et al., 2005).

The East Asian haze (EAH) has been disturbing in spring and winter around the East Asia due to the spread of anticyclones over Asian continent. (Fu et al., 2014; Yang et al., 2016). When the Asian anticyclones was formed at the Siberia moved southeastwards, the peripheral circulation usually transported EAH to downwind regions including Korea, Japan, and Taiwan (Zhang et al., 2015). Most literatures discussing the transport of EAH in recent years generally applied two methods: the trajectories statistics (TS) and the chemical transport modeling (CTM). The TS method calculated the frequency of the backward trajectories passing through specific surrounding regions. The frequency of the trajectories passing through a specific region implied the impact level of this region. The trajectories could be calculated from, for example the archived meteorological data of NOAA ARL (www.ready.noaa.gov/archives.php) or the model outputs of MM5 (Mesoscale Model version 5, Dudhia, 1993) or WRF (Weather Research and Forecasting, Skamarock and Klemp, 2008). Pawar et al. (2015) utilized the TS method to assess the impacts of short-range and long-range transport (LRT) PM_{2.5} on Mohali in north-west Indo-Gangetic plain. Similar method was applied to evaluate the contribution of LRT of PM_{2.5} to south-western Germany (Garg and Sinha, 2017) and eastern Germany (van Pinxteren et al., 2019). Yang et al. (2018) also used this method to evaluate the influence of PM_{2.5} from the Bohai Sea, Yangtze River Delta, and Pearl River Delta regions on Beijing. Although the TS method has been used widely, the passing frequency over some specific regions can only approximate statistics of the contributions from those regions. The plume transport from an upstream place to the receptor would exchange and react with air and pollutants along the path of transport. It suggests the plume arriving the receptor is no longer the plume emitted from the initial upstream place. The farther the upstream place is away from the receptor; the more uncertainty will be in the TS method. Therefore, the TS method would contain substantial uncertainty.

The application of CTM on the study of transport usually comprises two methods: the Brute Force Method (BFM) and the Apportionment Method (AM). The principle of BFM is to run two simulations: one control run and another one without specific source. The difference of the base case and the zero-out case is the reduction of the zero-out source. The reduction is approximate the contribution of that zero-out source only under the assumption when the contributions of each sources are additive. However, there is indirect contribution not considered in BFM method, i.e., the chemical reactions between the specific zero-out source and surrounding sources is neglected. The indirect contribution could be large if the zero-out sources and surrounding sources are both huge and have enough time to react. BFM method has been widely used for estimating the contribution of a specific source or the effect of a control strategy (Marmur et al., 2005; Burr and Zhang, 2011; Chen et al., 2014; Li et al., 2017) because this method is easy and straightforward. Nevertheless, this method is not perfect for potentially ignoring chemical reactions between the specific source with the remaining sources. Therefore, the BFM method is more reliable if the effect of the chemical reaction is minor. The AM method is more complex and applied the idea of apportionment technique into CTM model. The simulation consumes much computing resources, but it could estimate the contributions of different emission sources in a run. Skyllakou et al. (2014) applied the particulate matter source apportionment technique (PSAT, Wagstrom et al., 2008) in PMCAMx model (Fountoukis et al., 2011) to assess the impact of local pollution (LP), short distance transport (50-500 km), and LRT (>500 km) on Paris in France. Kwok et al. (2013) also developed a similar technique called Integrated Source Apportionment Method (ISAM) in CMAQ model (Byun and Schere, 2006). The AM method can be used to evaluate the contributions of different emission sources simultaneously; however, it does not comprehensively account for the non-linear chemical reactions between sources. BFM and AM methods both have their edge over the other. The CTM modeling requires large computer resources and contains many uncertainties like emissions, meteorology, chemical mechanisms, and numerical methods. However, the CTM especially the AM method is able to give clearer contributions from a specific source compared to the TS method.

The LRT of EAH has tremendous impact on the air quality in Taiwan. The following is a brief of such modelling studies. Chang et al. (2000) applied the CTM to simulate the influence of LRT acid pollutants from East Asian to Taiwan. In the six events of 1993, the average contribution accounted for 9 - 45% and 6 - 33% of total sulfur and nitrogen deposition on Taiwan, highest when the northeast monsoon prevailed. Lin et al. (2004) examined the meteorological and air quality data from

80 November 1999 to May 2000, and from November 2000 to May 2001 in Taiwan. They classified the LRT in winter into dust transport, frontal transport with pollutants, and LRT of background air masses which contributed an average PM_{10} level of $127.6 \mu\text{g m}^{-3}$, $85.0 \mu\text{g m}^{-3}$, and $32.8 \mu\text{g m}^{-3}$ respectively. Furthermore, the frequencies of LRT events and LP events were 25.2% and 71.7% (missing data accounts 3.1%). Chuang et al. (2008a) classified the pollution weather patterns for Taipei $PM_{2.5}$ events. They coined the weather system during LRT events as the “high-pressure pushing” in which the high-pressure systems

85 advected the pollutants from Asian continent to Taiwan. Subsequently, Chuang et al. (2008b) utilized CMAQ to simulate the chemical evolution of $PM_{2.5}$ compositions in the moving plume. They found that the proportion of nitrate and sulfate would decrease and increase respectively along the path. Chen et al. (2013, 2014) also applied the CMAQ to assess the $PM_{2.5}$ distribution in East Asia and subsequently estimated the impact of $PM_{2.5}$ from Asian continent on Taiwan. They suggested the direct and indirect LRT accounted for 27% and 10% of $PM_{2.5}$ in Taiwan in 2007. For the autumn and winter of 2007, the LRT

90 contributed 39% and 41% of total $PM_{2.5}$ in Taiwan. Wang et al. (2016) combined backward trajectories and AOD distribution to estimate the impact of EAH on Taiwan. Their results suggested the $PM_{2.5}$ level was $57.1 \pm 13.6 \mu\text{g m}^{-3}$ for haze event, which is four folds of the background events ($13.7 \pm 7.4 \mu\text{g m}^{-3}$) from 2005 to 2013. They also estimated pollution transport time from the Yangtze River Delta (YRD) to the northern tip of Taiwan was about 28 hours. Chuang et al. (2017) discussed three types of $PM_{2.5}$ episodes into the LRT, the LP and the LRT/LP mix. Both the simulation and observation showed the proportion of

95 NO_3^- in $PM_{2.5}$ was very small in the EAH and strong north to northeast wind increased the proportion of sea salt. Chuang et al. (2018) developed an efficient method to estimate the LRT- $PM_{2.5}$ and LP- $PM_{2.5}$ at any place in Taiwan. They classified the daily $PM_{2.5}$ into LRT-Event (high concentration events caused nearly by pure LRT), LRT-Ordinary (non-events caused nearly by pure LRT), and LRT/LP&Pure LP (other days influenced by mix of LRT and LP & pure LP), which were $31\text{--}39 \mu\text{g m}^{-3}$, $12\text{--}16 \mu\text{g m}^{-3}$, $4\text{--}13 \mu\text{g m}^{-3}$ at the northern tip of Taiwan from 2006 to 2015 for northeast monsoon period. On average, the ratio of

100 LRT- $PM_{2.5}$ and LP- $PM_{2.5}$ for LRT-Event was 70:30 for northern Taiwan, 50:50 for central Taiwan, and 30:70 for southern Taiwan; for LRT-Ordinary was 60:40 for northern Taiwan and 40:60 for central and southern Taiwan; for LRT/LP&Pure LP was 30:70 for northern Taiwan and 25:75 for central and southern Taiwan. Their results also showed the annual LRT- $PM_{2.5}$ decreased since 2013, which implied the emissions in Asian continent decreased since then.

The above studies all showed the East Asian continent was the dominant source of LRT $PM_{2.5}$ for Taiwan in winter period.

105 Therefore, if we can realize the sources contribute the most to LRT $PM_{2.5}$ and the transport pathway, then we can enhance the ability to predict the LRT $PM_{2.5}$, i.e. the EAH. From the emission map of Asia (Li et al., 2017; Kurokawa and Ohara, 2020), the largest emission source was power and industry sector. The three biggest industrial regions in mainland China are the Bohai Rim industrial region (BRIR), the Yangtze River Delta industrial region (YRDIR), and the Pearl River Delta industrial region (PRDIR), as illustrated in Fig. 1. The present study attempts to assess the impact of these three industrial regions on the $PM_{2.5}$

110 in Taiwan. It applied the CTM with BFM method to simulate four scenarios: the *Base* (control case with integrated emissions), *Brir* (all emissions except BRIR), *Yrdir* (all emissions except YRDIR), and *Prdir* (all emissions except PRDIR) scenarios and thus resulted in the contributions of each industrial region. As mentioned above, the difference of Base and sensitivity scenarios is the reduction of the specific source. Only when the chemical reactions are not important then the reduction can be approximate the contribution of that specific source. In this study, the pollutants from those three industrial regions transport

115 directly to Taiwan instead of meandering movement. Therefore, we can roughly estimate the contribution of BRIR, YRDIR, and PRDIR to $PM_{2.5}$ as the difference between the *Base* case and the *Brir*, *Yrdir*, and *Prdir* cases. In addition, this study applied the Integrated Process Rate (IPR) technique (Byun and Schere, 2006; Liu and Zhang, 2013; Zhu et al., 2015) in CMAQ to discuss the process analysis during transport from the industrial regions to Taiwan. The bottom 20 layers (below 1.7 km) were selected for IPR analysis since they have covered the boundary layer where the physical and chemical processes take place.

120 The climate in East Asia basically is divided into the northeast monsoon season in winter and southwest monsoon season in

summer. In order to understand the LRT in different seasons, the simulation periods for this study were January and July 2017. We also selected representative events to discuss in detail.

2. Methods

125 It is known that the EAH events mainly occur in winter (Chuang et al., 2008a; Wang et al., 2016). Although the high PM_{2.5} events in Taiwan caused by the EAH during spring period sometimes was enhanced by the Southeast Asian biomass burning aerosol (Yen et al., 2013; Chuang et al., 2016; Lin et al., 2017), the latter would implicitly complicate the transport of EAH and their co-occurrence has left to be a study in the near future. Therefore, this study chose January and July 2017 to represent the LRT in winter and summer period and the contrast between them. In addition, year 2017 was selected for this study is that it can reflect the impact of EAH lately because the anthropogenic emission in China has been decreasing obviously in recent
130 years (Zheng et al., 2018; Chuang et al., 2018).

2.1 Geographical location of meteorological and air quality observation sites

Taiwan is an island located in the West Pacific and separated from mainland China on the west by the Taiwan Strait. The north is the East China Sea and the south sits the Philippines across the Bashi Strait. For meteorology evaluation, we chose eight representative stations: Peng Jiayu (PJY, #1 in Fig. 1), Taipei (TPE, #2 in Fig. 1), Chupei (CP, #3 in Fig. 1), Taichung (TC, #4
135 in Fig. 1), Chiayi (CY, #5 in Fig. 1), Tainan (TN, #6 in Fig. 1), Kaohsiung (KH, #7 in Fig. 1), and Hengchun (HC, #8 in Fig. 1) stations to evaluate the modeling performance of temperature, wind speed, and wind direction. Since most residents lived at the relatively flat western Taiwan, the observations at the Banqiao (BQ, #9 in Fig. 1), Pingzhen (PZ, #10 in Fig. 1), Miaoli (ML, #11 in Fig. 1), Zhongming (ZM, #12 in Fig. 1), Chiayi (CY, #13 in Fig. 1), Tainan (TN, #14 in Fig. 1), Zuoying (ZY, #15 in Fig. 1), and Hengchun (HC, #16 in Fig. 1) stations were chosen for PM_{2.5} evaluation.

140 2.2 Models and modeling configuration

This study applied the WRF v3.9.1 (Skamarock and Klemp, 2008) and CMAQ v5.2.1 (Byun and Schere, 2006) for scenario simulations. The initial meteorological condition was from NCEP diagnostic fields. Horizontal resolutions of four domains from outer to inner were 81, 27, 9, and 3 km, respectively. The first domain covered the East Asia and Southeast Asia and the fourth domain contained only the Taiwan island. The vertical layers were 46, about 20 layers below 1.7 km, in which the
145 boundary layer was well resolved. The anthropogenic emissions for East Asia and Taiwan island were obtained from MIX (Multi-resolution Emission Inventory for China, Li et al., 2017) and TEDS 10.0 (Taiwan Emission Data System, TEPA, 2017), which are based on the years of 2010 and 2016, respectively. The MIX emissions of SO₂, NO_x, NMHC, NH₃, CO, PM₁₀, and PM_{2.5} were adjusted with change of -62%, -17%, 11%, 1%, -27%, -38%, and -35%, respectively, according to the change of annual emission between 2010 and 2017 (Zheng et al., 2018). This study assumes the emission of 2017 in Taiwan is the same
150 as that of 2016. The biogenic emissions were prepared by the Biogenic Emission Inventory System version 3.09 (BEIS3, Vukovich and Pierce, 2002) for Taiwan island and Model of Emissions of Gases and Aerosols from Nature v2.1 (MEGAN, Guenther et al., 2012) for regions outside Taiwan. While the biomass burning emissions imported the data of FINN v1.5 inventory (Wiedinmyer et al., 2011). All the remaining modeling configuration for this study is the same as that in Chuang et al. (2017).

155 2.3 Model evaluation

This study used statistical indexes such as MB (Mean Bias), MAGE (Mean Average Gross Error), and IOA (Index of Agreement) to evaluate temperature and wind speed, and used WNMB (Wind Normalized Mean Bias) and WNME (Wind Normalized Mean Error) for wind direction in the fourth domain. For PM_{2.5} performance in the same domain, we applied MB,

MFB (Mean Fractional Bias), and MFE (Mean Fractional Error), R (Correlation coefficient), and IOA indexes. All the formulas
160 for above indexes are from Emery (2001) and TEPA (2016), illustrated in Supplement S1.

2.3.1 Evaluation of WRF meteorological modeling

The MB performance shows that the temperature is slightly overestimated for PJY which is located in the outer sea of northern
Taiwan (Table 1). The MAGE appeals simulated temperature at all stations is reasonable in both months. While the IOA
indicates the simulated temperature at PJY and KH was not well enough. The deviation of simulated temperature for PJY and
165 KH could be influenced by the sea surface temperature since these stations are nearer the sea than other stations. The
performance of MB indicates the simulated wind speed was underestimated at TN, which led to the low IOA. In contrast, the
simulated wind speed was overestimated at HC, which could be due to the smoother terrain in the simulation than the actual
situation. The performance of wind direction at most stations are within the range of acceptance but not so well for TC and
CY. The deviation could potentially due to the influences of nearby buildings. In summary, the simulated temperature, wind
170 speed, and wind direction performed reasonably acceptable since most indices at many stations complied with the benchmark.
The comparisons of observed and simulated temperature, wind speed, and wind direction are illustrated in Fig. S2.1, S2.2, and
S2.3.

2.3.2 Evaluation of CMAQ chemical modeling

For the *Base* case, the simulated PM_{2.5} was overestimated in all stations except CY and HC in January 2017 (Table 2). The
175 performance of trend (correlation coefficient, R) is acceptable or good for all stations except HC. It is rather difficult to simulate
the wind speed at HC well which is located at the downwind south tip of Taiwan (Chuang et al., 2016). It is therefore reasonable
that overestimated wind speed in HC led to poor underestimation of PM_{2.5}. Because the performance of PM_{2.5} in HC is very
poor, the following discussion will exclude this station and leave it to future improvement. The comparison of observed and
simulated PM_{2.5} is illustrated in Fig. S2.4.

180 3. Results and Discussion

3.1 The impact of PM_{2.5} from the Chinese three major industrial regions in January 2017

As mentioned, the impact was considered as the reduction of specific source removed or roughly the contribution of that
specific source, i.e. the difference between the base and sensitivity scenarios. For the impact of three industrial regions on
PM_{2.5} in Taiwan in January 2017, the monthly mean impact from BRIR was about 0.7-1.1 $\mu\text{g m}^{-3}$ as illustrated in Fig. 2(a).
185 The impact was higher in the northern Taiwan, about 5% of total PM_{2.5}. The proportion of influence gradually decreased from
north to south (Fig. 2(b)). From the view of daily average, Fig. 3(a-1)-(a-7) show that the trend is similar for seven air quality
stations and the impact on northern Taiwan was higher than central and southern Taiwan. In January 2017, the proportion of
influence was higher on the 8th to 14th and the 20th to 23rd. It is found that the influence of EAH was closely related to the
intrusion of Asian anticyclones. This study selected Jan 9th and Jan 13th for discussion of PM_{2.5} events in section 3.5.
190 Comparing Fig. 2(a)/(b) with Fig. 2(c)/(d), it is apparent that the monthly mean influence from YRDIR was higher than BRIR.
The reason is that YRDIR was nearer to Taiwan than BRIR. The monthly mean impact from YRDIR was about 1.2-1.9 $\mu\text{g m}^{-3}$,
highest in northern Taiwan with the proportion of about 7.5% of total monthly average PM_{2.5} concentration. The spatial
influence from BRIR was similar to YRDIR since these two industrial regions are both located off the north of Taiwan, i.e.,
the upstream of Taiwan under prevailing northeast wind. For the daily mean influence, the impact of YRDIR was also higher
195 than BRIR and the influencing period were almost the same for both regions (Fig. 3(a-1)-3(a-3), Fig. 3(b-1)-3(b-3)). In
particular, the contributions from BRIR and YRDIR to northern Taiwan could reach 6~8 and 9~12 $\mu\text{g m}^{-3}$ on Jan 9th 2017.

The spatial distribution of influence from PRDIR was totally different from BRIR and YRDIR as shown in Fig. 2(e) and Fig. 2(f). Interestingly, the impact from PRDIR was higher on the mountains than on the ground. For the ground stations, there was minor influence on 8th to 12th January 2017 (Fig. 3(c-1)-3(c-3)). It is found that there is a stationary front from the sea north of Taiwan extended southwest to Fujian and Guangdong provinces on January 7th (Fig. S2.5(a)). The front passed Taiwan on January 8th (Fig. S2.5(b)). Fig. 3(c-1)-3(c-3) show that the influence on the southern Taiwan was higher than that on the northern Taiwan. Similar fronts passed Taiwan on January 10th (Fig. S2.5(c)) and 12th (Fig. S2.5(d)). From Fig. 4, it is found that the PM_{2.5} from PRDIR would transport pollutants upward above the top of boundary and then moved eastwards (Fig. 4(a-1), Fig. 4(b-1)). When pollutants ran into the mountains in Taiwan, most part was blocked and transported to the ground through vertical mixing (Fig. 4(a-2)-4(a-3), Fig. 4(b-2)-4(b-3)). This transport mechanism is quite similar to the biomass burning aerosols from Indochina to Taiwan (Yen et al., 2013; Chuang et al., 2016). **The boundary layer mixing was enhanced by the pass of cold surge and increased PM_{2.5} on the ground.**

3.2 The physical and chemical processes of LRT from the Chinese three major industrial regions to Taiwan in January 2017

This study applied the process analysis technique in the CMAQ model, in which the terms of Horizontal advection (HADV), Vertical advection (ZADV), Horizontal diffusion (HDIF), Vertical diffusion (VDIF), Emissions (EMIS), Dry deposition (DDEP), Cloud process and aqueous chemistry (CLDS), Gas chemistry (CHEM), and Aerosol chemistry (AERO) in the diffusion equation can be resolved (Byun and Schere, 2006). Each term contributes to the rate of change of PM_{2.5} level at the locations chosen in this study: the position #17 (Fig. 1) located between Bohai Sea and East China Sea, #18 (Fig. 1) located between East China Sea and Taiwan, #19 (Fig. 1) located in the middle of Taiwan Strait, #20 located in the northern South China Sea, BQ (#9 in Fig. 1) in northern Taiwan, ZM (#12 in Fig. 1) in central Taiwan, and CY (#13 in Fig. 1) in southwestern Taiwan. Those positions were chosen because they are on the path of northeast wind. Through the value of each term in the process analysis, we can understand whether each term can produce or remove PM_{2.5} at these positions and therefore realize the physical and chemical processes during LRT.

Similar to Fig. 2, we deduced the difference of base and sensitivity scenarios for IPR analysis. This study considered the reduction as the approximate contribution for each industrial region. Therefore, the reader should keep in mind that the following discussion is satisfied on when the chemical reaction between each industrial region and surrounding was ignored. The positive and negative contribution terms in Fig 5 (a-1) and Fig. (a-2) appealed synchronously although their magnitudes were not in equal proportions. It implies #17 was influenced by both BRIR and other nearby sources. The increase of PM_{2.5} was caused mainly by the process HADV, followed by ZADV and VDIF and the removal process was mainly AERO. The removal process is likely caused by the evaporation of ammonium nitrate in PM_{2.5} plume moving from high latitude regions to low latitude regions (Stelson and Seinfeld, 1982; Chuang et al., 2008b). In contrast, there was less PM_{2.5} occasionally from YRDIR (Fig. 5(a-3)) and nearly none from PRDIR (Fig. 5(a-4)). It is expected because northeast wind prevails in winter, the BRIR and YRDIR/PRDIR are located at the upstream and downstream of #17, respectively. From Fig. 5(b-1)-(b-4), among three industrial regions it is apparent that #18 was influenced by both the BRIR and YRDIR, mainly produced through non-uniform HADV, VDIF, ZADV, and CLDS; and removed through AERO and occasional HADV and DDEP processes, and almost unaffected by PRDIR. For #19, PM_{2.5} was influenced mainly by YRDIR (Fig. 5(c-2)) and occasionally by BRIR (Fig. 5(c-3)) for those three industrial regions, but it was also influenced by PRDIR from 8th to 12nd (Fig. 5(c-4)), which has been verified to be related to the intrusion of cold surge and transboundary transport in last section (Fig. 4). The production from BRIR and YRDIR were mainly attributed to CLDS; and removal process was mainly AERO and secondly DDEP. The positive and negative contribution of PM_{2.5} for #20 were very similar to #19 but slightly lower (Fig. 5(d-1)-5(d-4)) because it is farther from BRIR and YRDIR than #19. Although #20 is very near PRDIR, it was influenced more by YRDIR (Fig. 5(d-3)-5(d-4))

and other sources in the north other than three industrial regions since the prevailing wind was mainly northeast wind in January. From above, it is found that the PM_{2.5} plume transported southwards from BRIR or YRDIR in a three-dimensional path, i.e., horizontal and vertical advection, and vertical diffusion over Bohai Sea and East China Sea. During the southward transport, AERO was always the major removal process, i.e., evaporation of volatile species. When the plume transported to subtropical regions, cloud process became the major production process of PM_{2.5}. The reason was possibly the condensation in the mix of cold PM_{2.5} plume from high latitude regions to warm air/sea at low latitude regions.

The build-up of PM_{2.5} at BQ were mainly HADV with minor CLDS, and the removal processes were mainly ZADV with minor AERO (Fig. 5(e-1)). It suggests that the PM_{2.5} plume transported in a mainly horizontal when it was close to and reached northern Taiwan. Moreover, each industrial region contributed PM_{2.5} to BQ in very similar processes (Fig. 5(e-2)-(e-4)). In addition, certain PM_{2.5} was formed in northern Taiwan probably due to the high relative humidity, which was probably induced by the cloud or fog produced by terrain uplifting. The removal process of PM_{2.5} at BQ was mainly ZADV, which implies PM_{2.5} at BQ then transport up and reflects BQ is located in a basin. Comparing Fig. 5(f-1) with Fig 5(f-2)-Fig 5(f-3), it is obvious that the PM_{2.5} of ZM was produced more by local from vertical transport than BRIR or YRDIR, which only exerted less PM_{2.5} along with the cold surge, and removed by horizontal transport. In other words, the PM_{2.5} in upstream northern Taiwan was vertically advected and diffused southwards to central Taiwan and then horizontally advected to downwind areas. On the other hand, the influence from PRDIR was much less when the prevailing wind was northeast monsoon (Fig. 5(f-4)). However, when the cold surge passed Taiwan (Jan 8th and 10th), the influence from PRDIR could not be ignored, which has been illustrated in Fig 2(f), Fig. 4 and Fig. 5(f-4). On Jan 8th to 10th, the negative ZADV indicated the concentration was decreasing at the lower 20 averaged layers but the concentration gradient was positive ($\frac{\partial PM_{2.5}}{\partial z} > 0$, the concentration of PM_{2.5} from PRDIR was higher at high altitude than that at low altitude over Taiwan) implies the vertical velocity had to be negative, i.e. downward motion. Therefore, the boundary layer mixing of the aloft PM_{2.5} plume was enhanced by the pass of the cold surge. (Yen et al., 2013; Chuang et al., 2016). For CY located in southwestern Taiwan, VDIF and HADV mainly contributed to the gains of PM_{2.5}, and the removal processes were mainly ZADV and AERO; however, occasionally when the positive contribution to PM_{2.5} were ZADV and VDIF, the removal processes were HADV and AERO (Fig. 5(f-1)). Compared Fig. 5(f-2)-(f-4) and Fig. 5(g-2)-(g-4), it is obvious the positive and negative contribution to PM_{2.5} for CY were very similar to for ZM. The impact from BRIR and YRDIR was less and mainly from local. When the cold surge passed Taiwan, PRDIR influenced PM_{2.5} at CY as well.

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4 Analysis of the strong episodes occurring on 13th January 2017

On January 13th 2017, the Asian anticyclone transported pollutants from Asian continent to Taiwan and caused high PM_{2.5} episodes. Such LRT events occurred at a weather pattern as illustrated in Fig. 6. Although the impact of LRT on Jan 13th was less than Jan 8th, 9th, 20th or 22nd (Fig. 3), the physical and chemical processes during transport were similar for these days since the weather patterns were quite analogous to each other. The Asian anticyclone was moving from East Asian to the West Pacific. The peripheral circulation of the Asian anticyclone was the strong northeast wind on coastal areas and the sea. It was found the northeast wind formed lee wakes in southern Taiwan where PM_{2.5} accumulated (Fig. 7(a)-(b)). When the leading edge of Asian anticyclone arrived, the wind speed increased and therefore enhanced the dispersion of PM_{2.5} in southern Taiwan (Fig. 7(c)-(e)). Subsequently, the LRT haze arrived (Fig. 7(f)) and split to the east and west side of Taiwan due to the blocking of mountains, more on the west side. (Fig. 7(g)-(i)).

Fig. 8(a-1)-(a-4) shows that the influence of BRIR on #17 was more than YRDIR and PRDIR on Jan 13th, since BRIR is located at the upstream of #17 under northeast wind. The major production process was VDIF below 760 m (layer 14) and AERO above 760 m. It implies the transport path from BRIR to #17 could be horizontal between BRIR and #17 and then vertical at the location of #17. The removal process was AERO below 760 m and VDIF above. It suggests the ascent and

280 subsidence of air parcels might enhance the formation and removal of aerosol in upper and lower level, respectively. It is possibly that the ascent motion of air parcel near the warm surface moved to a cold environment in higher altitude. This may cause condensation and triggered heterogeneous reactions of aerosols. On the contrary, the descent motion of air parcel may cause the evaporation of aerosols. Although #17 was slightly influenced by YRDIR, the contribution of different processes from YRDIR on #17 was less and non-uniform (Fig. 8(a-3)). The contribution of different processes from PRDIR to #17 was also non-uniform and even less (Fig. 8(a-4)). From Fig. 8(b-1)-(b-4), it was found that #18 was mainly influenced by YRDIR on Jan 13th. The major processes below layer 9 (~310 m) contributing to the increase of PM_{2.5} were HADV, VDIF, and ZADV and removal processes were DDEP and AERO (Fig. 8(b-3)). #18 was slightly influenced by BRIR with major production process were VDIF and ZADV and removal process was AERO (Fig. 11(b-2)). On Jan 13th, #19 and #20 were less influenced by all industrial regions (Fig. 8(c-2)-(c-4), Fig. 8(d-2)-(d-4)). It implied that #19 was possibly influenced by nearby Fujian province on the north and west side of Taiwan Strait. On Jan 13th, #20 was also less influenced by three industrial regions probably due to BRIR and YRDIR was distant and PRDIR was located at the downstream of #20. Comparing Fig. 8(e-1) and Fig. 8(e-2)-8(e-4), it was found the BQ was much influenced by YRDIR. Although #18 and BQ were most affected by YRDIR, the major contribution process at BQ below 200 m (layer 7) was HADV, followed by AERO and above 200 m was either of VDIF, ZADV, CLDS and mixture of them. The major removal process was ZADV followed by VDIF below 200 m but HADV and AERO above. BQ was less influenced by BRIR due to long distance, deviation of wind direction and by PRDIR due to BQ is located at upstream of PRDIR. In this event, ZM and CY were less influenced not only by BRIR and PRDIR but also YRDIR (Fig. 8(f-1) - Fig. (g-4)). It explains the haze plume passed BQ and then transported to the west coast of Taiwan instead the inland ZM and CY on Jan 13th.

3.5 Analysis of the moderate episodes occurring on 9th January 2017

300 The PM_{2.5} event occurring in western Taiwan on Jan 9th was similar to that on Jan 13th, which were both LRT of EAH. However, there were still slightly differences between these two events. First, the impact of three industrial regions on PM_{2.5} in western Taiwan was much higher on Jan 9th than Jan 13th. Second, for the haze from BRIR and YRDIR, the positive and negative contribution processes on BQ were mainly HADV/AERO and ZADV/VDIF below 200 m (layer 7, Fig. 8(e-3)) and less different processes at different layers above 200 m on Jan 13th. While on Jan 9th, the major processes leading to the increase of PM_{2.5} at BQ was mainly HADV below 380 m (layer 10), AERO between 120 to 900 m (layer 5 to 15), and ZADV/CLDS between 650 to 1500 m (layer 13 to 19), as illustrated in Fig. 9(e-2)-(e-3). The removal process was mainly ZADV below 460 m (layer 11), HADV between 550 to 900 m (layer 12 to layer 15), and HADV/AERO between 1000 to 1300 m (layer 16 to 18). Third, the stronger event occurring on Jan 9th has more obvious impact on ZM and CY than that on Jan 13th. vv Jan 9th explains the rapid moving EAH. In contrast, the higher production of AERO near surface occurring on Jan 310 13th explains slow moving EAH had time to react with the local pollutants, e.g. HNO₃ in Asian plume reacted with local NH₃ to form NH₄NO₃, which has been discussed in Chen et al. (2014).

3.5 The impact of PM_{2.5} from the Chinese three major industrial regions in July 2017

The Fig. 10(a) and Fig. 10(b) reveals that the impact of BRIR on PM_{2.5} in Taiwan was negligible. The monthly contribution was less than 0.01 μg m⁻³ or less than 0.04% of total PM_{2.5} on the western Taiwan. The influence from YRDIR and PRDIR on Taiwan was equally small with BRIR (Fig. 10(c)-Fig. 10(f)). The daily contribution from three industrial regions to the western Taiwan was similar for all cities. The contribution from BRIR was only with less than 0.1 μg m⁻³ from 25th to 28th July (Fig. S2.6(a-1)-(a-7)), from YRDIR was about 0.1-0.3 μg m⁻³ from 27th to 29th July (Fig. S2.6(b-1)-(b-7)), and detectable on 28th July but rose to 0.2-0.5 μg m⁻³ on 30th to 31st July (Fig. S2.6(c-1)-(c-7)). Owing to the small impact from three industrial regions on the western Taiwan, the physical and chemical processes was small for all days in July 2017 except the last few days in that month, as illustrated in Fig. S2.6, The weather map revealed that there was a thermal low near Taiwan at the end 320

of July (Fig. S2.7). In short, during the period of prevailing southwest wind, the influence of BRIR, YRDIR, or PRDIR could be ignored unless there was special weather system like the aforementioned thermal low which could transport less PM_{2.5} from distant sources.

From Fig. S2.8(a-1) to Fig. S2.8(a-4), it was found that #17 was influenced more by YRDIR than BRIR or PRDIR on July 18th 2017. The positive and negative contribution processes were non-uniform below 80 m (layer 4). But from 120 m to 460 m (layer 5 to layer 11), the major processes to build-up of PM_{2.5} were AERO and ZADV and the removal process was mainly HADV. The Fig. S2.8 shows that the influence of three industrial regions on #18, #19, #20, BQ, ZM, or CY was almost ignorable. It suggested the PM_{2.5} was mainly from local pollution in July. On the other hand, the #19, #20, BQ, ZM, and CY was influenced by PRDIR at the end of July 30th (Fig. S2.8). As mentioned earlier, the thermal low over Taiwan Strait (Fig. S2.7) caused unstable wind field and transported pollutants from coastal areas of Asian continent to northern South China Sea and Taiwan strait Taiwan (Fig.S2.9). In July 2017, there is hardly amount of PM_{2.5} transported from three industrial regions to those specific locations on July 30th except from PRDIR to #20, as illustrated in Fig. S2.10.

3.6 Discussion of the chemical compositions and emissions

Lee et al. (2017) conducted PM_{2.5} sampling at BQ, ZM, and CY every six days in 2017. The sampling of Jan 13th was used to compare with simulated PM_{2.5} compositions, as indicated in Fig. 15 The previous studies (Chuang et al., 2008b ; Wang et al., 2016) suggested it took about 28 hours for the PM_{2.5} haze transported from Yangtze River estuary to the northern tip of Taiwan island. Therefore, the simulated PM_{2.5} compositions at #17 and #18 on Jan 12th were also illustrated. According to the main content, among those three industrial regions BRIR and YRDIR were the major sources of #17 and #19 - #20, respectively. As illustrated in Fig. 11, no matter on Jan 12nd or Jan 13th, the major compositions were sulfate and OC for #17 - #20. However, the proportion of nitrate in PM_{2.5} at #17 on Jan 12th was higher than those at #18, #19, and #20 on Jan 13th. It explains the nitrate would evaporate from aerosol phase to gas phase for PM_{2.5} plume transported from high to low latitude regions (Chuang et al., 2008b). The proportions of Na⁺ and Cl⁻ in PM_{2.5} at # 19 and #20 were higher than those at #17 and #18. The higher sea salt due to stronger wind speed is expected because the Taiwan Strait was a wind tunnel between Central Mountain Range in Taiwan and Wuyi Mountain Range in Fujian province (Lin et al., 2012). In addition, the proportions of nitrate in PM_{2.5} at BQ, ZM, and CY were higher than those over #17 - #20. That should be caused by the local pollution. The comparison between simulation and observation indicated that the performance of simulation was not bad. The simulated proportion of nitrate and ammonium in PM_{2.5} was slightly lower than the observations. While the simulated proportion of K⁺, Ca²⁺, Mg²⁺, Na⁺ was slightly overestimated. This suggested the emission of biomass burning and wind-blown dust over Taiwan island and the influence of sea salt still have room for improvement.

We also compared the simulated PM_{2.5} compositions with observations on July 18th 2017 (Fig. S2.11). As mentioned in main content, #17 was influenced by upstream YRDIR, the proportion of nitrate in PM_{2.5} #17 was higher than further upstream #18, #19, and #20. The proportion of nitrate in PM_{2.5} at #19 and #20 was higher than #18, it implies #19 and #20 were influenced more by PRDIR than #18. For BQ, ZM, and CY, the proportion of simulated OC in PM_{2.5} was slightly overestimated as compared with observation but nitrate, sulfate and others were underestimated. Since BQ, ZM, CY were less influenced by PRDIR on July 18th, the overestimation of OC and underestimation of nitrate should be related to the bias of local emission inventory. In addition to local emission inventory, the underestimation of sulfate could possibly be related to underestimation of emission from ships around Taiwan since the local emission of SO₂ is quite low. Moreover, the uncertainty of emission in the Southeast Asia is also another issue that needs to be improved.

On July 30th 2017, there was a thermal low which influence the circulation near Taiwan. The Fig. S2.12 illustrates that BQ, ZM, and CY were influenced by local pollution and therefore the proportions of EC and NH₄⁺ in PM_{2.5} at these three cities were higher than #17 - #20. It was not easy to form nitrate at BQ, ZM, and CY since the circulation was strong and cloud cover was intense (no PM_{2.5} sampling on July 30th due to bad weather condition).

4. Conclusions

This study evaluated the impact of the three biggest industrial regions in Asian continent on PM_{2.5} in Taiwan and discussed the process analysis during transport. It applied the CMAQ model with BFM method and process analysis technique. The simulation period was January and July 2017.

In January 2017, the LRT from Asian continent to Taiwan was substantial over northern Taiwan and gradually minor in central and southern Taiwan. The impact of monthly PM_{2.5} from BRIR and YRDIR on Taiwan was 0.7-1.1 μg m⁻³ and 1.2-1.9 μg m⁻³, about 5% and 7.5% of total concentration, respectively. The daily impact was the most on January 9. The contribution from BRIR and YRDIR on Taiwan was 6-8 and 9-12 μg m⁻³, respectively. In contrast, the influence of PRDIR to Taiwan was ignorable. However, when the cold surge passed Taiwan, the PM_{2.5} from PRDIR can influence Taiwan with monthly average impact of about 0.5 μg m⁻³ via transboundary transport and boundary layer mixing (VDIF). When the cold surge induced-events occurred, the impact from BRIR and YRDIR was substantial on BQ. The transport mechanism of EAH from BRIR and YRDIR was horizontal (HADV) and vertical (ZADV and VDIF) at Bohai Sea and East China Sea. When the EAH moved to Taiwan Strait and northern South China Sea, CLDS became the major production of PM_{2.5} under high relative humidity environment. Along the transport, AERO and DDEP were always the removal process for the EAH transporting from high latitude regions to low latitude regions. When the EAH moved to northern Taiwan, HADV and AERO were the major contribution processes of PM_{2.5} at BQ. The transport mechanism from northern Taiwan to central Taiwan and southern Taiwan was changeable due to complex terrain and complex land canopy. In addition, the intensity of EAH would have different production and removal processes in different height. The stronger the intensity of EAH, the impact on central and southern Taiwan was more obvious, the proportion of HADV contributed to PM_{2.5} budget was more obvious near surface.

In July 2017, the influence from three industrial regions on the PM_{2.5} was ignorable in Taiwan, i.e. PM_{2.5} was mainly come from upwind adjacent local sources unless if there was special weather system, e.g. a thermal low nearby which may carry small amount of pollutants from PRDIR to Taiwan.

In regards of performance of MIX emission inventory, this study compared the simulated and observed PM_{2.5} compositions on Jan 13th, July 18th, and July 30th. The simulated proportion of nitrate and ammonium in PM_{2.5} during the winter time was slightly overestimated but the simulated K⁺, Ca²⁺, Mg²⁺, Na⁺ was underestimated at BQ, ZM, and CY. It suggested the bias in the local emission inventory has lacked the correct information of local biomass burning. During the summertime, the simulated proportion of OC in PM_{2.5} was overestimated but underestimated for nitrate, sulfate, and others. In addition to the bias of local emission inventory, the LRT emission of sulfate is another reason that caused the difference.

Author contribution

Ming-Tung Chuang designed the experiment, carried out most part of the study and wrote the original draft.

Maggie Chel Gee Ooi helped produce half of figures and revised the manuscript.

Neng-Huei Lin is the project leader, provided consultation and acquired the financial support for this study.

Joshua S. Fu submitted valuable questions and helped enhancement of the writing.

Chung-Te Lee provided the PM_{2.5} compositions data and provided related consultation.

Sheng-Hsiang Wang provided beneficial consultation according to his previous publications.

Ming-Cheng Yen provided beneficial consultation according to his previous publications.

Steven Soon-Kai Kong helped part of the post processing of the simulation results.

Wei-Syun, Huang helped maintenance of the computing machine.

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Table 1 The performance of meteorological modeling results for the present study

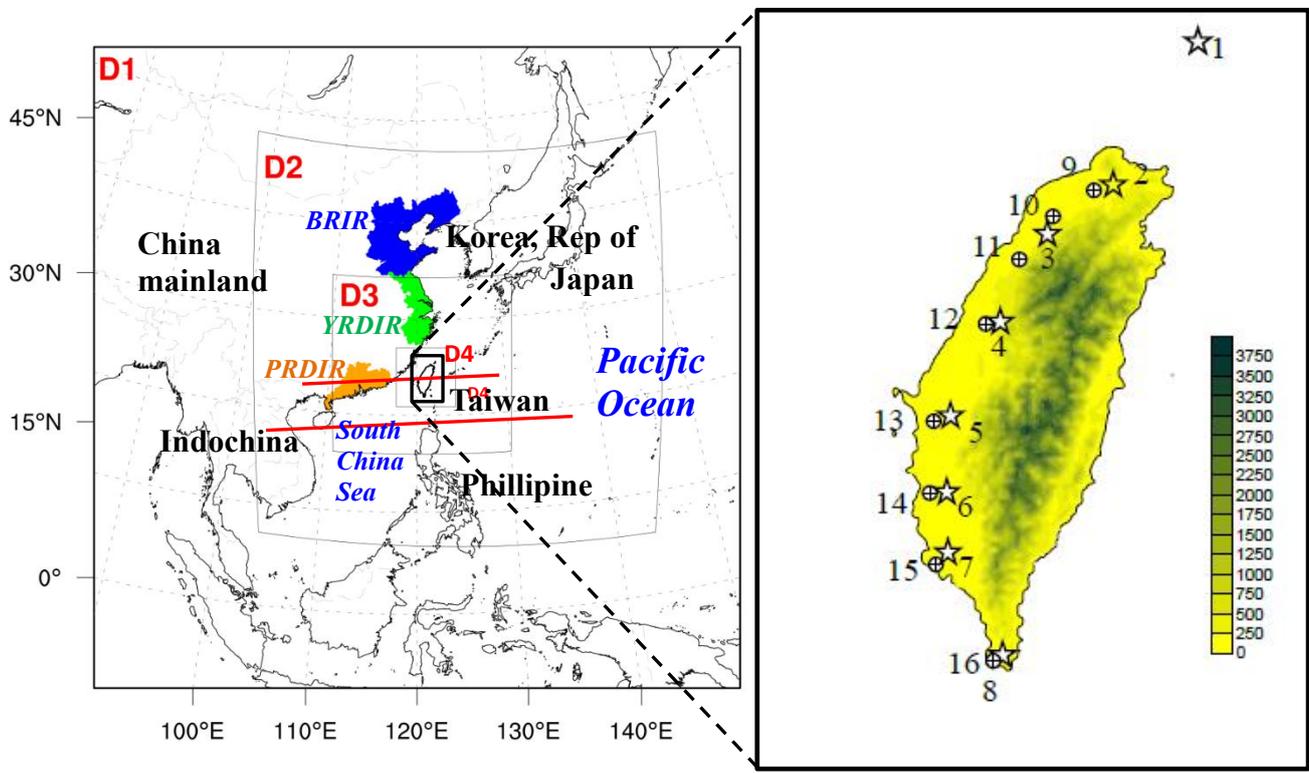
		Temperature			Wind speed			Wind direction	
Standard		MB (°C)	MAGE (°C)	IOA	MB (m s ⁻¹)	MAGE (m s ⁻¹)	IOA	WNMB	WNME
		±1.5	<3	>0.7	±1.5	<3	>0.6	±10%	<30%
PJY	Jan	1.54	1.63	0.90	-0.01	1.16	0.91	-2.09	5.91
	July	0.43	1.18	0.69	0.05	1.29	0.93	0.00	4.27
TPE	Jan	0.00	0.60	0.99	-0.75	1.10	0.74	8.91	13.16
	July	-0.31	0.98	0.91	-0.06	0.92	0.81	5.71	22.04
CP	Jan	0.12	0.61	0.98	0.52	0.86	0.84	2.70	13.85
	July	-0.02	0.73	0.95	0.16	0.68	0.80	4.50	19.01
TC	Jan	0.17	1.02	0.96	0.06	0.47	0.87	3.16	41.33
	July	0.61	1.19	0.92	0.05	0.56	0.80	6.84	25.30
CY	Jan	0.05	0.83	0.98	-0.21	0.61	0.83	12.34	32.40
	July	0.02	1.06	0.93	-0.35	0.83	0.78	5.61	21.18
TN	Jan	0.18	0.83	0.97	-1.82	1.84	0.52	9.42	20.26
	July	-0.14	0.85	0.93	-0.97	1.12	0.69	-1.33	20.76
KH	Jan	-0.07	0.94	0.93	1.15	1.26	0.60	4.22	23.40
	July	-1.27	1.47	0.66	1.19	1.56	0.73	4.84	12.81
HC	Jan	-1.29	1.39	0.88	2.17	2.31	0.80	-0.60	7.39
	July	-0.79	1.13	0.90	1.88	1.96	0.66	1.01	8.58

Note: The standard of statistical evaluation is based on [Emery \(2001\)](#) and [TEPA \(2016\)](#).

Table 2 Simulated PM_{2.5} at eight air quality stations in western Taiwan

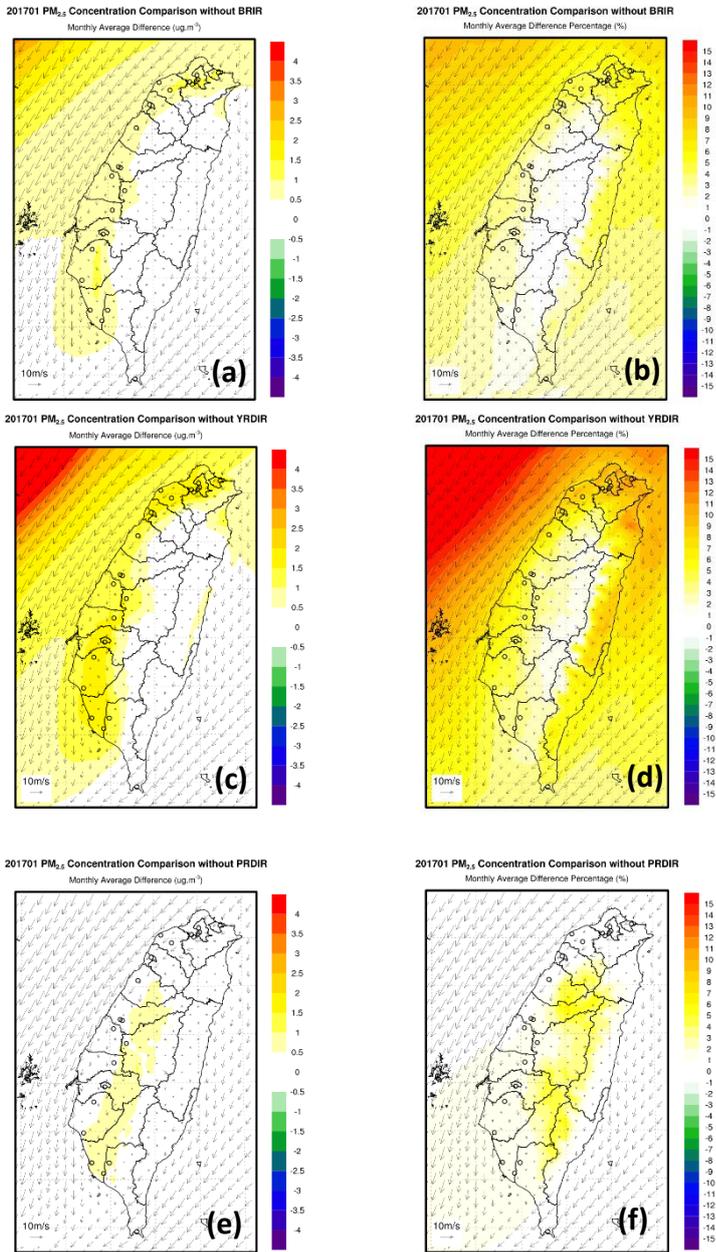
		MB	MFB (%) <±65	MFE (%) <85	R >0.5	IOA >0.6
BQ	Jan	5.0	10%	38%	0.85	0.82
	July	5.3	40%	49%	0.46	0.55
PZ	Jan	5.1	9%	38%	0.71	0.68
	July	3.2	17%	29%	0.63	0.67
ML	Jan	0.2	-17%	42%	0.73	0.77
	July	4.8	22%	40%	0.76	0.65
ZM	Jan	5.5	12%	29%	0.82	0.83
	July	3.3	16%	33%	0.68	0.76
CY	Jan	-2.6	-10%	23%	0.69	0.80
	July	0.3	5%	30%	0.52	0.70
TN	Jan	0.5	-2%	22%	0.64	0.77
	July	7.4	46%	46%	0.69	0.68
ZY	Jan	1.1	1%	17%	0.67	0.79
	July	1.7	12%	35%	0.52	0.72
HC	Jan	-4.1	-62%	77%	0.14	0.43
	July	0.4	-18%	53%	0.19	0.26

Note: the standard of statistical evaluation is based on [Emery \(2001\)](#) and [TEPA \(2016\)](#).



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Figure 1: Geographic location of three major industrial regions (BRIR (blue line enclosed region), YRDIR (green) and PRDIR (orange)) in East Asia and meteorological and air quality stations in Taiwan. Meteorological stations: #1: PJY, #2: TPE, #3: CP, #4: TC, #5: CY, #6: TN, #7: KH, and #8: HC; air quality stations: #9: BQ, #10: PZ, #11: ML, #12: ZM, #13: CY, #14: TN, #15: ZY, and #16: HC. The circular, triangle, diamond, and rectangular symbols are #17, #18, #19, and #20, respectively. The red line is the cross-section plot for Figure 4



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Figure 2: The monthly average wind field and impact of PM_{2.5} from BRIR: concentration (a) and percentage (b) ; YRDIR: concentration (c) and percentage (d) ; PRDIR: concentration (e) and percentage (f) on Taiwan in January 2017

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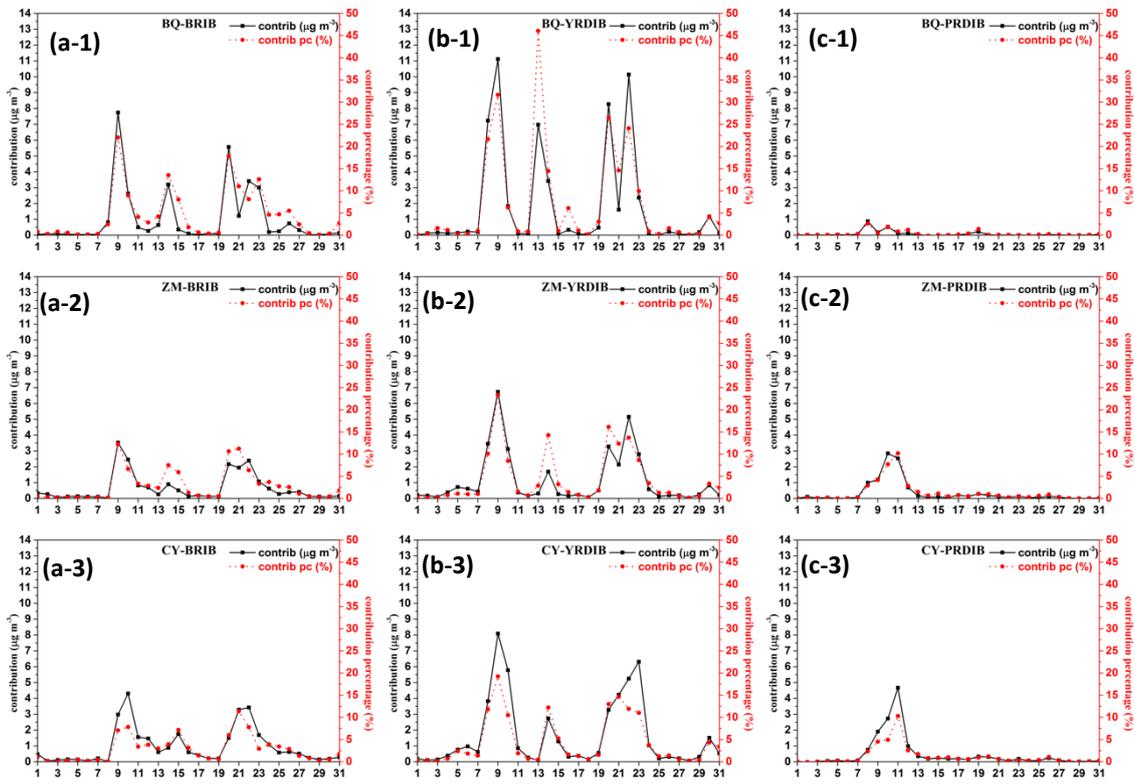


Figure 3: The daily average impact of PM_{2.5} from BRIR, YRDIR, PRDIR on air quality stations in Taiwan in January 2017

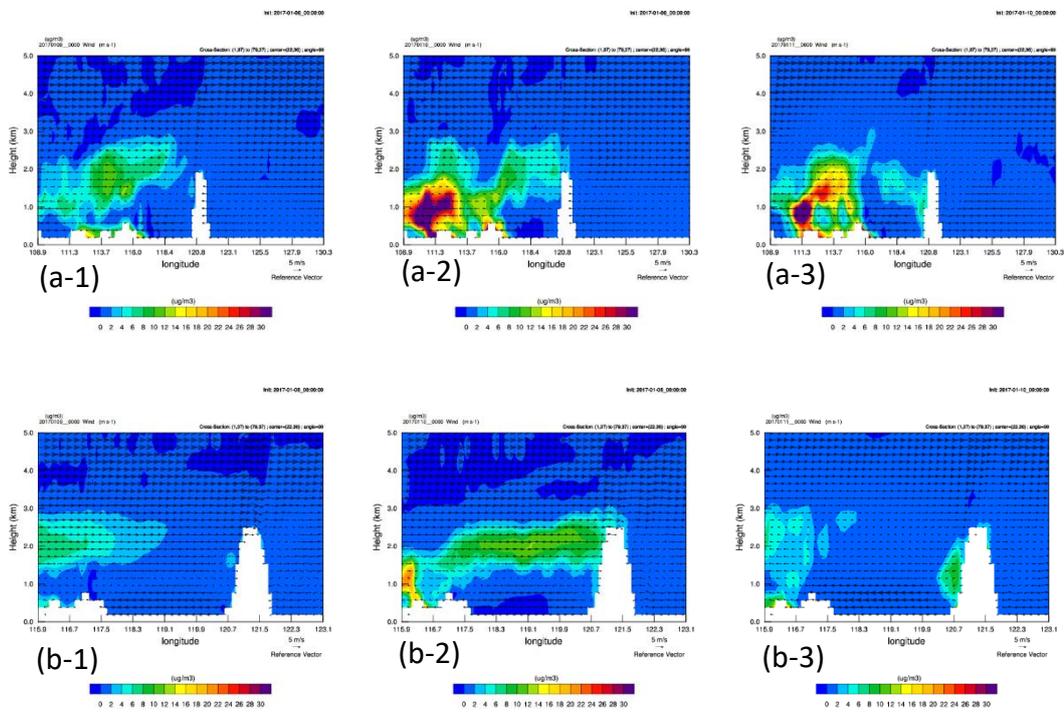
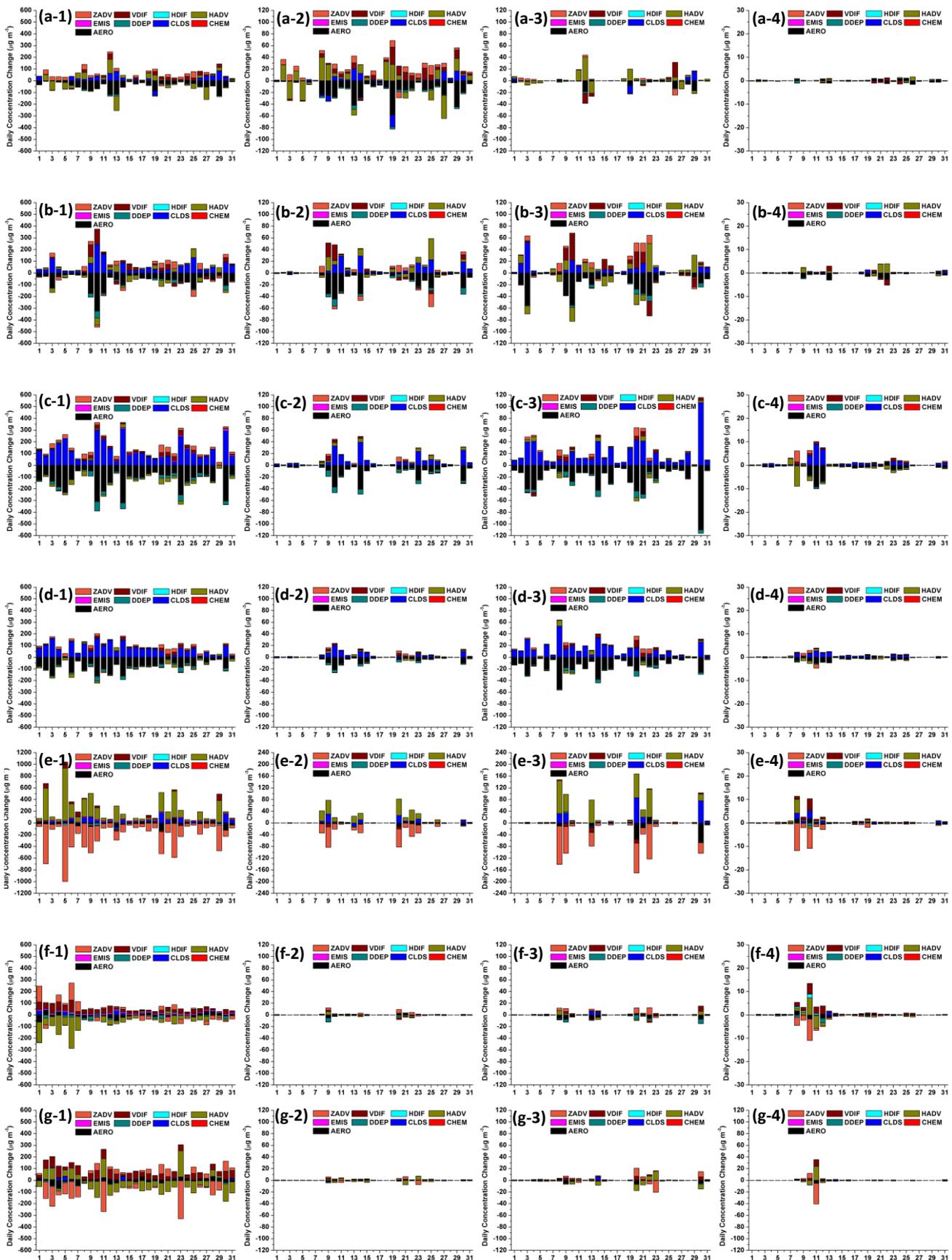


Figure 4: Cross-section plot of PM_{2.5} along the red line of Fig. 1 at 08:00 LT (Local Time) on Jan 9th (a-1), 08:00 LT on Jan 10th (a-2), 08:00 LT on Jan 11th (a-3) of domain 2 for Base case minus Prdir case. Synchronized plots for domain 3 are (b-1) to (b-3)

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Figure 5: The daily contributions of individual processes to the concentrations of PM_{2.5} in January 2017, a,b,c,d,e,f, and g represent #17, #18, #19, #20, BQ, ZM, and CY, respectively ; 1, 2, 3, and 4 represent influence of total emissions, BRIR, YRDIR, and PRDIR, respectively

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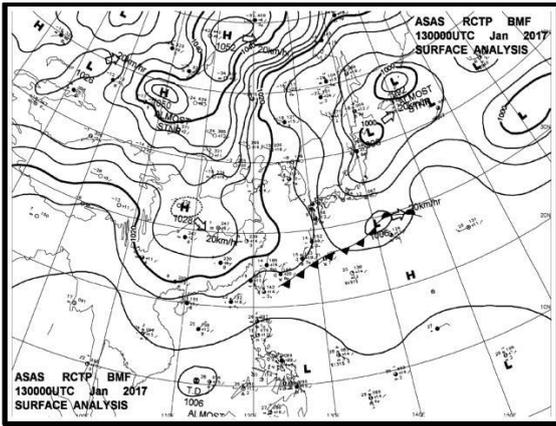
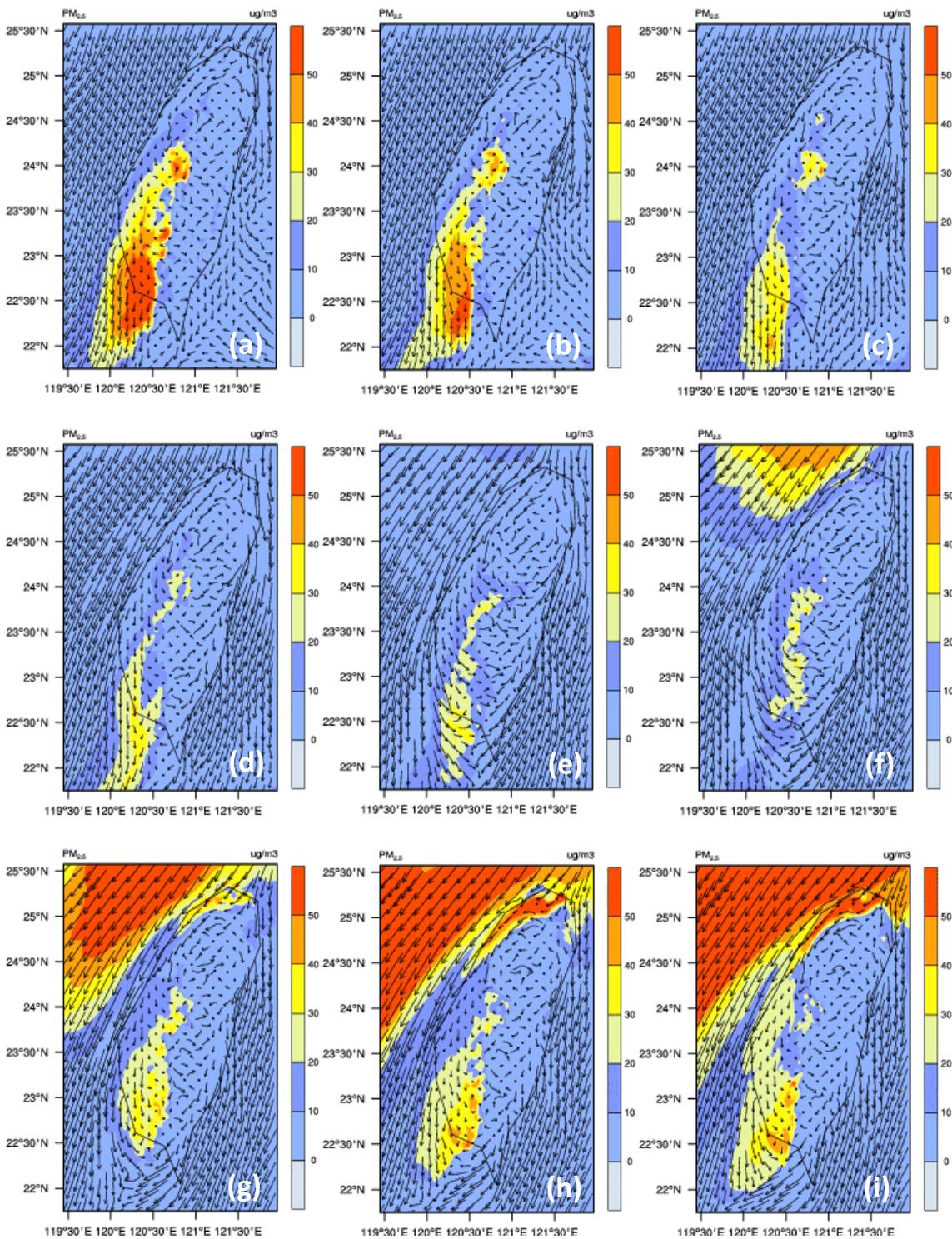


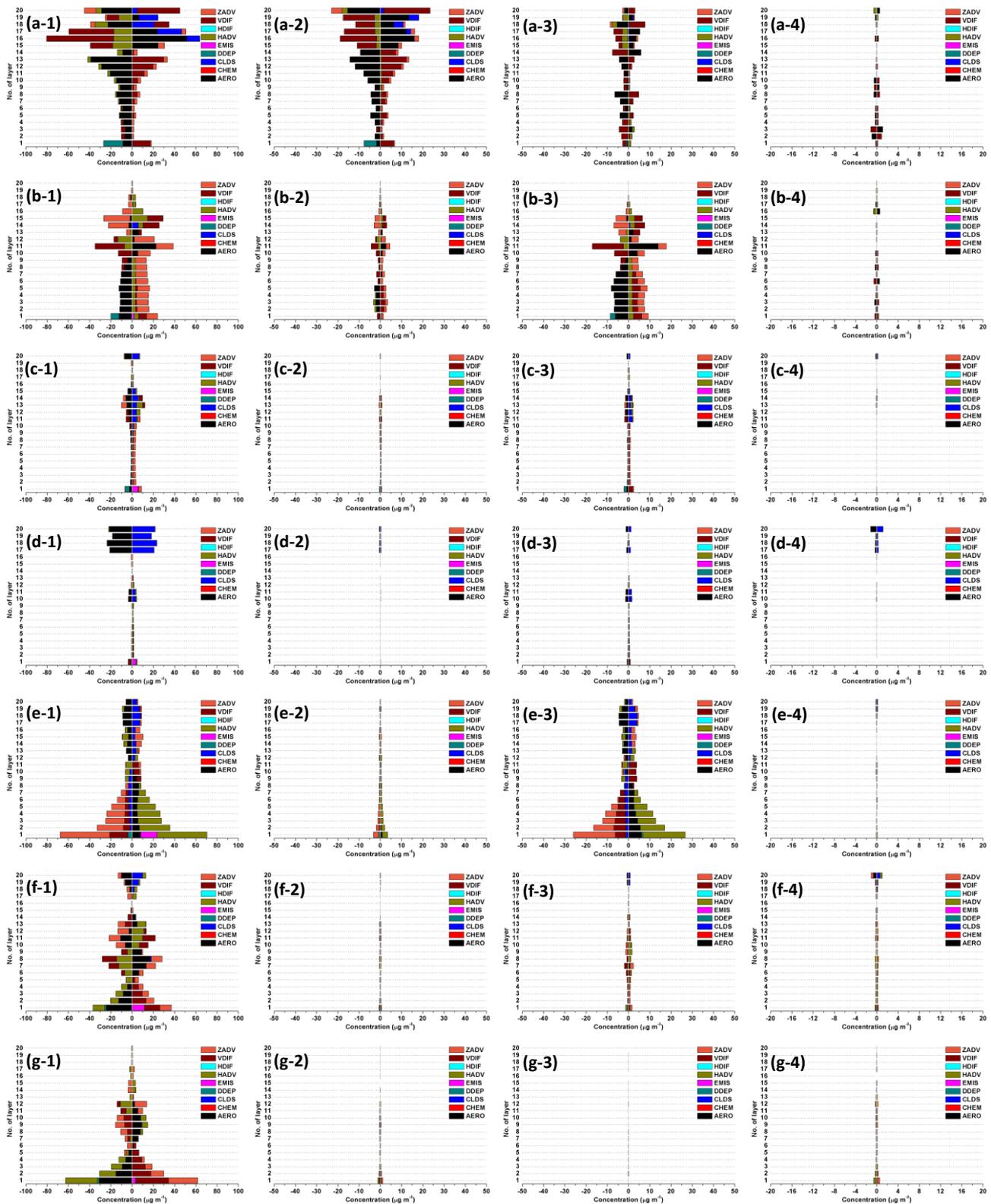
Figure 6: The surface weather map on 08:00 LT Jan 13th 2017



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Figure 7: The every 3 hour simulated wind vector and PM_{2.5} distribution on the event at 00:00 LT (a) 03:00 LT (b) 06:00 LT (c) 09:00 LT (d) 12:00 LT (e) 15:00 LT (f) 18:00 LT (g) 21:00 (h) Jan 13th and 00:00 LT (i) Jan 14th 2017

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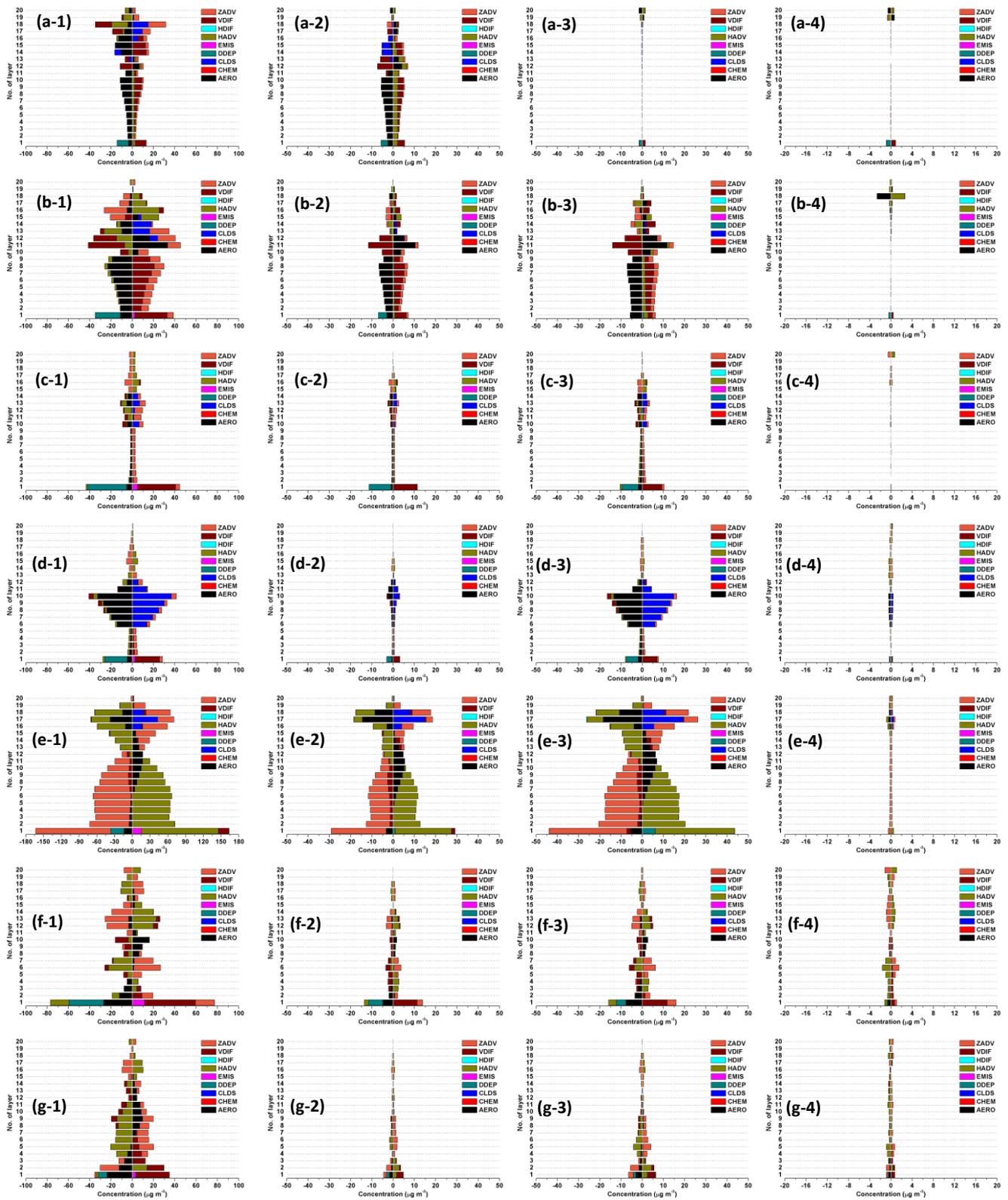


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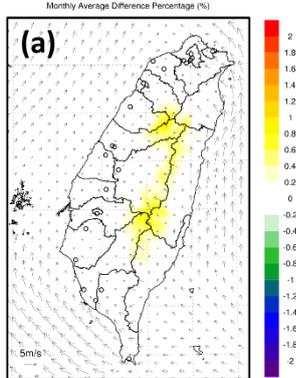
Figure 8: The hourly average contribution of physical process at each layer on Jan 13th 2017, a,b,c,d,e,f, and g represent #17, #18, #19, #20, BQ, ZM, and CY, respectively ; 1, 2, 3, and 4 represent influence of total emissions, BRIR, YRDIR, and PRDIR, respectively

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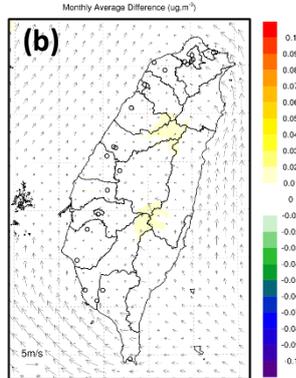


625 Figure 9 The hourly average contribution of physical process at each layer on Jan 9th 2017, a,b,c,d,e,f, and g represent #17, #18, #19, #20, BQ, ZM, and CY, respectively ; 1, 2, 3, and 4 represent influence of total emissions, BRIB, YRDIB, and PRDIB, respectively.

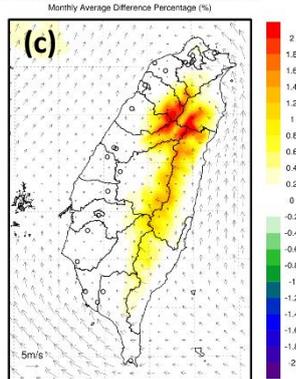
201707 PM_{2.5} Concentration Comparison without BRIR



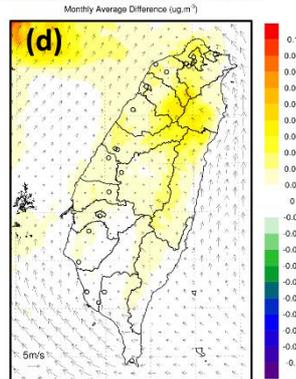
201707 PM_{2.5} Concentration Comparison without BRIR



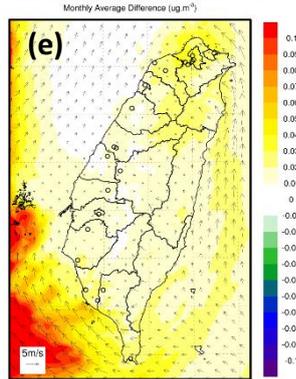
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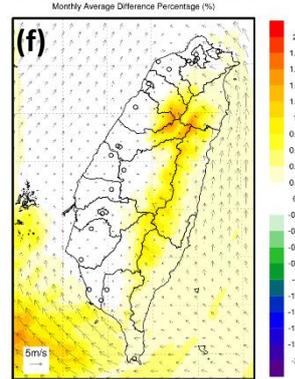
201707 PM_{2.5} Concentration Comparison without YRDIR



201707 PM_{2.5} Concentration Comparison without PRDIR



201707 PM_{2.5} Concentration Comparison without PRDIR



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635 **Figure 10: The monthly average impact of PM_{2.5} from BRIR: concentration (a) and percentage (b) ; YRDIR: concentration (c) and percentage (d) ; PRDIR: concentration (e) and percentage (f) on Taiwan in July 2017**

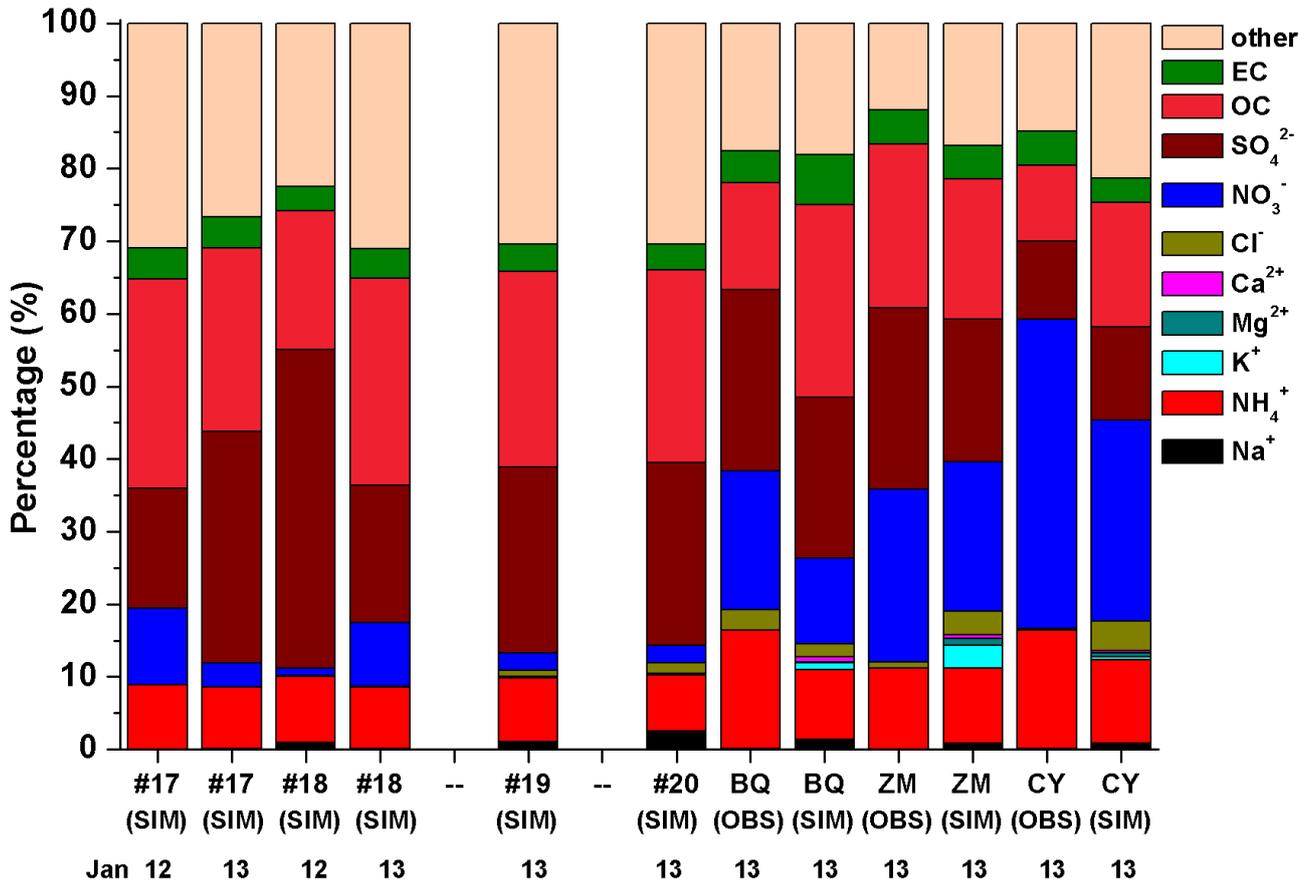


Figure 11: The comparison of simulation (SIM) and observation (OBS) of PM_{2.5} compositions at #17-#20 and BQ, ZM, and CY on Jan 12th and 13th 2017

